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CAROLINA BLEFARI BATISTA

**BIOGEOGRAFIA E CONSERVAÇÃO DE MORCEGOS NA
MATA ATLÂNTICA**

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2020

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Tese apresentada ao Programa de Pós-graduação em Ciências Biológicas da Universidade Estadual de Londrina - UEL, como requisito parcial para a obtenção do título de Doutor.

Orientador: Prof. Dr. Marcos Robalinho Lima.
Co-orientador: Prof. Dr. Isaac Passos de Lima.

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“Explicar completamente a natureza é uma tarefa difícil demais para qualquer homem, e mesmo para qualquer época...”

É muito melhor fazer um pouco com certeza e deixar o resto para os que vierem depois do que explicar todas as coisas por conjectura sem se certificar de nada...”
(Isaac Newton)

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RESUMO

Entender como as espécies estão distribuídas no espaço, quais são os fatores limitantes de suas distribuições, e como isso tudo varia de acordo com a escala de análise, é importante para ajudar a prever a dinâmica das populações em resposta às mudanças climáticas e outras perturbações no ambiente. Essas informações são úteis para guiar o planejamento da conservação prática, de forma a aumentar a eficiência da relação custo-benefício nos investimentos à longo prazo. Para compreender padrões de distribuição é necessário que a base de dados de ocorrência de espécies utilizada nas análises seja uma fonte confiável e atualizada, com o máximo de informações precisas (ex. taxonomia, local georreferenciado, indicação do esforço amostral). Dessa forma, o primeiro capítulo objetivou revisar as ocorrências de morcegos para a região de Londrina, Paraná, nos últimos 36 anos, e compilar um banco de dados com informações sistematizadas, além de identificar as lacunas de amostragem para ajudar a direcionar onde e como novos esforços devem ser empregados. Essa revisão foi baseada em dados publicados e não publicados. O segundo capítulo, objetivou identificar os padrões e os preditores (ambientais e espaciais) da beta diversidade de morcegos na Mata Atlântica, em diferentes escalas de análise. Foi utilizada uma análise de particionamento da beta diversidade (em aninhamento e substituição) e um Modelo de Dissimilaridade Generalizado para correlacionar a beta diversidade com os preditores. O terceiro capítulo, objetivou responder: 1) Onde estão localizados os centros de maior concentração e restrição de diversidade de morcegos dentro dos limites do *hostspot* global Mata Atlântica? e 2) Quanto em extensão dessas áreas já encontram-se protegidas por lei? Para isso foi utilizada uma análise de Interpolação Geográfica de Endemismo (GIE) e uma sobreposição do mapa de Unidades de Conservação (UCs) ao mapa resultante da GIE. Os resultados apresentados nos três capítulos: 1- servem de base para estudos que queiram utilizar o banco de dados da distribuição de morcegos na região de Londrina; 2- Indicam que esta região ainda está subamostrada e direcionam novos esforços de amostragem para morcegos; 3- identificam as ecorregiões mais ricas em morcegos para a Mata Atlântica (costa e interior da Bahia e Serra do Mar); como a beta diversidade varia ao longo do gradiente latitudinal e longitudinal do bioma; e como a percepção dessa variação muda de acordo com a escala de análise; 4- ajudam a prever como as mudanças ambientais podem interferir na distribuição das espécies, ao indicarem a importância dos fatores ambientais na composição das espécies de morcegos; 6- apresentam uma área de endemismo para 63% das espécies de morcegos na Mata Atlântica, localizada no centro do bioma; 6- demonstra como a GIE, usando morcegos como modelo, pode ser uma análise que resulta em uma estratégia mais eficaz no planejamento das UCs na Mata Atlântica; e 7- indicam a urgência na criação de UCs dentro das regiões da costa e interior da Bahia.

Palavras-chave: Chiroptera; banco de dados; padrões de distribuição; áreas de endemismo.

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ABSTRACT

Understanding how species are distributed in space, which factors limit their distributions, and how these factors vary according to the scale of analysis, are important to help predict population responses to climate change and other environmental disturbances, for example. This information is useful to guide practical conservation planning in order to increase the cost-effectiveness of long-term investments. To understand distribution patterns, it is necessary that the species occurrence database used in the analyzes is reliable and up to date, with the maximum of accurate information (eg. taxonomy, georeferenced sites and indication of the sampling effort). Thus, the first chapter aimed to review the occurrences of bats for the Londrina region, Paraná, in the last 36 years, and compile a database with systematized information. In addition, we identified sampling gaps to help direct where and how new efforts must be employed. This review was based on published and unpublished data. The second chapter aimed to identify the patterns and predictors (environmental and spatial) of bats' beta diversity in the Atlantic Forest, at different scales of analysis. A beta diversity partitioning analysis (in nestedness and turnover) and a Generalized Dissimilarity Model were used to correlate beta diversity with the predictors. The third chapter, aimed to answer: where are the centers of greatest concentration and restriction of diversity of bats located within the limits of the Atlantic Forest hotspot? and, how much of these areas are already protected by current legislation? For this, an analysis of Geographical Interpolation of Endemism (GIE) was used and an overlay of the map of Protected Areas (PAs) on the map resulting from the GIE. The results presented in the three chapters: 1- a data base for studies on bat distribution in the region of Londrina; 2- data indicate that the Londrina region is still under-sampled and direct new sampling efforts are needed; 3- identify the more species-rich ecoregion for the Atlantic Forest for bats (coast and interior of Bahia and Serra do Mar); explaining how beta diversity varies along the biome's latitudinal and longitudinal gradient; and how the perception of this variation changes according to the scale of analysis; 4- help to predict how environmental changes can interfere in species distribution, by indicating the importance of environmental factors in the composition of bat species; 5- identification of an endemism area for 63% of bat species in the Atlantic Forest, located in the center of the biome; 6- demonstrate how GIE, using bats as a model, can be an analysis that results in a more effective strategy in planning PAs in the Atlantic Forest; and 7- indicate the urgency of creating PAs within the coastal and interior Bahia ecoregions.

Keywords: Chiroptera; database; distribution patterns; areas of endemism.

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LISTA DE ABREVIATURAS E SIGLAS

A	Amazon
AF	Atlantic Forest
A/SE	Abundance/Sampling effort
C	Carnivorous
Ca	Caatinga
CAPES	Coordenação de Aperfeiçoamento de Pessoal de Nível Superior
CNPQ	Conselho Nacional de Desenvolvimento Científico e Tecnológico
CR	Critically Endangered
DD	Data deficient
EECon Lab	Evolutionary Ecology & Conservation Lab (Universidade Estadual de Londrina)
F	Frugivorous
GDM	Generalized dissimilarity model
GIE	Geographical Interpolation Endemism
H	Hematophagous
I	Insectivorous
ICMBio	Instituto Chico Mendes da Conservação
IN	International status conservation
IUCN	International Union for Conservation of Nature's
LC	Least concern
LEPIB	Laboratório de Ecologia de Peixes e Invasões Biológicas
MZUEL	Museu de Zoologia da Universidade Estadual de Londrina
N	Nectarivorous
NI	Not included
NT	Near Threatened

O	Omnivorous
P	Piscivorous
Pa	Pampas
PA	Protected Areas
PEMG	Parque Estadual Mata dos Godoy
PMAT	Parque Municipal Arthur Thomas
RE	Regional satatus conservation
RGB	Red, green, blue color pallet
UEL	Universidade Estadual de Londrina
UC	Unidade de Conservação
VU	Vulnerable
β_{sim}	Turnover
β_{sne}	Nestedness
$\beta_{s\emptyset r}$	Total beta diversity

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APRESENTAÇÃO

O Capítulo 1 “Thirty-six years of research on bats in the region of Londrina, Brazil: what have we learned?” Encontra-se nas normas para ser submetido ao periódico “Neotropical Biology and Conservation” (Anexo A).

O Capítulo 2 “Beta diversity patterns of bats in the Atlantic Forest: How does the scale of analysis affect the importance of spatial and environmental factors?” Encontra-se nas normas e aceito no periódico “Journal of Biogeography” (Anexo B).

O Capítulo 3 “Downscaling the Atlantic Forest biodiversity hotspot: using the distribution of bats to find smaller hotspots of conservation priority”. Encontra-se nas normas para ser submetido ao periódico “Biological Conservation” (Anexo C).

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2. INTRODUÇÃO GERAL

Estudos que buscam compreender os padrões em biodiversidade têm sido cada vez mais abundantes (ex. SILVA et al., 2012; SAITER et al., 2016; ANJOS et al., 2018; LÓPEZ-AGUIRRE et al., 2019). Isso porque o aumento do impacto das atividades humanas no ambiente (principalmente fragmentação de habitat e mudanças climáticas) tem causado uma drástica defaunação quando comparado com as taxas de perda de biodiversidade animal durante eventos de extinções naturais (DIRZO et al., 2014; PIMM et al., 2014; VELLEND et al., 2017), implicando na urgência em entender os padrões de distribuição das espécies para que tomadores de decisões possam ter suporte e criar medidas visando minimizar esses impactos, e que venham a auxiliar na conservação da biodiversidade.

Com base nisso, ecologistas e biogeógrafos tentam responder, dentre outras questões, porque as espécies ocorrem em determinados locais, mas não em outros. Diversas hipóteses podem ser testadas, como a de que o ambiente impõe limitações climáticas ou de recursos (filtros ambientais), ou que as espécies evoluíram em determinados locais e não conseguiram se dispersar (DOBROVOLSKI et al., 2012; GIANUCA et al., 2017). Esta última, seja por causa da distância geográfica entre os locais ou a existência de barreiras geográficas (filtro geográfico), ou pela baixa capacidade de dispersão das espécies (filtro ecológico). Para isso é necessário entender a relação da ocorrência das espécies com o ambiente, investigando como os filtros ambientais, geográficos e ecológicos influenciam nos padrões de distribuição. Por exemplo, sabe-se que a latitude e a altitude estão relacionadas ao clima e são preditores ambientais na distribuição e riqueza das espécies (WILLIG et al., 2003; STEVENS, 2013; ARITA et al., 2014; SOININEN et al., 2018; Batista et al., 2020), sendo os locais em latitudes baixas/médias e em altitudes mais baixas, mais ricos do que as regiões mais frias em altas latitudes ou altitudes. Ainda, espécies que possuem baixa capacidade de dispersão, podem ter suas distribuições fortemente influenciadas pelos filtros geográficos (DOBROVOLSKI et al., 2012; VARZINCZAK et al., 2019).

Entender os padrões de distribuição das espécies e como os fatores ambientais, geográficos e ecológicos estruturam a biodiversidade, pode fornecer

34 importantes perspectivas para a conservação (WRIGHT, 2005; SIGRIST;
35 CARVALHO, 2008; CAÑADAS et al., 2014; BATISTA et al., 2020). Por exemplo, a
36 identificação de como as espécies estão distribuídas revela quais são endêmicas,
37 raras, ameaçadas, e onde elas ocorrem. Isso possibilita a delimitação de zonas
38 biogeográficas com biotas únicas (ex. OLSON et al., 2001), além de indicar onde
39 estão as áreas de endemismo (SIGRIST; CARVALHO, 2008; CARNAVAL et al.,
40 2014; CAMPOS; LOURENÇO-DE-MORAES, 2017; OLIVEIRA et al., 2019;
41 CASTUEIRA-OLIVEIRA et al. 2020). Também é possível prever como as espécies
42 responderiam às perturbações ambientais em seus habitats, como às mudanças
43 climáticas (SIMON et al., 2013; AGUIAR et al., 2016; LOURENÇO-DE-MORAES et
44 al., 2019). Assim, diante da escassez de investimentos para a conservação, os
45 estudos sobre padrões em biodiversidade são uma base sólida para que os
46 conservacionistas escolham estratégias mais eficientes e com melhores relações de
47 custo/benefício (SIGRIST; CARVALHO, 2008). Por exemplo, ao entender melhor a
48 distribuição de espécies possibilita a indicação de áreas prioritárias para a
49 conservação que consigam salvaguardar o máximo de espécies (de diferentes *taxa*,
50 as raras, endêmicas e ameaçadas), serviços ecossistêmicos, ou fitofisionomias
51 únicas em longo prazo (BRUM et al., 2017; CAMPOS; LOURENÇO-DE-MORAES,
52 2017; SOBRAL–SOUZA et al., 2018).

53 As ferramentas para investigar os padrões de distribuição das espécies estão
54 cada vez melhores e mais acessíveis, com bons programas gratuitos que permitem
55 medidas e modelagens de dados de biodiversidade (OLIVEIRA, 2018; SIMPSON,
56 2018; LORTIE et al., 2020). A princípio, apenas índices simples para medir a
57 riqueza, diversidade e equabilidade estavam disponíveis (veja revisão em
58 COLWELL; CODDINGTON, 1994). No entanto, as inferências que esses índices
59 permitem fazer são limitadas para explicar as diferentes escalas de diversidade (alfa,
60 gama e beta). A diversidade local (ou diversidade alfa) e a diversidade regional (a
61 soma da diversidade de todos os locais dentro de uma região, ou diversidade gama)
62 são facilmente medidas por esses índices.

63 Como resultado, conseguimos avaliar quais são as comunidades mais ricas
64 ou diversas e quão semelhantes elas são entre si em uma escala regional. Mas não
65 conseguimos analisar como a composição das espécies varia entre as comunidades
66 dentro de uma região (diversidade beta). Existem duas maneiras pelas quais as

67 comunidades podem se diferenciar e que compõe a medida de beta diversidade
68 (BASELGA; ORNE, 2012). A primeira é a substituição de determinadas espécies
69 num local por outras espécies em outro local, ou seja, ambos locais podem ter o
70 mesmo índice de riqueza, mas a composição da comunidade é diferente. A segunda
71 implica no aninhamento entre locais, ou seja, a perda ou ganho de espécies em
72 determinados locais em detrimento de outros. Baseado nisso, Baselga; Orne (2012)
73 propuseram um método que foi implementado no pacote “Betapart” (R-project)
74 capaz de particionar a beta diversidade (em substituição e aninhamento),
75 proporcionando uma análise mais eficiente que revela como as comunidades se
76 diferenciam, possibilitando identificar a escala espacial da perda ou ganho de
77 diversidade (SOCOLAR et al., 2015).

78 Aliar diferentes métodos de análises em biodiversidade podem fornecer
79 informações mais complexas ainda (ex. OLIVEIRA et al., 2019). Além de ser
80 possível identificar os gradientes de riqueza e composição de espécies entre as
81 comunidades (a beta diversidade), podemos atribuir variáveis ambientais e
82 geográficas ao gradiente espacial da análise, usando um Modelo de Dissimilaridade
83 Generalizado (GDM), por exemplo. Este modelo correlaciona as mudanças das
84 comunidades em um gradiente regional às variáveis climáticas e espaciais do
85 habitat, sugerindo uma relação das diferenças na composição das espécies entre
86 locais (FITZPATRICK et al., 2013; FITZPATRICK; LISK, 2016). Ainda, podemos
87 encontrar as áreas onde ocorre a sobreposição na distribuição das espécies,
88 indicando alta diversidade, e que possuem também espécies endêmicas (as
89 chamadas áreas de endemismo) (ex. Interpolação Geográfica de Kernel, ou GIE -
90 OLIVEIRA et al., 2015). Analisar a diversidade utilizando diferentes métricas em
91 conjunto, auxilia no mapeamento de áreas de relevância biológica e pode clarear o
92 direcionamento dos mecanismos envolvidos na conservação da biodiversidade.

93 Todavia, a correta aplicação desses métodos depende de um banco de dados
94 extenso e confiável, que contenha informação sobre a distribuição das espécies com
95 dados georreferenciados e com informações taxonômicas corretas ou ao menos que
96 permitam a conferência de *vouchers* depositados em coleções seguras e
97 estruturadas (DiCAMILLO et al., 2018). Atualmente diversos bancos de dados estão
98 disponíveis (GBIF, SpeciesLink, e até os mapas da IUCN são utilizados como
99 ocorrência de espécies) e é em parte devido a isso que tem sido possível o

100 desenvolvimento de modelos sobre padrões de biodiversidade (SANTOS et al.,
101 2010). No entanto, a falta de amostragens realizadas de forma padronizada (com o
102 emprego de esforço amostral suficiente e em áreas aleatórias e não concentradas
103 em poucos locais), as identificações taxonômicas imprecisas ou incorretas e falta de
104 precisão dos dados georreferenciados, geram lacunas nos mapas de ocorrência e
105 podem comprometer a validade dos modelos (HORTAL, 2008; MALDONADO et al.,
106 2015), dificultando o trabalho de ecólogos, biogeógrafos e conservacionistas. Com
107 base nisso, grupos de pesquisadores têm concentrado esforços em reunir os dados
108 de ocorrência das espécies, revisar criteriosamente e compilar informação em
109 bancos de dados confiáveis (Um exemplo é o conjunto de dados organizado pelos
110 pesquisadores Mauro Galetti e Milton Ribeiro para o bioma Mata Atlântica “*Atlantic:
111 Data Papers from a biodiversity hotspot*”. Disponível em:
112 [https://esajournals.onlinelibrary.wiley.com/doi/toc/10.1002/\(ISSN\)1939-
113 9170.AtlanticPapers](https://esajournals.onlinelibrary.wiley.com/doi/toc/10.1002/(ISSN)1939-9170.AtlanticPapers)).

114 A partir da compilação de dados confiáveis de ocorrência das espécies e a
115 aplicação de modelos ecológicos e biogeográficos, diversos estudos tem indicado
116 áreas prioritárias para a conservação, em diferentes regiões e escalas de análise
117 (global – CEBALLOS; EHRLICH, 2006; JENKINS et al., 2013; região neotropical -
118 SIGRIST; CARVALHO, 2008; Brasil – OLIVEIRA et al., 2019; biomas inteiros –
119 SIGRIST; CARVALHO, 2008; CARNAVAL et al., 2014; ecorregiões - CAMPOS et al.,
120 2017). Estudos que analisam a distribuição do mesmo taxa mas em diferentes
121 escalas de análise, demonstram que a escala influencia na percepção dos
122 resultados (STEVENS et al., 2007; STEVENS, 2013; LÓPEZ-GONZÁLEZ et al.,
123 2015). Quanto maior a escala de análise, mais facilmente se observa a influência da
124 distância geográfica na distribuição das espécies (VARZINCZAK et al., 2018;
125 LÓPEZ-AGUIRRE et al., 2019). Por outro lado, análises em escalas regionais,
126 principalmente pequenas regiões, podem superestimar a influência das variáveis
127 climáticas locais na ocorrência das espécies, deixando de lado fatores geográficos
128 (STEVENS, 2013; STEVENS et al., 2007; LÓPEZ-GONZÁLEZ et al., 2015). Isso
129 ressalta a importância de se investigar os padrões de diversidade em diferentes
130 escalas, aumentando a percepção da complexidade da distribuição das espécies
131 que, por sua vez, facilitará os esforços de conservação das espécies e biomas. Por
132 exemplo, atualmente são indicados 36 *hotspots* globais de biodiversidade, que são

133 áreas com alta taxa de endemismo e grande número de espécies criticamente
134 ameaçadas (MYERS et al., 2000; WILLIAMS et al., 2011; NOSS et al., 2015; CEPF,
135 2016). No entanto, essas áreas podem ser tão extensas como biomas inteiros, e a
136 falta de interesse e recursos para a conservação, junto com a exploração econômica
137 dos remanescentes florestais, inviabiliza a aplicação de regras que protejam um
138 bioma como um todo, tornando imprescindível a identificação de áreas críticas
139 dentro de uma área que já possui importância intrínseca (chamados *hotspots*
140 dentro de *hotspots*) (CAÑADAS et al., 2014).

141 A Mata Atlântica, por exemplo, é considerada um dos *hotspots* globais mais
142 ameaçados (MYERS et al., 2000; SCARANO; CEOTTO, 2015). Possui grande
143 extensão latitudinal (do nordeste do Brasil até o leste do Paraguai e nordeste da
144 Argentina), abrigando dez diferentes ecorregiões que refletem um alto gradiente de
145 diversidade (OLSON et al., 2001). Análises em escalas globais e regionais, já
146 apontaram diversas áreas de refúgio, de alta diversidade e endemidade dentro da
147 Mata Atlântica para diferentes *taxa* (CEBALLOS; EHRLICH, 2006; CARNAVAL;
148 MORITZ, 2008; SIGRIST; CARVALHO, 2008; CARNAVAL et al., 2014; CAMPOS;
149 LOURENÇO-DE-MORAES, 2017; COSTA et al., 2018), sugerindo a criação de
150 novas Unidades de Conservação e a manutenção das já existentes ao longo de todo
151 o bioma (desde a Serra do Mar, no sul, até as ecorregiões de Pernambuco, no norte),
152 dependendo do *taxa* avaliado. No entanto, a conservação da biodiversidade não tem
153 sido uma prioridade para os tomadores de decisões no Brasil, ao contrário, o
154 afrouxamento das legislações ambientais (como o novo Código Florestal, lei 12.651
155 de 25 de maio de 2012) e recentes propostas do Ministério do Meio Ambiente para
156 reduzir a proteção da Mata Atlântica (ver em:
157 [https://www1.folha.uol.com.br/ambiente/2020/04/salles-anistia-desmatadores-da-](https://www1.folha.uol.com.br/ambiente/2020/04/salles-anistia-desmatadores-da-mata-atlantica-em-meio-a-pandemia-de-covid-19.shtml)
158 [mata-atlantica-em-meio-a-pandemia-de-covid-19.shtml](https://www1.folha.uol.com.br/ambiente/2020/04/salles-anistia-desmatadores-da-mata-atlantica-em-meio-a-pandemia-de-covid-19.shtml) Acessado em 24 de abril de
159 2020), colocam em risco não apenas espécies, mas os serviços ecossistêmicos que
160 mantém o funcionamento do bioma. Isso torna urgente o direcionamento de
161 pesquisas que consigam identificar áreas prioritárias mais pontuais, para
162 salvaguardar não apenas um ou outro *taxa*, mas sim locais que representem grande
163 diversidade taxonômica e funcional de diferentes grupos. Para isso é preciso
164 escolher um grupo cuja distribuição represente também a distribuição de outros *taxa*
165 (OLIVEIRA et al., 2019).

166 Nesse sentido, morcegos podem ser um bom modelo para estudos
167 biogeográficos e conservacionistas. Dentre os mamíferos, os morcegos — com mais
168 de 1100 espécies — só perdem em diversidade para os roedores (SIMMONS, 2005)
169 sendo encontrados em quase todos os continentes (com exceção dos polos). Essa
170 diversidade não é apenas taxonômica, mas também funcional existindo espécies
171 nectarívoras, frugívoras, insetívoras, carnívoras (incluindo consumo de peixes) e
172 hematófagas, e que utilizam diferentes recursos do habitat (cavernas, troncos de
173 árvores, folhagens e rochas, em matas primárias ou secundárias) (KALKA et al.,
174 2008). Isso sugere que entender como a distribuição do grupo varia ao longo de um
175 gradiente ambiental e apontar onde estão as áreas de endemismo, pode ser um
176 método para identificar locais com uma complexa rede de interações implicando em
177 alta diversidade de diferentes grupos e que poderiam ser o foco dos esforços de
178 conservação.

179 Em vista desse cenário, essa tese foi estruturada em três capítulos. O
180 primeiro, objetivou revisar os registros de ocorrência de morcegos para a região de
181 Londrina e compilar um banco de dados com informações sistematizadas. Isso
182 porque a região é amostrada para o grupo desde a década de 80, o que gerou
183 números extensivo de dados dispersos em literatura publicada e não publicada
184 (monografias, dissertações e teses). Além disso, nesse capítulo identificamos e
185 apontamos as lacunas de amostragem a fim de ajudar a direcionar onde e como
186 novos esforços de amostragem devem ser realizados. O segundo capítulo, objetivou
187 identificar os padrões e os preditores (ambientais e espaciais) da beta diversidade
188 de morcegos na Mata Atlântica, em diferentes escalas de análise (o bioma como um
189 todo, grandes regiões, e ecorregiões dentro das grandes regiões), partindo das
190 hipóteses: H1- A substituição de espécies está associada à distância geográfica e é
191 o principal componente da diversidade beta no gradiente latitudinal; H2- O processo
192 principal que determina a dissimilaridade da composição ao longo de um gradiente
193 longitudinal (da costa para o interior) é o aninhamento; H3- As ecorregiões
194 localizadas no sul da Mata Atlântica, em latitudes mais altas e climas mais frios,
195 terão a menor riqueza e substituição de espécies de morcegos; e H4- As ecorregiões
196 localizadas em latitudes mais baixas, no centro e norte, apresentarão a maior
197 riqueza e substituição de espécies de morcegos. Finalmente, o terceiro capítulo
198 objetivou responder duas questões centrais: 1) Onde estão localizados os centros de

199 maior concentração e restrição de diversidade de morcegos dentro dos limites do
200 *hostspot* Mata Atlântica? e 2) Quanto em extensão dessas áreas já encontram-se
201 protegidas por lei?

202 Logo, os resultados apresentados nos três capítulos: 1- servem de base para
203 estudos mais complexos que queiram utilizar o banco de dados da distribuição de
204 morcegos na região de Londrina, Paraná; 2- direcionam novos esforços de
205 amostragem para morcegos na região de Londrina; 3- apontam as ecorregiões mais
206 ricas em morcegos para a Mata Atlântica, como a beta diversidade varia ao longo do
207 bioma e como a percepção dessa variação muda de acordo com a escala de
208 análise; 4- indicam a importância dos fatores ambientais na composição das
209 espécies de morcegos e que eles devem ser levados em conta em estudos
210 biogeográficos, ecológicos e conservacionistas; 5- apresenta uma área de
211 endemismo para 63% das espécies de morcegos que ocorrem na Mata Atlântica; e
212 6- demonstra como a análise de Interpolação Geográfica de Endemismo (GIE),
213 usando morcegos como modelo, pode ser uma análise que resulta em uma
214 estratégia mais eficaz no planejamento de zoneamento ecológico-econômico das
215 áreas prioritárias para conservação na Mata Atlântica.

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1 3. CAPÍTULO 1

2 **Thirty-six years of research on bats in the region of Londrina, Brazil: what have we** 3 **learned?**

4 Trinta e seis anos de pesquisas com morcegos na região de Londrina, norte do Paraná, Brasil:
5 o que sabemos?

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11 **Abstract**

12 Providing reliable databases on biodiversity is important, as they are the raw material
13 for complex studies in ecology, biogeography and conservation. The region of Londrina,
14 Paraná state (Brazil), has had its bat fauna studied for over 36 years, by a group of researchers
15 from the Laboratório de Ecologia de Mamíferos of the Universidade Estadual de Londrina.
16 This generated an extensive published and unpublished literature (monographs, master's and
17 PhD theses) on the diversity of bats in the region. We compiled and reviewed all of the bat
18 studies carried out in the Londrina region and made a database with information on richness,
19 distribution and sampling gaps. Our compilation resulted in 41 species, of which 29 occur
20 within Protected Areas. Despite the long sampling time, we found that the region still has a
21 lack of information on bat fauna for most of its forest remnants, especially in the northern
22 region. In addition, the forest fragment with the greatest number of sampling nights (Parque
23 Estadual Mata dos Godoy) is undersampled. Thus, we present a reliable database that can
24 serve as a basis for more complex studies and as a guide of where sampling effort should
25 occur in future projects with bats in the region.

26 **Keywords:** Chiroptera, Paraná state, Brazil, diversity, sampling gaps.

27 **Resumo**

28 Construir e disponibilizar bancos de dados confiáveis sobre a biodiversidade é

29 importante, pois eles são a matéria prima para estudos complexos em ecologia, biogeografia e
30 conservação. A região de Londrina, Paraná, teve sua fauna de morcegos estudada ao longo de
31 36 anos, por um grupo de pesquisadores do Laboratório de Ecologia de Mamíferos da
32 Universidade Estadual de Londrina. O que gerou uma extensa literatura publicada e não
33 publicada (monografias, dissertações e teses) sobre a diversidade de morcegos na região. Nós
34 compilamos e revisamos todos os trabalhos com morcegos na região de Londrina e
35 construímos um banco de dados com informações sobre riqueza, distribuição e lacunas de
36 amostragens. Nossa compilação resultou em 41 espécies, das quais 29 ocorrem dentro de
37 Unidades de Conservação. Descobrimos que apesar do longo tempo de amostragem, a região
38 permanece com a maior parte de seus remanescentes florestais sem informações para a fauna
39 de morcegos, principalmente na porção norte. O fragmento com maior número de noites de
40 amostragem (Parque Estadual Mata dos Godoy) está subamostrado. Dessa forma,
41 apresentamos um confiável banco de dados que pode servir de base para estudos mais
42 complexos e orientar o direcionamento do esforço de amostragem em futuros projetos com
43 morcegos na região.

44 **Palavras-chave:** Chiroptera, Paraná, Brasil, diversidade, lacunas de amostragem.

45 **3.1. Introduction**

46 Databases on the occurrence and abundance of species aim at compiling and making
47 available taxonomic data that is scattered throughout the literature, museum collections and
48 other non-published sources, such as monographs, dissertations and theses (Hortal *et al.*,
49 2007). It is undeniable that the availability of biodiversity databases has greatly advanced and
50 improved studies on ecological and biogeographical patterns (Santos *et al.*, 2010). However,
51 these online data bases (e.g. GBIF and SpeciesLink) are based on the digitalization of natural
52 history collections of different museums, where a formal control process is nonexistent, and
53 the amount of detail and quality of data is strongly dependent on the data supplier (Maldonado
54 *et al.* 2015). Therefore, it is not uncommon to find species occurrence records with erroneous
55 georeferencing or mistakes in their taxonomic identification, and it is recommended to verify
56 the data with substantial taxonomic knowledge before applying biodiversity analysis.

57 Limitations on the accuracy and precision of the data (e.g., erroneous georeferencing and
58 taxonomic misidentification) may lead to two types of errors on species distribution (Hortal,
59 2008). Error 1 leads to false positive, or the attribution of a species to a site where it does not
60 occur; Error 2 leads to false negative, or the omission of a species distribution on a site where

61 it actually occurs. The above errors addresses issues in more complex ecological,
62 biogeographic and conservation studies. For example, distribution or niche models use species
63 occurrence and predictor variables of occurrence site to build their projections (i.e.
64 Dobrovolski *et al.*, 2012; Aguiar *et al.*, 2016; Delgado-Jaramillo *et al.*, 2020; Batista *et al.*,
65 2020). Further, knowledge on the biodiversity of forest remnants is crucial to choose new
66 Priority Areas and to maintain already existing ones (i.e Sigrist and Carvalho, 2008; Campos;
67 Lourenço-De-Moraes, 2017; third chapter). Consequently, databases should be developed by
68 experts to guarantee information that is reliable, of good quality and systematized. The
69 compilation of species occurrence records into local lists, with as much information as
70 possible and revised, constitute the smallest unit of the most accessed global databases. That
71 is why local lists are important to maintain the quality of the global databases and, therefore,
72 of the more complex studies developed from them.

73 Bats rank second in mammal diversity, after rodents (Simmons, 2005). Due they are
74 distributed worldwide, to have diversified life habits (living in caves, trees, tree hollows, and
75 even human building, and which feed on fruits, nectar, insects and other invertebrates, small
76 vertebrates and blood) (Kalka *et al.*, 2008), and great dispersion capacity (Kalka *et al.*, 2008),
77 bats have been widely employed in ecological and biogeographic studies, such as species
78 distribution models (Stevens *et al.*, 2007; Stevens, 2013; López-González *et al.*, 2015;
79 Varzinczak *et al.*, 2018; Varzinczak *et al.*, 2019, Batista *et al.*, 2020), which require good
80 quality data both on taxonomic identification and georeferencing of species occurrences.

81 There are at present 182 bat species in Brazil (Nogueira *et al.*, 2018) and, at least, 63
82 species occur in the state of Paraná (Passos *et al.*, 2010; Garbino and Nogueira, 2017). Bat
83 fauna in the region of Londrina, northern of Paraná, has been investigated since 1982 by a
84 research group at the Laboratório de Ecologia de Mamíferos of the Universidade Estadual de
85 Londrina being one of the most researched location for these taxa in the state of Paraná
86 (Miretzki, 2003). These three decades of sampling efforts have had public financial
87 investment (CAPES, CNPQ and Fundação Araucária) and personnel, which only further
88 highlights the importance of retrieving this information and making it public. Therefore, in
89 this study we compile and organize all available information on published and non-published
90 data of bat species in the region of Londrina to provide a systematized and taxonomically
91 verified database. In addition, the state of Paraná is in third place in the deforestation ranking
92 of the Atlantic Forest (SOS Mata Atlântica, 2018), the north of Paraná being one of the most
93 affected and fragmented regions. This makes it urgent to increase efforts for cataloging of

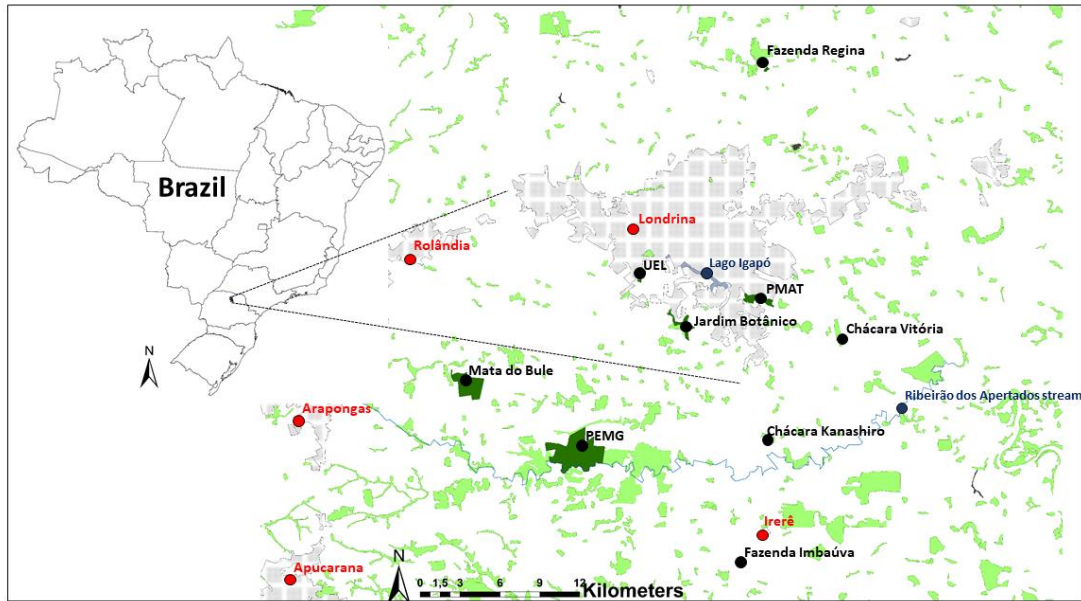
94 new occurrences of species that may not have been recorded for the region yet because of the
 95 sampling effort. Therefore, we also detect and point out information gaps to direct where and
 96 how new inventories should be undertaken.

97 **3.2. Methodology**

98 *3.2.1. Londrina region*

99 The municipality of Londrina lies in the northern region of the state of Paraná, Brazil,
 100 with an area of approximately 1,650,809 km² (IBGE, 2010). Climate (Cfa) (Köppen, 1948)
 101 features humid subtropical, with hot summers, and rainfall throughout the year, decreasing in
 102 the winter. The region belongs to the Atlantic Forest biome within the Alto Paraná ecoregion,
 103 covered by Semi-deciduous Seasonal Forest. Mean yearly rainfall is 1,888 mm and mean
 104 yearly temperature lies at 21.6° C (IAPAR, 2019). Londrina's zoning directive (Prefeitura de
 105 Londrina, 2019) acknowledges 24 green areas in the municipality, some of which have been
 106 extensively sampled for bats since 1982: non-identified valleys; the banks of the Igapó Lake
 107 on the south-eastern segment of Londrina; Jardim Botânico de Londrina (23°21'57.60"S;
 108 51°10'39.72"W) (~73 ha); the campus of the Universidade Estadual de Londrina (UEL)
 109 (23°19'19"S; 51°12'04"W) with its Horto Florestal (~10 ha). The sites are covered with
 110 secondary forest remnants, featuring native and exotic vegetation, inserted within the
 111 metropolitan region's urban matrix.

112 Several Legal Reserves within private farmland, with native and exotic secondary
 113 vegetation, have also been studied for bats. They comprise of Mata do Bule (~300 ha)
 114 (23°24'10.80"S; 51°20'13.20"W), Fazenda Regina (~6.2 ha) (23°11'1,70"S; 51° 7'39,70"W),
 115 Chácara Vitória (~1 ha) (23°22'48,52"S; 51°4'34,39"W), Fazenda Imbaúva (area unknown
 116 and without any geo-reference, in the district of Irerê - Reis et al., 1993) and Chácara
 117 Kanashiro (area unknown and without any geo-reference, between the municipality of
 118 Londrina and the district of Irerê – Reis et al., 1993). Further, Londrina has two Protected
 119 Areas (PAs) acknowledged by the Instituto Ambiental do Paraná (IAP), extensively sampled
 120 for bats. The Parque Municipal Arthur Thomas (PMAT) (23°20'39"S; 51°8'6"W) lies within
 121 the south-west urban matrix and features 85.47 ha of secondary forest, whilst the Parque
 122 Estadual Mata dos Godoy (PEMG) (23°26'53"S; 51°15'21"W), albeit with an area close to
 123 680 ha, is contiguous with other forest fragments which together make up an area of ~ 2.800
 124 ha. PEMG, within agricultural and stock-breeding culture areas, is one of the most important
 125 forest remnants in the northern region of the state of Paraná, Brazil (Figure 1).



126

127 **Figure 1.** Londrina region, Paraná, Brazil. The black dots point to areas of the region that have already
 128 been sampled for bats since 1982. The two Protected Areas in the region: PMAT – Parque Municipal
 129 Arthur Thomas (23°20'39"S; 51°8'6"W); and PEMG – Parque Estadual Mata dos Godoy (23°26'53"S;
 130 51°15'21"W). The metropolitan green areas: UEL – Campus da Universidade Estadual de Londrina
 131 (including Horto Florestal) (23°19'19"S; 51°12'04"W) and Jardim Botânico de Londrina
 132 (23°21'57.60"S; 51°10'39.72"W). The Mata do Bule (23°24'10.80"S; 51°20'13.20"W), Fazenda
 133 Regina (23°11'1,70"S; 51° 7'39,70"W), Fazenda Imbaúva (area unknown and without any geo-
 134 reference, in the district of Irerê - Reis et al., 1993), Chácara Kanashiro e Chácara Vitória, are Legal
 135 Reserves within private farmland. The location of Fazenda Imbaúva and the Chácara Kanashiro are in
 136 approximate locations, based on the map available in Reis et al., (1993), since the exact coordinates of
 137 these locations were not found.

138 3.2.2. Collection data

139 We listed the occurrences (species richness; presence) of bat species registered in
 140 Londrina between 1982 and 2018 throughout the scientific literature and non-published data.
 141 The latter were retrieved from field notes, monographs and dissertations at the Laboratório de
 142 Ecologia de Mamíferos of the Universidade Estadual de Londrina (Table 1 in pag. 87). The
 143 first step consisted of verifying all records and establishing filters for the exclusion of false
 144 occurrences. Exclusion criteria comprised lack of voucher specimen deposited in the Museu
 145 de Zoologia (MZUEL) or at the Laboratório de Ecologia de Mamíferos of the Universidade
 146 Estadual de Londrina; expected geographic distribution of species based on the literature; and
 147 the lack of precise site where the species was captured.

148 A Venn-Euler diagram was used to illustrate the occurrence of bat species in the
149 region of Londrina. Species taxonomy followed updated and commented list of Brazilian bats
150 (Nogueira et al., 2014). Conservation status of each species was based on international
151 (IUCN, 2020), national (ICMBio/MMA 2018) and regional lists “Livro vermelho da fauna
152 ameaçada no Estado do Paraná” (Margarido and Braga, 2004).

153 3.2.3. Sample gaps in the region of Londrina

154 Data were tabulated in our database (Table 1-3 in pag. 87). In this database we
155 compiled information about the number of capture nights (the only common information
156 among published and non-published data available for the sampling effort) and the
157 environments where mist nets were installed in each sample site. Therefore, we evaluate
158 whether sampling effort was heterogeniously distributed in the different environments within
159 the forest fragments.

160 Using ArcGIS 10.5 from ESRI (ArcGIS trial), a 30 km² buffer from downtown
161 Londrina was constructed. The distance was large enough to comprise all of the sampled
162 forest fragments for bats in the region. Based on this vector layer, the area (ha) of all
163 fragments located within a 30 km²-radius was added, and only for the fragments that had an
164 exact georeferenced location (namely, PEMG, PMAT, Mata do Bule, Jardim Botânico, Horto
165 Florestal in UEL and Fazenda Regina). This allowed us to calculate the percentage of sampled
166 area for the 30 km²-radius. In the case of non-sampled fragments, we employed the ArcGis
167 option for the classification of the vector layer according to certain characteristics. The
168 variable ‘fragment size’ coupled with the use of six classes was selected. Non-sampled
169 fragments were divided into: smaller than 30 ha; between 30 and 48 ha; between 48 and 70
170 ha; between 70 and 114 ha; between 144 and 196 ha; between 196 and 504 ha. This allowed
171 us to locate and determine the number of fragments in each class.

172 3.3. Results

173 Forty-one bat species were listed, distributed into four families, nine subfamilies and
174 27 genera. Records were obtained from ten georeferenced sites and other green areas in the
175 Londrina (non-identified valleys) and human building (sites without exact indication), in 577
176 sampling nights since 1982. All of the bat guilds (nectarivores, frugivores, insectivores,
177 carnivores, omnivores, hematophages and piscivores) are present (Table 2 in pag. 90). We
178 excluded only one record - *Uroderma bilobatum* - previously listed for the region of

179 Londrina, according the exclusion filter (lack of voucher specimen and expected geographic
180 distribution of species based on the literature - Garbino and Nogueira, 2017).

181 The PEMG and PMAT harbor, respectively, 68.3% and 41.50% of the bat fauna in the
182 region of Londrina. Seven out of the 12 species which were not recored in the two PAs
183 (Figure 2) were listed only in the green areas of the metropolitan region; two species were
184 registered on private farms and other two species in human building; there was no
185 environmental information of locality indication of where *Eumops perotis* was captured in the
186 metropolitan region of Londrina (Table 2 in pag. 90). Four species in our list had a certain
187 degree threat, with *Diaemus youngii* classified as “critically endangered” (Margarido and
188 Braga, 2004) *Diphylla ecaudata*, *Chrotopterus auritus* and *Chiroderma doriae* being listed as
189 “vulnerable (Margarido and Braga, 2004); and *Myotis ruber* as “near threatened” (IUCN,
190 2020).

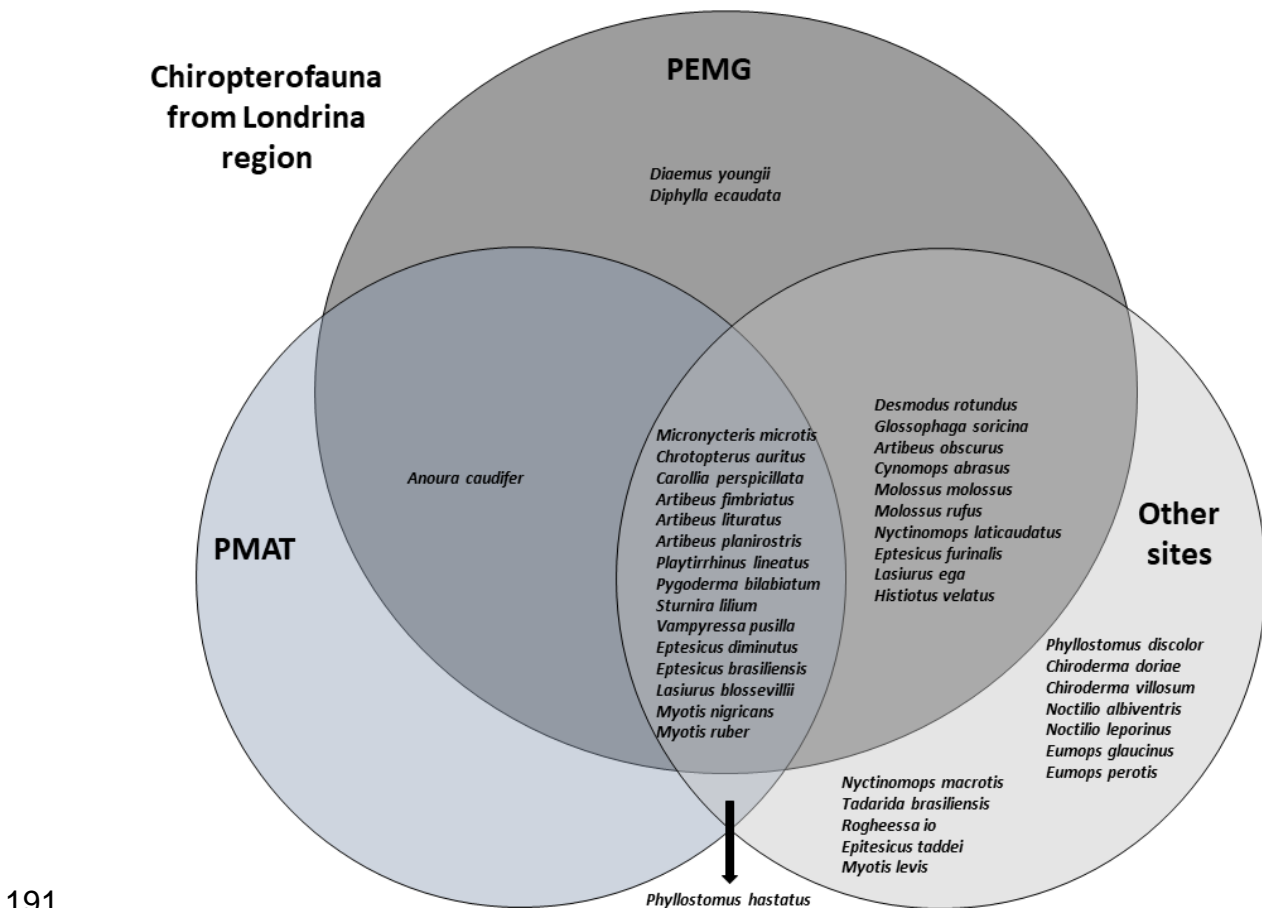
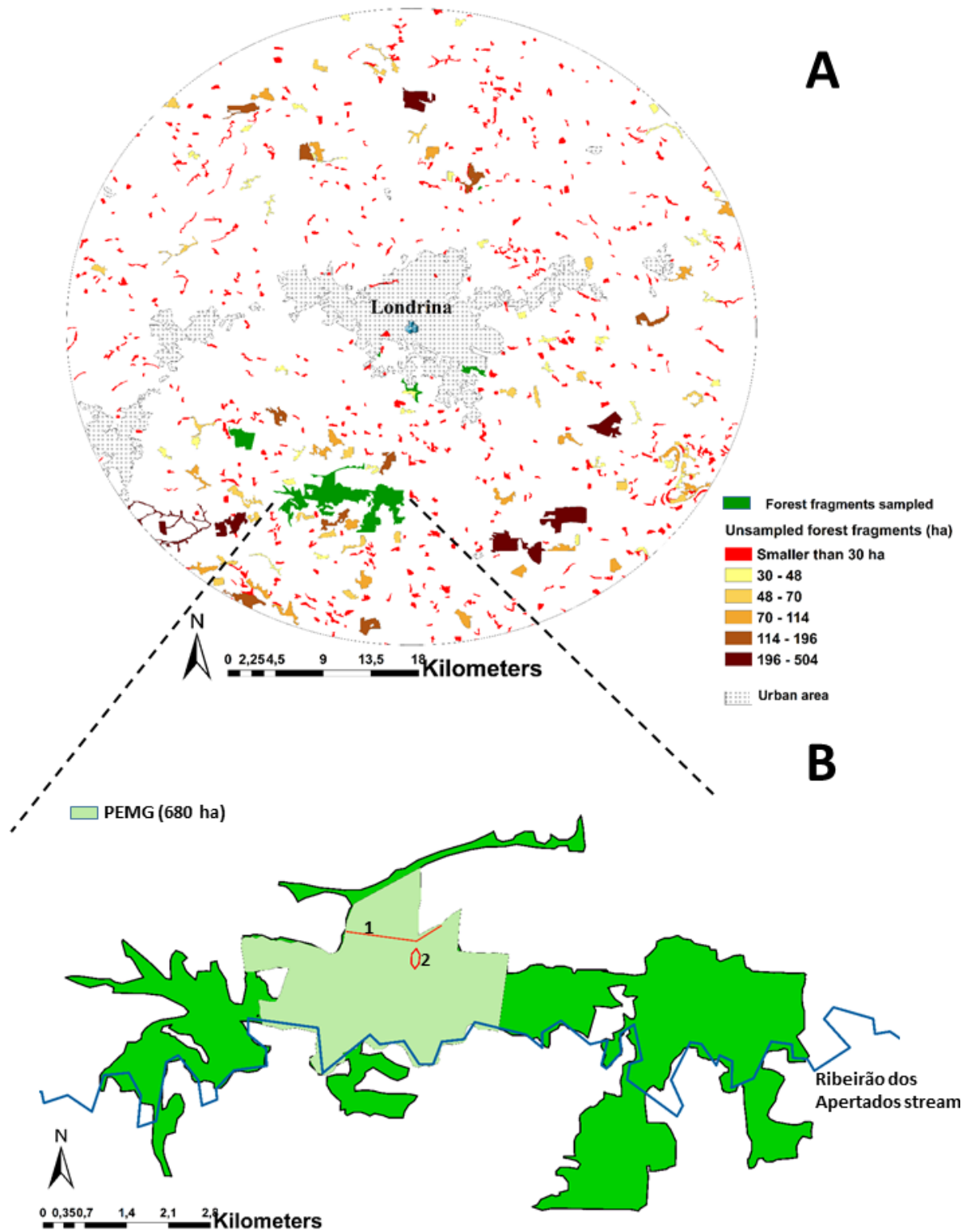


Figure 2. Venn-Euler diagram to illustrate occurrence of bat species in the region of Londrina, Paraná, Brazil. Two Protected Areas - Parque Estadual Mata dos Godoy (PEMG) and Parque Municipal Arthur Thomas (PMAT). “Other sites” indicate Legal Reserves (Mata do Bule, Fazenda Imbaúva, Chácara Kanashiro, Chácara Vitória and Fazenda Regina), the campus of Universidade Estadual de

Londrina (UEL) (including the Horto Florestal); the Jardim Botânico; banks of the Igapó Lake; other green areas in the municipality (non-identified valleys) and human building (sites without exact indication). The arrow indicates where *Phyllostomus hastatus* is in the diagram, within the “PMAT” and “Other sites” set.

192 We managed to retrieve information on the environment of captured sites for five
193 forest fragments in the region of Londrina: PEMG, PMAT, UEL, Jardim Botânico and Mata
194 do Bule. In the case of PEMG, data are not available in the published literature and they were
195 provided by researchers from the Laboratório de Ecologia de Mamíferos of the Universidade
196 Estadual de Londrina. Bat mist-nets in the park were placed along two trails (Trilha das
197 Perobas and Trilha dos Catetos) during 238 sampling nights. PMAT was sampled for 106
198 nights, while the UEL campus was sampled for 133 nights. Mist-nets in both areas were
199 distributed throughout the area. In the Jardim Botânico, captures were undertaken for 16
200 nights in several places, but was restricted to the eastern region where the area is open to
201 visitors. Mist-nets were installed in the Mata do Bule at random, in all regions of the forest
202 fragment (north, south, east and west), with a total of 14 sampling nights. Table 3 (in pag.
203 100) gives all available data regarding sampling effort and locality of mist-nets.

204 Spatial analysis revealed that less than 4% of forest fragments in the region of
205 Londrina, within a 30 km radius, were sampled for bats. Moreover, most of the sampling
206 (with the exception of Fazenda Regina) took place in southern part of the region of Londrina.
207 There are seven fragments with an area between 196 and 504 ha, of which five are greater
208 than 300 ha; nine fragments between 196 and 114 ha; 21 fragments between 114 and 70 ha;
209 27 fragments between 70 and 48 ha; 37 between 48 and 30 ha; and several fragments with less
210 than 30 ha. All of them have yet to be sampled for bats (Figure 3A).



211

Figure 3. A) The 30 km² buffer from downtown Londrina, Paraná, Brazil. Green forest fragments are the areas that have already been sampled for bats. Unsampled fragments are classified in different colors depending on the size of the area in hectares: smaller than 30 ha; between 30 and 48 ha; between 48 and 70 ha; between 70 and 114 ha; between 144 and 196 ha; between 196 and 504 ha. B)

Highlights the largest forest fragment in the region of Londrina (~ 2000 ha), where the Parque Estadual Mata dos Godoy (PEMG) (23°26'53"S; 51°15'21"W) (which has 680 ha) is located. The figure also indicates the Ribeirão dos Apertados stream and the two visitor's track where bats were captured since 1982: Trilha dos Catetos (1) and Trilha das Perobas (2).

212 3.4. Discussion

213 Our review demonstrated that the region of Londrina has at least 65% of the bat fauna
214 found in the state of Paraná (Passos et al., 2010; Garbino and Nogueira, 2017), and 43% of the
215 species found in the Atlantic Forest (Muylaert et al., 2017). However, most of the Londrina
216 region has never been inventoried and the region may have larger species richness than
217 currently registered. *Uroderma bilobatum* was not included, since no voucher specimen has
218 been deposited in the MZUEL or in the Laboratório de Ecologia de Mamíferos of the
219 Universidade Estadual de Londrina. Exclusion of *U. bilobatum* was also recommended by
220 Garbino and Nogueira (2017) on the basis of the species' natural history and the lack of
221 voucher specimens deposited in public records. Therefore, the occurrence of this species
222 should be excluded from reports of bat fauna for the state of Paraná. *Eumops perotis* should
223 also be excluded due to the criteria used in this study (no voucher specimen has been
224 deposited in the Zoological museums of the region). However, see Suckow et al. (2010) - the
225 source for this species' registration for Londrina – and Figure 4. Briefly, the specimen was
226 captured and forwarded to the Laboratório de Ecologia de Mamíferos of the Universidade
227 Estadual de Londrina on August 3, 2004 by an unknown person on the recommendation of the
228 health secretary of Londrina (Lima, I. P. personal communication).

229 The metropolitan green areas and Legal Reserves (rural forest areas) are relevant sites
230 to maintain bat diversity in the Londrina region. In fact, when they were analyzed together,
231 they harbored 95% of the bat species. The two PAs (PEMG and PMAT) are sites with high
232 richness and a safe haven for all species, including threaten species. The PEMG is the sole
233 occurrence site for two threatened hematophageous species, *D. ecaudata* and *D. younggii*.
234 However, our results have shown that despite the great sampling effort in the region, there are
235 still many sampling gaps. The region's biggest forest fragment comprises 2000ha, but
236 samplings are limited to PEMG, which has an extension of only 680 ha within this 2000 ha
237 forest fragment (Figure 3B). PEMG is one of the main forest remnants in the state of Paraná
238 and one of the main genetic forest reserves in south Brazil (IAP, 2015), with the presence of
239 different micro-habitats (Willrich et al., 2019). In spite of the great number of sampling nights
240 (238) in the area, the sites where mist-nets were installed are mainly in the visitors tracks

241 located in the northern part of the park (Figure 3B). The southern region on the banks of the
242 Ribeirão dos Apertados stream has not been systematically sampled. There were only two
243 sampling nights with zero specimen deposited in collections and without any field entries
244 (Lima, I. P. personal communication). The region lies along an elevated gradient associated
245 with changes in vegetation structure (Willrich et al., 2019), which with the water body
246 provides a different set resources when compared to the northern part of the PEMG. For
247 example, resources for *Noctilio leporinus*, a piscivore bat species, which has never been
248 captured in the park. The above suggests that there are probably more bat species in the park
249 area which have never been registered.

250 Smaller and and unprotected fragments, including metropolitan green areas, can also
251 be important for the maintenance of bat diversity in the region of Londrina. However, these
252 areas have low or no sampling, which suggests that the Londrina region may actually have a
253 higher richness of bat species richness. A good example is the Mata do Bule Legal Reserve.
254 In only 14 sampling nights, a total of nine species were recored, including the near threatened
255 species *M. ruber* (Solaris, 2019). Further, 96% of the area covered by forest fragments have
256 never been sampled, while the remaning 4% has an uneven distribution of sampling nights.
257 For example, only one forest fragment (Fazenda Regina) has been sampled in the northern
258 region of Londrina; all the other forest fragments lie in the southwestern region of Londrina.
259 Moreover, the southeastern region contains three of the five largest fragments (over-300ha) in
260 a 30km radius that have never been sampled (Figure 3A), showing the huge sampling gap for
261 the region of Londrina.

262 There is a predominance of data gaps in biodiversity independent of taxa, and because
263 acquisition of biodiversity data is based on local taxonomic surveys, these gaps accumulates
264 at increasing scales (Hortal et al., 2007). A recent review on species distribution of
265 vertebrates, arthropods and angiosperms in Brazil showed that the greatest sampling effort is
266 concentrated in easily reached places, such as PAs and forest fragments close to roadways
267 (Oliveira et al., 2017). Further complicating matters is the fact that even the current
268 knowledge on biodiversity within PAs is still scanty, with less than 1% of the area being
269 sampled for vertebrates, arthropods and angiospers in Brazil (Oliveira et al. 2017).

270 In the case of bats, there is still an additional issue with geographic sampling gaps. It is
271 very common for bat inventories to be carried out using mist-nets, which may limit the
272 detectability of bat families that are less prone to be sampled by this method, such as

273 Molossidae, Embalonuridae and Vespertilioniae (Simmons and Voss 1998, Carvalho et al.,
274 2013). For instance, there are six insectivore species (*Eumops glaucinus*, *Eumops perotis*,
275 *Nyctinomops macrotis*, *Tadarida brasiliensis*, *Epitesicus taddei*, *Rogheessa io* and *Myotis*
276 *levis*), which have not been captured within the PAs of the region of Londrina. Lack of
277 precision on the localities where these species were captured only allows us to assert that
278 these species were captured within the metropolitan area. Urban areas usually present an
279 abundance of insects that can be easily caught (insects flying around illuminated objects) and
280 roosting areas (human buildings), which attracts insectivore bats (Pacheco et al., 2010).
281 Moreover, it is easy to capture insectivorous bats in urban environment by actively searching
282 for day roosting areas and/or placing mist-nets close to their exit. It is highly probable that the
283 absence of occurrence records for these species within PAs and in the legal reserves of the
284 region to be a type 2 Error - the false absence of species in places where they may actually
285 occur.

286 Bat capture undertaken merely by mist-nets in forest fragments may increase type 2
287 error for insectivore bats of the families Molossidae, Vespertilionidae and Embalonuridae.
288 Bats from these families usually fly very high and have a more efficient eco-localization
289 system, making their capture with mist-nets a difficult task (Nogueira et al., 1999; Nogueira et
290 al., 2008). This sample bias can lead to information gaps, for example, six species of
291 insectivorous bats on our list (*Cynomops abrasus*, *Eptesicus diminutus*, *Eptesicus taddei*,
292 *Rogheessa io*, *Histiotus velatus* and *Myotis ruber*) of the Molossidae and Vespertilionidae
293 families, are categorized as "data deficient" by at least one of the lists that assessed the degree
294 of threat to species (Table 2 in pag. 90). Therefore, it is imperative to sample bats in the
295 Londrina region with alternative and complementary methods (Bergallo et al., 2003) because
296 this will surely increase the number of species recorded for the area. Methods include
297 searching for roosting sites during the day (Simmons and Voss, 1998), the construction of
298 artificial roost to attract bats (bat houses) (Tuttle and Donna, 1993), the monitoring of eco-
299 localization callings to allow the identification of species by ultrasound (Arias-Aguilar et al.,
300 2018), and the installation of canopy mist-nets (Carvalho and Fabián, 2011).

301 Our results are based on reviewing the scientific literature and careful compilation of
302 several small monitoring projects in Londrina since 1982, which allowed us to determine the
303 bat species that occurs in the area and the sampling gaps. Besides being a reliable dataset, our
304 review also provides the basis for more complex ecological studies, as well as guidance to
305 where future bat surveys should be conducted. Lastly, we have shown the conservation

306 relevance of several forest fragments of different types (e.g., urban green areas, Legal
307 Reserves and Protected Areas) in the landscape should be underscored to maintain bat
308 diversity in the region of Londrina.



316
317 **Figure 4.** Specimen of *Eumops perotis* received by the Laboratório de Ecologia de Mamíferos of the
318 Universidade Estadual de Londrina on August 03, 2004 by an unidentified person. The animal was
319 alive, was identified and released. This record is published at Suckow et al. (2010). Picture by Isaac
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1 4. CAPÍTULO 2

2 **Beta diversity patterns of bats in the Atlantic Forest: How does the scale of analysis affect the** 3 **importance of spatial and environmental factors?**

4 **Short running title: Beta diversity of bats in the Atlantic Forest**

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17 **ABSTRACT**

18 **Aim**

19 Environmental and spatial factors are broadly recognized as important predictors of beta diversity
20 patterns. However, the scale at which beta diversity patterns are evaluated will affect the outcoming
21 results. For example, studies at larger scales will usually find spatial processes as the main predictor of
22 beta diversity patterns. In this study we evaluate how beta diversity patterns change when analyses are
23 conducted at different scales by reducing the scale of analysis in a hierarchical manner.

24 **Taxon**

25 Chiroptera.

26 **Location**

27 Atlantic Forest biome.

28 **Methods**

29 Information on the occurrence of 59 bat species were obtained from the Atlantic Bats and Species
30 Link database. We partitioned beta diversity into its two components (nestedness and turnover) and
31 calculated these indexes hierarchically: the biome in its entirety (all ecoregions); between larger
32 regions (north, central and south); and between ecoregions within each region. We performed a
33 Generalized Dissimilarity Model (GDM) to identify and predict the turnover of bat species in the
34 Atlantic Forest based on geo-climatic predictors. We obtained 19 geo-climatic data from AMBDATA,
35 an environmental dataset based on different data sources commonly used in species distribution
36 modeling.

37 **Results**

38 We found that turnover was the main component influencing a latitudinal gradient when the biome
39 was analysed in its entirety. However, when the scale of the analysis was reduced, we found that
40 species loss (nestedness component) had a large effect in determining beta diversity dissimilarity. We
41 also found that nestedness was the main pattern explaining beta diversity dissimilarity along a
42 longitudinal gradient.

43 Main conclusions

44 Beta diversity patterns changed with the scale of analysis, which indicates that bat species composition
45 does not follow the same pattern throughout the Atlantic Forest. This corroborates the importance of
46 analysing beta diversity patterns at different scales in order to understand how environmental
47 dissimilarity across geographic space can influence species distribution patterns.

48 **Keywords:** biogeography, Chiroptera, tropical forest, species dissimilarity, ecological predictors,
49 ecoregions.

50

51 4.1. INTRODUCTION

52 Identifying patterns of species distribution and the potential factors involved in the shaping of
53 biodiversity are central for the understanding of the evolutionary and macroecological processes
54 influencing the structuring of communities (Leigh, O'Dea, & Vermeij, 2014). For example, by
55 exploring the way in which the spatial divergence of taxa is associated with ecological and climatic
56 gradients, allows the identification of potential filters (e.g., dispersal and environmental filters)
57 involved in the shaping of biodiversity (Dobrovolski, Melo, Cassemiro, & Diniz-Filho, 2012; Gianuca,
58 Declerck, Lemmens, & De Meester, 2017). However, the scale at which biodiversity is considered will
59 influence the perception of species distribution patterns and their drivers (McGill, Dornelas, Gotelli, &
60 Magurran, 2015). If the analysis is limited to measuring only local diversity (alpha diversity) without a
61 regional analysis, inferences will be limited to identifying sites with higher biological diversity and at
62 best the environmental and climatic factors associated with species richness gradients (Stevens &
63 Willig, 2002; Dobrovolski et al., 2012). Conversely, comparing inter-site differences in species
64 composition (beta diversity) at larger scales, allows for the evaluation of the effect of ecological
65 gradients on biodiversity patterns, which provides a better understanding of the potential mechanisms
66 involved in the distribution of species at a larger scale.

67 The regional component of biodiversity can be included by performing a beta diversity
68 analysis, which concerns variation in species composition among different sites (Whittaker, 1960;
69 Chao & Ricota, 2019). Beta diversity can be explained by two distinct components: nestedness and
70 species turnover. Nestedness is characterized by the loss or gain of species, while species turnover is
71 the replacement of species along an environmental gradient owing to historical processes (Baselga &
72 Orme, 2012; Socolar, Gilroy, Kunin, & Edwards, 2016; Coelho, Carneiro, Branco, Borges, &
73 Fernandes, 2018). Environmental dissimilarity should be greater for spatially large regions. For
74 example, biomes are not homogeneous throughout their extent, with variations in topography and
75 microclimate creating environments with different conditions and resources, which in turn can affect
76 the structuring of communities at the regional scale. The partitioning of total beta diversity into the
77 nestedness and turnover components allows a more refined understanding of how these factors affect
78 biological communities (Rosauer et al., 2014; Legendre, 2014; Soininen et al., 2018) and enables the
79 testing of potential mechanisms. If species are responding to environmental dissimilarities, then a
80 greater turnover of species will be expected for regions that have greater environmental dissimilarity.
81 It is well-established that climate patterns are associated with latitude; they can be used as a proxy for
82 environmental heterogeneity, and are associated with species richness gradients (Willig, Kaufman, &
83 Stevens, 2003; Soininen et al., 2018). Consequently, a greater turnover of species is expected for
84 regions located at lower latitudes, whereas nestedness should be greater for regions located at higher
85 latitudes.

86 Several studies have found that spatial distance and environmental differences between sites
87 are the main factors explaining beta diversity dissimilarity along a geographical gradient (Melo,
88 Rangel, & Diniz-Filho, 2009; Rodriguez-Artigas, Ballester, & Corronca, 2016; Hu et al., 2018; Coccia
89 & Fariña, 2019). Species' life history traits are important predictors in explaining beta diversity
90 patterns. For example, communities comprised of species with limited dispersal capabilities are more
91 likely to be influenced by spatial distances (Dobrovolski et al., 2012; Varzinczak, Moura, Lima, &
92 Passos, 2019). Conversely, highly vagile species should be able to follow their environmental niche,

93 being more affected by climatic and environmental factors and less so by spatial dynamics
 94 (Dobrovolski et al., 2012; Varzinczak, Lima, Moura, & Passos, 2018). However, recent studies on bat
 95 communities (i.e., species with high dispersal capability) along a geographical gradient have yielded
 96 contradictory results. While some studies have shown that differences in communities are mainly due
 97 to different environmental preferences (Stevens, López-González, & Presley, 2007; Stevens, 2013;
 98 López-González, Presley, Lozano, Stevens, & Higgins, 2015), others have shown that the spatial
 99 pattern was more effective in explaining bat diversity (Varzinczak et al., 2018) or both factors
 100 (environmental and geographical) influenced similarity in community composition (Varzinczak et al.,
 101 2019). A potential explanation for this controversy could be the scale of analysis. For instance, studies
 102 that evaluated bat diversity considering a larger scale, such as the neotropical region, showed that the
 103 spatial patterns was the most influential factor explaining differences in the diversity of bats
 104 (Varzinczak et al., 2018; López-Aguirre, Suzanne, Laffan, & Archer, 2019), whereas studies on a
 105 smaller scale, covering a single biome, demonstrated that climate was the most influential factor
 106 explaining bat species diversity (Stevens, 2013; Stevens et al., 2007; López-González et al., 2015). It
 107 is possible that spatial patterns will be more important at a larger scale because this increases the
 108 chances of including geographical barriers (Varzinczak et al., 2018) that are capable of imposing
 109 stronger physiological and metabolic constraints (Zuloaga & Kerr, 2017).

110 Contradictory results, such as the ones presented above for bat species, lead to questions such
 111 as “how do species distribution patterns change when the analyses are conducted at different spatial
 112 scales?” One should expect spatial factors to have a higher influence in explaining species diversity as
 113 one increases the scale of analysis, but should we also expect to find changes in the processes
 114 (nestedness or turnover) explaining beta diversity differences? Because of the growing availability of
 115 data papers in ecology, particularly those regarding species distribution data, it is now possible to test
 116 the above hypotheses. Recently, a collection of species occurrence data from the Atlantic Forest has
 117 been made available. Known as “Atlantic: Data Papers from a biodiversity hotspot,” these datasets are
 118 organized by researchers Mauro Galetti and Milton Ribeiro, and all papers have been reviewed by
 119 specialists, one of which consists of a dataset for bat species (Muylaert et al., 2017).

120 In addition to the availability of these data, other factors make the bats of the Atlantic Forest
 121 biome an important topic for study. First, bats are distributed almost everywhere in the world (except
 122 at the poles) and have diverse life-histories and great dispersion capacity (Kalka, Smith, & Kalko,
 123 2008). Secondly, the Atlantic Forest has a highly endemic biodiversity and is one of the most
 124 threatened ecosystems in the world (Scarano & Ceotto, 2015). This biodiversity hotspot suffers from
 125 strong landscape conversion, with remaining forest fragments often being too small for species
 126 persistence (Scarano & Ceotto, 2015). The biome covers the north of the Brazilian coast to west of
 127 Paraguay and Argentina (Ribeiro, Metzger, Martensen, Ponzoni, & Hirota, 2009). Because of its large
 128 latitudinal range (> 30 degrees), the Atlantic Forest includes a diverse mosaic of habitats with a rich
 129 biodiversity. It is currently classified into ten distinct ecoregions (biogeographic zones) (Olson et al.,
 130 2001), which were delimited on the basis information concerning both resident taxa and geo-climatic
 131 data. Because species respond to environmental and spatial gradients, the ecoregions should present a
 132 correlation between beta diversity components and geographic and environmental factors (Stevens,
 133 2013; Anjos et al., 2018; López-Aguirre et al., 2019).

134 Considering the different interpretations of distribution patterns of bat species (Stevens et al.,
 135 2007; Stevens, 2013; López-González, 2015; Varzinczak et al., 2018; Varzinczak et al., 2019) and
 136 the availability of a data set on the occurrence of bat species within the Atlantic Forest (Muylaert et
 137 al., 2017), we evaluated the association between bat community (beta diversity) composition and
 138 environmental and geographical variables. We did this hierarchically, first for the entire extent of the
 139 biome, and then by dividing it into three regions: north, centre and south. In table 1 we provide a
 140 summary of our hypotheses and their main predictions.

141 4.2. MATERIALS AND METHODS

142 4.2.1. The Atlantic Forest biome

143 Our study covers the Atlantic Forest, the second largest South American rainforest and one of

144 the world's biodiversity hotspots. The biome covers the north of the Brazilian coast to the east of
 145 Paraguay and Argentina (Ribeiro et al., 2009). It embraces a diverse mosaic of habitats with a rich
 146 biodiversity and was classified by Olson et al. (2001) into ten distinct ecoregions (Fig. 1a.). In the
 147 southern region: (1) Alto Paraná Atlantic forests (Alto Paraná), composed by deciduous and semi-
 148 deciduous forests; (2) Moist *Araucaria* pine forests (*Araucaria*); (3) Serra do Mar coastal forests (Serra
 149 do Mar), which consists of ombrophilous dense forest and (4) Campos Rupestres montane savanna
 150 (Campos Rupestres), which are high-elevation grasslands. In the central region: (5) Bahia coastal
 151 forests (Bahia Coastal) and (6) Bahia interior forests (Bahia Interior), both of which are semi-
 152 deciduous forests; and (7) Atlantic dry forests (Dry Forests), which are deciduous forests with
 153 enclaves of Cerrado savanna. In the northern region: (8) Pernambuco coastal forests (Pernambuco
 154 Coastal) and (9) Pernambuco interior forests (Pernambuco Interior), both of which are a mix of
 155 ombrophilous dense forest and ombrophilous open forest. Lastly, (10) Atlantic Coast restinga (Coast
 156 Restinga), which are *restingas* and mangroves that extend throughout the coast of the biome. The Dry
 157 Forests ecoregion was excluded from this study due to a lack of data for bat occurrences there.

158 4.2.2. Occurrence and geo-climatic data

159 Our study was based on 2,634 bat occurrence data points for 526 sites (coordinates) within the
 160 Atlantic Forest. Our data came from two sources. We extracted 1,795 occurrence data points for 160
 161 sites from Muylaert et al. (2017). This dataset was compiled by bat specialists who also reviewed the
 162 taxonomy of the species and the coordinates of sampling sites. We then used the species list provided
 163 in Muylaert et al. (2017) to search for other occurrence records in speciesLink (data downloaded from
 164 <http://splink.cria.org.br/>). We obtained 838 occurrence records of bat species for 365 sites. We
 165 reviewed the dataset obtained from speciesLink according to how reliable the information was
 166 regarding: i) coordinates and site correspondence (we used google maps to check if the coordinates
 167 were referring to the places indicated), ii) correct taxonomy (we excluded species with “sp”, “ssp”,
 168 “cf” and “aff”), and iii) voucher specimens (we only considered records with specimens that were
 169 deposited in a museum). We also included a single occurrence record of *Natalus macrourus* (Trajano,
 170 1984) from Parque Estadual Turístico do Alto Ribeira (PETAR), which was not considered by
 171 Muylaert et al. (2017) or SpeciesLink. Occurrence records belonging to the bat families Molossidae,
 172 Vespertilionidae and Emballonuridae were not included in our study because they are seldom captured
 173 in mist-nets (Nogueira, Pol & Peracchi, 1999; Nogueira, Pol, Monteiro & Peracchi, 2008), which was
 174 the predominant method used for sampling bat species represented in our data sources.

175 We obtained geo-climatic data from AMBDATA (available at
 176 <http://www.dpi.inpe.br/Ambdata/index.php>). The AMBDATA is an environmental dataset
 177 systematized from different data sources and commonly used in species distribution modelling. It
 178 consists of 19 bioclimatic variables at 30 arc-sec resolution (approx. 1 km). These are: 1) annual mean
 179 temperature (°C); 2) mean diurnal range (°C); 3) isothermality (mean diurnal range divided by annual
 180 temperature range, and multiplied by 100); 4) temperature seasonality (standard deviation *100); 5)
 181 maximum temperature of warmest month (°C); 6) minimum temperature of coldest month (°C); 7)
 182 temperature annual range (°C); 8) mean temperature of wettest quarter (°C); 9) mean temperature of
 183 driest quarter (°C); 10) mean temperature of warmest quarter (°C); 11) mean temperature of coldest
 184 quarter (°C); 12) annual precipitation (mm); 13) precipitation of wettest month (mm); 14) precipitation
 185 of driest month (mm); 15) precipitation seasonality (coefficient of variation); 16) precipitation of
 186 wettest quarter (mm); 17) precipitation of driest quarter (mm); 18) precipitation of warmest quarter
 187 (mm); and 19) precipitation of coldest quarter (mm). We also included three non-climatic
 188 environmental variables from AMBDATA: 1) tree cover at a 500 m resolution (percentage); 2)
 189 elevation (m) at 3 arc-sec horizontal resolution (about 90 m) and a vertical resolution of 1 m and
 190 lastly, 3) declivity (degrees) generated from the elevation grid.

191 4.2.3. Beta diversity and the Generalized Dissimilarity Model (GDM)

192 There are various dissimilarity indices to measure changes in species composition between
 193 assemblages. We used the Sorensen index (β_{SOR}) as implemented in Betapart package – R-project
 194 (Baselga & Orme, 2012). The input data table consists of the presence and absence of bat species for
 195 each study site (latitude and longitude). The package computes the total dissimilarity across all sites,

196 and calculates turnover (Simpson's index, β_{sim}) and nestedness (the difference between the Sorensen
197 and Simpson index, β_{sne}) components. Betapart returns cluster and dissimilarity matrices (between
198 pairwise sites, and pairwise matrices of shared and non-shared species between sites) of turnover and
199 nestedness.

200 First, we computed total beta diversity and its two components, nestedness and turnover,
201 among the ten Atlantic Forest ecoregions proposed by Olson et al. (2001). Then we used the cluster
202 and GDM results (Fig. 1a and 1b) to split the Atlantic Forest into three larger regions (southern,
203 central and northern). Lastly, nestedness and turnover were calculated among the ecoregions making
204 up each of the three regions. Each region was treated separately.

205 We used a species presence data frame with the coordinates of the occurrence sites to perform
206 Generalized Dissimilarity Modeling (GDM), which analyses spatial patterns of dissimilarity in species
207 composition pairwise between sites, using a nonlinear regression matrix. GDM quantifies dissimilarity
208 using the Soresen Index (total beta diversity), then associates the turnover component (β_{sim}) with
209 biological distance (predictor variables) between sites (Fitzpatrick et al., 2013). The GDM procedure
210 was used to predict bat species turnover across the Atlantic Forest based on environmental data. We
211 used the R package "GDM" (Fitzpatrick & Lisk, 2016) to fit a GDM with the 22 environmental
212 variables and the geographical distance (decimal degrees) between occurrence sites. The latter was
213 calculated using the option "geo=T" in the function "gdm" of the GDM package. We used the
214 parameter "weightType= richness" to weight sites relative to the number of species to minimize
215 sampling bias. We chose not to exclude sites with few species (i.e. less than five) because in our
216 database over 360 occurrence sites had five or fewer species, 90 sites had 10 species or fewer. Only 11
217 sites recorded 50% or more of the total number of species, so that excluding sites with few species
218 would lead to a significant loss of data. Patterns of species turnover can be visualized on a raster with
219 RGB colour standards; areas with similar colours contain similar assemblages. The GDM matrix
220 regression used was I-spline with three basic functions, meaning that we used three points (the
221 minimum) to form the I-spline curve (Fitzpatrick & Lisk, 2016). I-Splines can be visualized in a graph
222 showing the relationship of predicted biological distance versus observed biological distance,
223 providing an indication of how species composition changes along each environmental gradient
224 (Fitzpatrick & Lisk, 2016). The selection of the best subset of predictors for our model followed
225 Williams, Belbin, Austin, Stein, & Ferrier (2012): the initial model included all predictors; variables
226 that contributed less than 2% to model explanation were iteratively removed. Variable removal was
227 done on a stepwise basis beginning with the elimination of the variable that contributed the least to
228 model explanation. Variables were reassessed regarding their importance and significance during each
229 step of model reduction (i.e., backward elimination). Our model started with 23 predictor variables and
230 ended with 11.

231 4.3. RESULTS

232 We included 59 bat species belonging to five families (Phyllostomidae, Noctilionidae,
233 Natalidae, Thyropteridae and Furipteridae), 10 subfamilies and 35 genera. The Bahia Coastal
234 ecoregion had the highest number of species (53 species, 89.83%), followed by Serra do Mar (45
235 species, 76.27%), Bahia Interior (38 species, 64.40%), Alto Paraná (37 species, 62.71%), Pernambuco
236 Coastal (32 species, 54.23%), Coast Restinga (19 species, 32.20%), Pernambuco Interior (18 species,
237 30,50%), Araucaria moist forests (17 species, 28.81%) and Campos Rupestres (15 species, 25.42%)
238 (See table S1.1 in pag. 106).

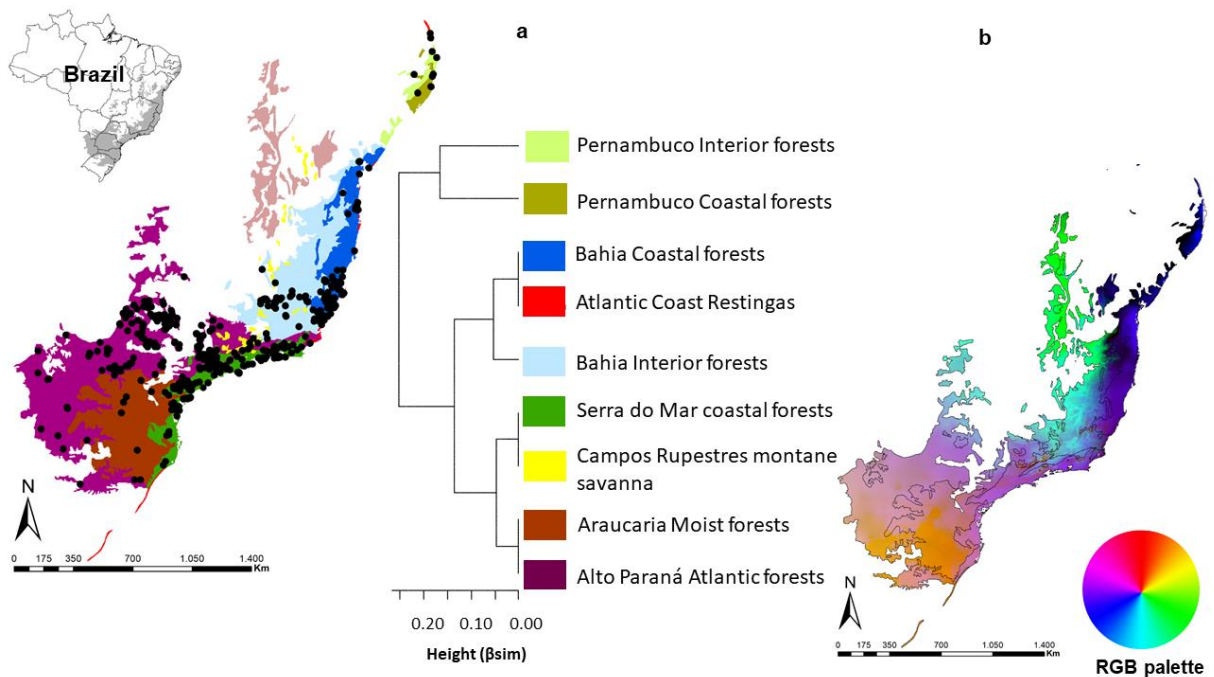
239 Our results indicate that geographic distance between sites as well as the spatial variation in
240 climate and elevation are correlated with bat species turnover in the Atlantic Forest (Fig. 1 and 2; and
241 figures in appendix S2 in pag. 111). The turnover component increases with the latitudinal distance
242 between ecoregions (see pairwise indexes matrix between ecoregions in Table 2 and Fig. 1b). The
243 mean dissimilarity (β_{sOr}) between ecoregions was 0.65, with a mean turnover (β_{sim}) of 0.36 and
244 nestedness (β_{sne}) of 0.29. The cluster and GDM analysis indicate three large groups of higher
245 similarity in species composition for the Atlantic Forest: southern, central and northern regions (Fig.
246 1a, b). Dissimilarity in species composition among these regions presented similar values for the
247 nestedness and turnover components ($\beta_{sOr} = 0.28$; $\beta_{sim} = 0.13$ and $\beta_{sne} = 0.15$). The turnover

248 component was larger and more important in separating the southern region from the central and
 249 northern regions (Table 3).

250 When beta diversity between ecoregions was analysed separately for each region (southern,
 251 central and northern), they presented higher values of nestedness, with the central and southern region
 252 having the lowest values of turnover (the turnover represented 13% of total beta diversity in the center
 253 region and 15% in the southern region) (see Table S2.3 in pag. 121, table S3.3 in pag. 123 and table
 254 S4.3 in pag. 125). The nestedness component was highest for ecoregions Araucaria and Campos
 255 Rupestres, which nested with Alto Paraná, Serra do Mar and Bahia Coastal (Table 2; and Table S2.3 in
 256 in pag. 121. For cluster see Fig S11.3 in pag. 123). Species dissimilarity between ecoregions that are
 257 located on the coast with ecoregions located in the interior by region (longitudinal gradient in the
 258 southern, central and northern regions of the Atlantic Forest), also presented higher nestedness.
 259 However, neighbouring ecoregions in the north, Pernambuco Coastal Forest and Pernambuco Interior
 260 forests, presented the highest turnover in species composition (42.5% of total beta diversity), similar to
 261 the values found among the northern, central and southern regions (Fig. 1a; Table 2; Table 3; Fig. S8.3
 262 in pag. 120; and Table S4.3 in pag. 125).

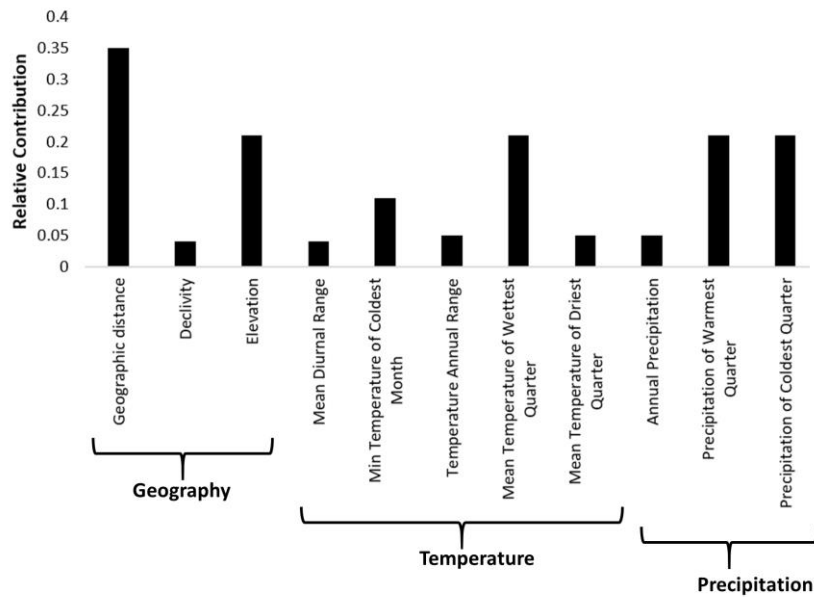
263 The GDM explained 40% of the observed species turnover for the Atlantic Forest. The
 264 variables with greatest predictive power were: geographic distance, elevation, precipitation of warmest
 265 quarter, precipitation of coldest quarter, mean temperature of wettest quarter and minimum
 266 temperature of coldest month (Fig. 1b, Fig. 2 and see Appendix S2 in pag. 114). The non-explained
 267 variation in GDM could be the result of biotic interactions and other factors (e.g., historical processes),
 268 which were not considered.

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271 **Figure 1. a)** Atlantic Forest ecoregions with bat study sites (coordinates) (black dots). The cluster of
 272 the turnover component (β_{sim}) provides values (height) from zero (same species found in both
 273 ecoregions) to one (no shared species between ecoregions). **b)** Map of bat species turnover based on
 274 geo-climatic predictors, the result of a Generalized Dissimilarity Model (GDM). The map is
 275 represented by RGB palette, colours represent the species composition gradient (β_{sim}) based on
 276 environmental variables. Similar colors represent communities with similar species composition.
 277 GDM explained 40% of the model, so the color differences in the RGB map are a reflection of this
 278 explained part. There may be differences in the RGB map that the other 60% of the model did not
 279 explain.



280 **Figure 2.** Eleven geo-climatic predictors that were significant predictors of bat species composition
 281 (β_{sim}) in the Atlantic Forest according to a Generalized Dissimilarity Model (GDM): geographical
 282 distance (decimal degrees); declivity (degrees); elevation (m); mean diurnal range ($^{\circ}\text{C}$); minimum
 283 temperature of coldest month ($^{\circ}\text{C}$); annual temperature range ($^{\circ}\text{C}$); mean temperature of wettest
 284 quarter ($^{\circ}\text{C}$); mean temperature of driest quarter ($^{\circ}\text{C}$); annual precipitation (mm); precipitation of warmest
 285 quarter (mm) and precipitation of coldest quarter (mm). The relative contribution of each predictor
 286 indicates its relative importance in contributing to biological turnover while holding all other variables
 287 constant. Hence, the relative contributions will not sum to 100%.

288 4.4. DISCUSSION

289 Our results indicate that bat species turnover is the main component in the dissimilarity among
 290 bat communities in the Atlantic Forest, under the influence of geographical distance and latitude. The
 291 larger the difference in latitude between occurrence sites, the larger was the turnover of species. This
 292 biogeographical pattern is common for many regions of the globe and for several taxa (Soininen et al.,
 293 2018). What is unique about our study, however, is the importance of species loss (nestedness
 294 component) in determining beta diversity dissimilarity when communities are analysed at smaller
 295 scales – among regions (north, centre and south), and among ecoregions within each region. The
 296 southern region was the most dissimilar, being highly nested with the centre and north regions, which
 297 explained why occurrence sites located furthest from the south had the greatest turnover of species.
 298 This distinctive pattern is clearly observed in the orange patch of Figure 1b, which covers the
 299 ecoregions Araucaria forest and the southern part of the Alto Paraná ecoregion, separating the warmer
 300 (centre and north) from the coldest (south) regions of the Atlantic Forest biome. We also found that
 301 nestedness was the main pattern explaining dissimilarity longitudinally, independent of latitude. This
 302 gradient in species loss from the coastal ecoregions to the interior could be explained by differences in
 303 temperature, elevation and precipitation, which are known environmental filters for bat species
 304 (Stevens, 2013). The GDM analysis was able to explain 40% of the turnover based on environmental
 305 and geographic variables, but we emphasize that part of this 40% may be mere chance. The other 60%
 306 that is not explained by the model may be due to historical, evolutionary and ecological filters (eg.
 307 biotic interactions), which were not tested.

308 In the southern region of the Atlantic Forest, pronounced changes in species composition
 309 occurred above 1000 m a.s.l., showing a gradient of species richness loss in ecoregions at high
 310 elevation: Araucaria and Campos Rupestres. The low temperatures associated with higher elevations
 311 separate the Araucaria and Campos Rupestres ecoregion from Alto Paraná and Serra do Mar, and

could be a limiting factor for the occurrence of several bat species. Flying is a costly activity, and because bats have a high ratio of body surface area to mass (because they are winged), they are more likely to lose heat to the environment and need a higher metabolic rate to maintain their body temperatures (Stones & Wiebers, 1965). There is also an indirect influence, since temperature and precipitation influence plant phenology, limiting the availability of food resources (Newton, 2003; Marques, Roper, & Savalaggio, 2004; Stevens, 2013). Thus, lower resource availability coupled with higher energetic demands for bats living in colder climates will be a strong limiting filter for several bat species, and only species adapted to colder climates will occur in Araucaria forests. Our study did find that all of the bat species occurring in the Araucaria ecoregion (see table S1.1 in pag. 106) have a broad distribution and are commonly sampled in almost all of the Brazilian biomes (Reis, Peracchi, Batista, Lima, & Pereira, 2017). Species with large range sizes face greater heterogeneity in environmental conditions and are capable of maintaining viable populations under different climatic and environmental conditions (Brown, 1984; Devictor, Romain, & Frédéric, 2008; Estrada et al., 2015). Also, most of the bat species from the Araucaria ecoregions (e.g. species of the genera *Artibeus*, *Sturnira*, *Carollia*, *Anoura*, *Platyrrhinus*) are known to occur in primary and secondary forests, as well as anthropogenic habitats (Reis et al., 2017), suggesting that bat species occurring in the Araucaria forests have a more plastic response to climatic and environmental conditions. The fact that of the 48 bat species that occur in the southern Atlantic Forest, only two (*Macrophyllum macrophyllum* and *Noctilio albiventris*) were not found in the central and northern areas, supports this interpretation. Conversely, 11 bat species were exclusively associated with the two warmer (central and northern) regions (*Glyphonycteris daviesi*, *Dryadonycteris capixaba*, *Phyllostomus elongatus*, *Lichonycteris degener*, *Trinycteris nicefori*, *Rhinophylla pumilio*, *Dermanura gnoma*, *Platyrrhinus incarum*, *Uroderma bilobatum*, *Micronycteris schmidtorum* and *Lophostoma brasiliense*). These species are also present in other warmer regions of Brazil, such as the Amazon (including savannas), Cerrado and Caatinga, which suggests that temperature is an important limiting factor in their distribution (Reis et al., 2017).

The lowest hierarchical level – ecoregions within each region – indicated that nestedness was the main factor influencing differences in beta diversity. In this case, the interior ecoregions were nested within the coastal ecoregions. This longitudinal pattern seems to follow an elevational gradient, particularly in the south and north of the Atlantic Forest (see elevation spline curve in Figure S1.2 in pag. 114). In the south, there is a system of mountains and escarpments that runs for 1,500 km along the Atlantic coast. The ecoregion Serra do Mar occurs on the eastern side of the mountain range along a 100 km wide strip with lowland forests on the coastal plains and upper montane forests in mountainous areas (~ 1,500 m a.s.l.). This creates a strong vegetation gradient and this ecoregion is known for its remarkable richness and endemism in plants (Werneck, Sobral, Rocha, Landau, & Stehmann, 2011), amphibians (Campos & Lourenço-de-Moraes, 2017), birds (Silva, Souza, & Castelletti, 2004) and mammals (Costa, Leite, Fonseca, & Fonseca, 2000). The interior forests are located on the western side of this mountainous region, and all have lower bat species richness and comprised subsets (i.e., were highly nested) of the bat fauna from the Serra do Mar ecoregion. For example, of the 37 bat species that occurred in Alto Parana, 34 of them also occurred in Serra do Mar ecoregion.

The same longitudinal pattern was recovered between Coast Restinga (an ecoregion that extends from the far north to the southern end of the Atlantic Forest coast) and the interior ecoregions. This longitudinal pattern seems to follow the precipitation regime, which in the central and northern region of the biome decreases dramatically from the coast to the interior (see spline curves Fig S2.2 in pag. 115). Coastal regions hosted several bat species that did not occur in the interior (such as *Lampronnycteris brachyotis*, *Lonchophylla mordax*, *Carollia brevicauda*, *Uroderma magnirostrum*, *Vampyrodes caracioli*, *P. incarum*, *D. gnoma*, *T. nicefori*, *G. daviesi*, *L. degener*, *P. elongatus* and *M. schimidtorum*), which resulted in greater dissimilarity between bat communities. Moreover, ecoregions in the interior presented a high number of shared bat species with coastal ecoregions and only a low number of species replacement (or turnover), which indicates that many species are lost along the longitudinal gradient of the Atlantic Forest. However, the Pernambuco ecoregions, which are located in the lower latitudes of the biome, have some climatic peculiarities that are not found in other regions of the Atlantic Forest: 1. there is no marked difference in elevation between the Pernambuco

366 Coastal and Pernambuco Interior; 2. despite there being a drop in temperature during the wettest
 367 season of the year in the Pernambuco Interior ecoregion, temperatures are not as low as in other
 368 regions of the Atlantic forest (Fig S3.2 in pag. 116); and 3. there is a pattern of higher precipitation
 369 during colder periods in the north that does not occur in other regions of the Atlantic forest . These
 370 environmental conditions may explain why Pernambuco Coastal and Pernambuco Interior had the
 371 highest turnover of bat species (40% dissimilarity), relative to neighbouring ecoregions of the Atlantic
 372 Forest.

373 Determining species distribution is a challenging task. We acknowledge that there are
 374 sampling gaps for the chiropteran from the Atlantic forest, as can be seen in Fig. 1a. However,
 375 sampling points are not concentrated in small patches for most of the ecoregions. In the centre of the
 376 biome, the southern portions of the Bahia Coastal and Bahia Interior ecoregions are better sampled,
 377 but, the central region is more similar in composition to the northern than the southern region. The two
 378 ecoregions that make up the north, Pernambuco Interior and Pernambuco Coastal, are small and have
 379 few study sites, but over 50% of the species included in our study were present in Pernambuco Coastal
 380 region, which suggests that sampling gaps are a minor issue. The GDM analysis explained less than
 381 50% of the dissimilarity among communities, which indicates the importance of historical factors in
 382 determining species distributions. Despite our results uphold previous studies that indicated that the
 383 central and northern regions served as refugia during past climatic changes (i.e., they have higher
 384 species richness) in the Atlantic Forest (Carnaval & Moritz, 2008; Carnaval et al., 2014), this
 385 hypothesis has yet to be tested for bats. Our findings showed that the composition of bat communities
 386 varied tremendously throughout the Atlantic Forest. This corroborates the importance of partitioning
 387 beta diversity (turnover and nestedness) and of analysing it at different scales to understand how
 388 environmental variation in geographic space influences species distribution patterns. The patterns
 389 identified here have important implications for understanding how bats respond to changes in the
 390 environment, which in turn helps predict how environmental change can interfere with species
 391 distributions. For example, we found that climatic factors have a strong influence on the distribution
 392 pattern of bats, and changes in precipitation and temperature may drive the configuration of beta
 393 diversity components. Species present in the coastal ecoregions of the Atlantic Forest, which are
 394 warmer, wetter and where forest refugia have been predicted, may respond negatively to climate
 395 change if the climate of the coastal ecoregions were to become more similar to the more unstable
 396 climatic zones of the interior ecoregions of the Atlantic Forest. Although this potentially negative
 397 impact needs to be tested, with the use of species distribution models and different climate scenarios,
 398 our results indicate the importance of environmental factors as predictors of bat species composition,
 399 and that they should be taken into account in biogeographic, ecological and conservation studies.

400 TABLES

Table 1. Summary of our hypotheses and their main predictions. The table shows the predictions and respective key references supporting the hypothesis. The hypotheses 1 and 2 were accepted, and hypotheses 3 and 4 were partially accepted.

Hypothesis	Predictions	Key references
H ₁ . Species turnover will be associated with geographical distance and will be the main component of beta diversity in the latitudinal gradient; for example, the more distant the latitudinal sites, the greater the	Because of flight ability bats should manage to track their environmental preferences. Therefore, variation in bat species composition should be more influenced by geographical distance and less so by local factors. Composition will be affected by environmental	Soininen et al., 2018; Varzinczak et al., 2019 and López-Aguirre et al., 2019.

turnover.	changes at a large scale, for example, the latitudinal gradient. Thus, the greater the distance between ecoregions, the bigger the turnover will be.	
H ₂ . The main process determining compositional dissimilarity at a longitudinal gradient (from the coast to the interior) will be nestedness.		
H ₃ . The ecoregions located in the south of the Atlantic Forest, which are at higher latitudes and have colder climates, will have the lowest richness and turnover rate of the Atlantic Forest.	Climate patterns are associated with latitude, which is also a proxy for species richness and turnover gradients. Also, bats have high physiological demand (energy and metabolism) because of their flight capability. Together with their high ratio of body surface area to mass, several bat species will have difficulties coping with cold climates.	Stevens & Willig, 2002; Willig et al., 2003; Dobrovolski et al., 2012 and Soininen et al., 2018.
H ₄ . The ecoregions located at lower latitudes, in the centre and north, will present the highest richness and species turnover index.		

401

402 **Table 2.** Pairwise dissimilarity index matrix of the turnover (β_{sim}) and nestedness (β_{sne}) component
403 of bats species between the Atlantic Forest ecoregions (according to Olson et al. 2001) of Brazil. This
404 matrix shows the results of the Beta diversity analysis (Betapart/R-project) and the index was
405 calculated using the presence and absence of bat species for each ecoregion. Total turnover and
406 nestedness are also presented.

	Alto Paraná	Araucaria	Coast Restinga	Bahia Coastal	Bahia Interior	Campos Rupestres	Pernambuco Coastal	Pernambuco Interior
	Turnover (β_{sim}) - 0.361194							
Araucaria	0.00000000							
Coast Restinga	0.10526316	0.35294118						
Bahia Coastal	0.11111111	0.00000000	0.00000000					
Bahia Interior	0.25000000	0.05882353	0.05263158	0.05263158				

Campos Rupestres	0.00000000	0.13333333	0.40000000	0.00000000	0.06666666			
Pernambuco Coastal	0.25000000	0.23529412	0.26315789	0.06250000	0.25000000	0.33333333		
Pernambuco Interior	0.16666667	0.52941176	0.33333333	0.00000000	0.16666666	0.66666666	0.16666666	
Serra do Mar	0.05555555	0.00000000	0.05263158	0.11111111	0.13157895	0.00000000	0.21875000	0.05555555
Nestedness (β_{sne}) - 0.2980416								
Araucaria	0.35849057							
Coast Restinga	0.27655502	0.03594771						
Bahia Coastal	0.16978777	0.51428571	0.47222222					
Bahia Interior	0.02027027	0.35935829	0.31578947	0.15615963				
Campos Rupestres	0.41176471	0.05416667	0.07058824	0.55882353	0.40503145			
Pernambuco Coastal	0.04411765	0.23409364	0.18782250	0.23161765	0.06428571	0.24113475		
Pernambuco Interior	0.27777778	0.01344538	0.01801802	0.49295775	0.29761905	0.03030303	0.23333333	
Serra do Mar	0.10493827	0.45161290	0.38486842	0.07256236	0.07324033	0.50000000	0.13189935	0.40476190

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Table 3. Pairwise index matrix of the turnover (β_{sim}) and nestedness (β_{sne}) component of bat species among three regions (south, central and north) of the Atlantic Forest of Brazil. South region includes

425 the Serra do Mar, Alto Parana, Araucarias and Campos Rupestres ecoregions; Central region includes
 426 Bahia Coastal, Bahia Interior and Coast Restinga ecoregions; and the North region includes
 427 Pernambuco Coastal and Pernambuco Interior ecoregions. This matrix shows the result of the Beta
 428 diversity analysis (Betapart/R-project) and the index was calculated using the presence or absence of
 429 bat species for each of the three regions. Total turnover, nestedness and total beta diversity are also
 430 presented.

	South	Central
	Tunover (β_{sim}) - 0.1318681	
Central	0.08333333	
North	0.17142857	0.05714286
	Nestedness (β_{sne}) - 0.1564202	
Central	0.06229773	
North	0.12977625	0.20952381
	Beta diversity ($\beta_{s\emptyset r}$) - 0.2882883	
Central	0.1456311	
North	0.3012048	0.2666667

431 DATA AVAILABILITY STATEMENT

432 The authors are willing to share the data, including distribution points and data analysis scripts, in a
 433 public repository (Dryad) if the article is accepted for publication. We choose not to share prior to
 434 acceptance for the sake of ensuring that our results are not reproduced by third parties. We are willing
 435 to supply any further information and we appreciate your understanding.

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559 biotic similarity? *Ecography*, 40, 478–486.
- 560 **BIOSKETCH**
- 561 Carolina Blefari Batista is broadly interested in the biogeography of bats. This work represents a
562 component of her PhD work at Universidade Estadual de Londrina (UEL), Londrina/PR, Brazil, on the
563 ecology, biogeography and conservation of bats in the Atlantic Forest.
- 564 Author contributions: CBB, IPL and MRL conceived the ideas; CBB conducted the data collection
565 and analysed the data (Beta diversity and Generalized Dissimilarity Model); CBB led the writing with
566 assistance from IPL and MRL.
- 567

1 5. CAPÍTULO 3

2 **Downscaling the Atlantic Forest biodiversity hotspot: using the distribution of bats to** 3 **find smaller hotspots of conservation priority**

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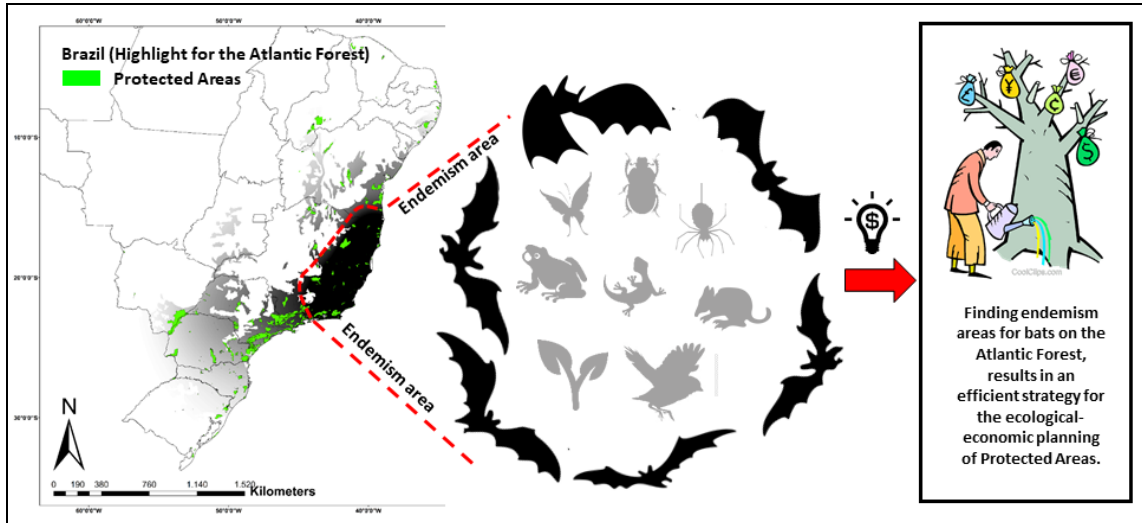
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11 **Abstract**

12 The Atlantic Forest (AF) is one of most important and threatened biodiversity hotspots
13 in the world and despite its wide geographical extension it has only 9.1% of its area under
14 legal protection. This fact linked to the low investment of resources in conservation makes it
15 necessary to find smaller hotspots of conservation priority in the biome. Finding an endemic
16 area for bats can be an effective strategy in the indication of priority areas. Because bats have
17 a high taxonomic and functional diversity, implying complex networks of interactions, this
18 strategy can represent sites with high diversity of other taxa as well. We used a Geographic
19 Interpolation of Endemism (GIE) analysis and indicated a congruence area for 58 of the 59
20 bat species included in this study, located in the center of the AF. Five species occur strictly in
21 this area. We also superimposed the map of legally protected areas (PAs) with our map of
22 endemism. We found that only 14 species had more than 50% of their records in PAs across
23 the biome. Considering only PAs within the endemism area, five species have no occurrence
24 in PAs, including a vulnerable species. Our area of endemism corroborates hotspots for other
25 taxa. Therefore, we present an efficient strategy in the planning of PAs in the AF, implying
26 that efforts for the conservation of biodiversity will be more efficient an increase in the cost-
27 benefit relationship in the establishment of PAs.

28 **Keywords:** Endemism area, Chiroptera, GIE, protected areas, efficient strategy.

29 **Graphical abstract**



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32 5.1. Introduction

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Human activities have triggered biodiversity loss at higher rates than those seen for natural extinctions. In fact, mankind is facing very fast extinction rates, mainly caused by climate change and habitat loss (Dirzo et al., 2014; Vellend et al., 2017). Meanwhile, there are still several deficits in global strategy for biodiversity conservation, such as low budgets and incorrect application of scarce resources (Johnson et al., 2017). In the rise of a future economic crisis due to the COVID-19 pandemic, experts are concerned that investment on conservation will be reduced, drawing attention to the importance of conservation biologists communicating the benefits of biodiversity conservation to society (Corlett et al., 2020). Thus, one of the greatest challenges for conservationists is to pinpoint where limited resources should be concentrated and applied, prioritizing certain areas in detriment of others (Di Minin et al., 2017).

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Several strategies have been employed to indicate priority conservation areas whose criteria vary according to the objective. Focus may be on conservation of rare species, endemic species, endangered species, general biodiversity maintenance or habitat conservation to keep ecosystem services and factors (Asaad et al., 2016). Since resources are scarce, the efficiency of priority areas regarding costs depends on a selection of strategies that comprises a multiple criteria dimension to properly conserve biodiversity (including endemic, rare and vulnerable species) and the ecosystem services for the longest period possible. Thirty-six critical areas of biodiversity concentration (hotspots) have been pinpointed (Myers

52 et al., 2000; Williams et al., 2011; Noss et al., 2015; CEPF, 2016). However, these
53 biodiversity hotspots were determined for a larger scale, despite conservation efforts and
54 strategies occurring on a much smaller scale (Jenkins; Pimm, 2006). In order to facilitate the
55 focus of conservation efforts, it is necessary to find and pinpoint high diversity sites at a
56 smaller scale, or rather “hotspots within hotspots” (Cañadas et al., 2014).

57 Regardless of the strategy employed, pinpointing priority areas and conserving
58 biological diversity requires knowing and understanding how species are distributed and
59 which factors determine their spatial distribution (Sigrist; Carvalho, 2008; Cañadas et al.,
60 2014; Batista et al., 2020). Despite recent progress in computer and analytic methods for a
61 better understanding of the processes involved in the distribution of species, this is still a
62 challenging issue in ecology and biogeography. First, species are not randomly distributed
63 and there are specific areas with great biodiversity concentration (Sigrist; Carvalho, 2008).
64 Second, most studies determine areas of greater biodiversity or endemism by employing
65 online-published databases or distribution maps of species (IUCN, for instance) (Sigrist;
66 Carvalho, 2008). This may add mistakes when delimiting priority areas. Species do not
67 respond in the same way to the limiting factors within the different habitats. Actually, it is
68 quite challenging to find congruent areas of species distribution (Oliveira et al., 2015).
69 Databases and maps on species distribution can have several gaps (e.g. low sampling effort
70 and lack of georeferenced information of the occurrence site), which leads to false absences
71 (Hortal, 2008; Stokland et al., 2011). Databases may also be biased, since they are constructed
72 over an arbitrary design, making it difficult to be used for biogeographic conservation studies
73 (Hortal et al., 2007). Further, these gaps may also prove to be the result of a greater
74 concentration of sampling efforts closer to human population centers, as has been reported for
75 Angiosperms (Nelson et al. 1990; Werneck et al. 2011), or by the uneven representativeness
76 of *taxa* in scientific collections (Ponder et al. 2001).

77 The Geographical Interpolation of Endemism (GIE) is one of the ways to delimitate
78 priority areas by determining sites with the greatest biodiversity and number of endemic
79 species, minimizing effects of species distribution gaps. This analysis was proved and tested
80 (Oliveira et al., 2015) and allows one to overlap areas in species distribution bands that are
81 sustained by at least one endemic species. Since GIE does not use grid cells in a distribution
82 map, it allows the use of occurrence data with spatial gaps (Oliveira et al., 2015). As
83 important as choosing the analytical method is to choose a group of organisms that could
84 represent diversity and distribution of other *taxa* (Oliveira et al., 2019). GIE analysis has been

85 employed for several *taxa*, such as plants (Castueira-Oliveira et al. 2020), invertebrates and
86 vertebrates (Oliveira et al., 2015; Oliveira et al., 2017), and for different biogeographic
87 regions (Castueira-Oliveira et al., 2020).

88 Due to the characteristics of bats and the Atlantic Forest global hotspot, the
89 investigation on sites of endemism may be important to help choose priority areas for
90 conservation. Areas of endemism comprise sites with high species distribution congruence
91 and the presence of restrict species (Harold; Mooi, 1994). Bats are the second most diverse
92 group in mammals, after rodents (Simmons, 2005). Such diversity is not just taxonomic, since
93 the group includes different feeding habits, including items such as fruits, nectar, insects and
94 other invertebrates, small vertebrates and even fish and blood (Kalka et al., 2008). They also
95 use several types of shelters such as leaves, trunks, trunk hollows, rocks, karstic environments
96 and even houses (Kalka et al., 2008). The above adaptations and the ability to fly makes bats a
97 good model for biogeographic studies focused on conservation (Aguiar et al., 2016; Herkt et
98 al., 2016), since its distribution may reflect other *taxa* distributions. The Atlantic Forest is one
99 of most important biodiversity hotspots in the world (Myers et al., 2000), with areas
100 considered to be biodiversity refuges due to its greater climatic stability, allowing a wider
101 diversification of species (Carnaval et al., 2009). On the other hand, 72% of Brazilians live
102 within the Atlantic Forest area (Ministério do Meio Ambiente, 2017) and only 9.1% of the
103 area is under legal protection (Sobral-Souza et al., 2018), making the biome vulnerable to
104 anthropogenic activities. Further, a comprehensive database on bat distribution in the Atlantic
105 Forest (Atlantic Bats – Muylaert et al., 2017) was built and extensively reviewed by experts.
106 This database is available and allows the development of more complex research with bats in
107 this biome.

108 Since bats may be useful to identify smaller priority areas for conservation within a
109 wider hotspot and that they could represent other *taxa*, increasing the cost-benefit ratio, this
110 study looked at two main issues. First, where are the centers of greatest concentration and
111 restriction of diversity of bats located within the limits of the global Atlantic Forest hotspot?
112 and second, how much of these areas are already protected by current legislation? To answer
113 those questions, we used the Bat distribution data for the Atlantic Forest and GIE analytic
114 tools to determine the main congruence areas for the occurrence of bat species. Congruence
115 regions were compared with legally protected areas at federal and state levels. This allowed us
116 to quantify the extent of these areas of congruence that are under legal protection. The two
117 strategies enabled the identification of conservation priority areas that are not yet protected by

118 current legislation.

119 **5.2. Methods**

120 5.2.1. Database

121 Two databases were constructed: a database for bat richness and another for
122 abundance indexes controlled by sampling effort. Both were required, since not all
123 georeferenced sites used for richness data provide data on species abundance and sampling
124 effort.

125 *5.2.1.1. Bat richness in the Atlantic Forest*

126 The *Atlantic Bats* database (Muylaert et al., 2017) was consulted. This database was
127 prepared by experts who also reviewed species taxonomy and sampling sites coordinates of
128 bat species occurrence records for the entire Atlantic Forest. From Muylaert et al. (2017)
129 database, we retrieved a species list that allowed searching another repository (speciesLink -
130 <http://splink.cria.org.br/>), because the database “Atlantic Bats” only takes into account records
131 that include abundance data. The dataset retrieved from speciesLink was analyzed according
132 to data reliability: i) correspondence of coordinates and sites (Google Earth maps were used to
133 verify whether coordinates referred to the indicated sites); ii) correct taxonomy (species with
134 “sp”, “ssp”, “cf” and “aff” were excluded); and iii) only records with specimens deposited in a
135 museum as vouchers were taken into account. Furthermore, we identified the absence of an
136 important record of *Natalus macrourus*, which was missing in Muylaert et al. (2017) and in
137 speciesLink. Thus, our database also includes the single record of *N. macrourus* (by Trajano,
138 1984) at the Parque Estadual Turístico do Alto Ribeira (PETAR). We noticed that our
139 database had few occurrence points for the northern region of the Atlantic Forest, so we also
140 included bat occurrence records from the Atlantic Forest of northeastern Brazil (Souza et al.,
141 2004; Campos et al., 2018; Nunes et al., 2018; Vargas-Mena et al., 2018). Records of bats
142 species occurrence from families Molossidae, Vespertilionidae and Emballonuridae were not
143 included in this database due to the difficulty in capturing them with the use of mist nets
144 (Nogueira et al., 1999; Nogueira et al., 2008). Mist nets are the predominant method used to
145 sample bat species in researched databases. To remove duplicates or very close coordinates of
146 occurrence records, we applied a latitude and longitude filter with three and two decimal
147 places, respectively. Then, we joined the columns of latitude and longitude plus species. Both
148 the filter with two and three decimal places resulted in the exclusion of 470 duplicate

149 occurrence records. Consequently, 2,829 bat occurrence data were retrieved, distributed in
150 535 sites (coordinates) for 60 species in the Atlantic Forest.

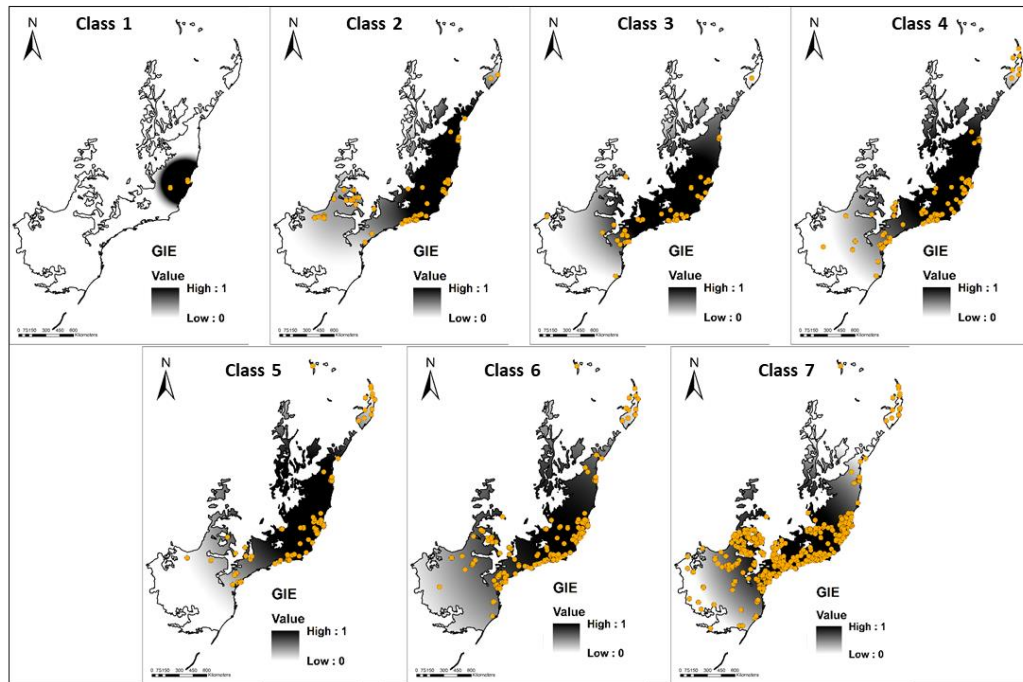
151 5.2.1.2. *Abundance controlled by sampling effort (A/SE)*

152 The *Atlantic Bats* (Muylaert et al., 2017) provides species' abundance data. However,
153 since sampling effort varied for each sampling site (coordinate), we filtered the database to
154 select only the occurrence records provided with species abundance and sampling effort. We
155 used this data to calculate the abundance index controlled by sampling effort
156 (Abundance/Sampling effort or "A/SE"). Occurrence sites (coordinates) with no sampling
157 effort data were recorded as 0 for this index. We elaborated "A/SE" indexes for 58 species
158 (see distribution map for each species in pag. 127), distributed in 1,397 records for 133
159 occurrence sites, and retrieved from the *Atlantic Bats* database only.

160 5.2.2. Congruence area of bat distribution in the Atlantic Forest

161 We used GIE method as suggested by Oliveira et al. (2015). This method consists in
162 five steps; (1) Determination of species distribution centroid, based on occurrence sites, using
163 the bat richness database for the Atlantic Forest, compiled from published databases (see "Bat
164 richness for the Atlantic Forest", section 5.2.1.1). Centroid is determined by mean latitude and
165 longitude values of the species' occurrence sites. (2) Determination of species' distribution
166 range, or rather, the distance between the centroid and the most distant occurrence site.
167 Distance is determined by Pythagoras theorem where the hypotenuse is the distance between
168 the centroid and the most distant site. (3) Species are categorized into different groups by
169 distance, with six distance classes of centroid influence: Class 1, species range of 0 to 300
170 km; Class 2, from 301 to 600 km; Class 3, from 601 to 900 km; Class 4, from 901 to 1,200
171 km; Class 5, from 1,201 to 1,500 km; Class 6, from 1,501 to 1,800 km; Class 7, from 1,801 to
172 2,171 km (Fig. 1; Table 1A in pag. 148). The above classification is important since several
173 species may have the same or very close centroids, albeit with very distinct distributions.
174 Categorization within different distance radii favors the comparison of congruence areas for
175 species with similar distribution ranges. (4) Determination of the area of endemism for each
176 class of distribution range using kernel algorithm. (5) Preparation of a consensus map
177 involving the maps of the six distance classes. Microsoft Excel (version 2016) and GIE
178 package inserted in ArcGIS 10.5 of ESRI (ArcGIS trial) were employed for this analysis.

179



180
 181 **Fig. 1.** Classification of 60 Atlantic Forest bat species from the distribution radius of the Geographic
 182 Interpolation of Endemism (GIE). The endemism area index varies from a scale of 0 to 1 (endemic
 183 area). Class 1, species range of 0 to 300 km; Class 2, from 301 to 600 km; Class 3, from 601 to 900
 184 km; Class 4, from 901 to 1,200 km; Class 5, from 1,201 to 1,500 km; Class 6, from 1,501 to 1,800 km;
 185 Class 7, from 1,801 to 2,171 km. Orange dots represent georeferenced occurrence sites of each
 186 species' class.

187 5.2.3. Analysis of sampling effort

188 Database for bat richness and the shape file of the Atlantic Forest limits were
 189 employed to prepare hotspot maps indicating the greatest sampling efforts for bat capture.
 190 Analysis was undertaken using BioDinamica package (Oliveira, 2018) in Dinamica Ego
 191 software. The function employs kernel algorithm to interpolate georeferenced occurrence sites
 192 and generate a sampling effort map. Adjustments in the analysis includes Search radius, or
 193 rather, the sampling influence area. In other words, what is the influence area of the mist nets
 194 sampling for the capture of bats? As number, size and distribution of mist nets may vary
 195 according to sampled sites, three analyses were undertaken with different sampling
 196 influences: 10 km², 50 km² and 100 km².

197 5.2.4. Relationship between the distribution latitude band of Atlantic Forest species, richness 198 and abundance.

199 We prepared rarefaction curves by extrapolation of species based on *Hill* series using

200 iNEXT package (Hsieh, 2019) on an R environment (R Core Team, 2020). Rarefaction
 201 indicates observed bat richness as a function of sampling effort employed (occurrences sites –
 202 coordinates were used) and the *Hill* series-based extrapolation shows the number of species
 203 that may be introduced into the sample when the sampling effort is doubled. It is a modern
 204 technique which replaces previous ones that made inferences based on non-parametric metrics
 205 as jackknife, chao and bootstrap. Further, richness was calculated with the function
 206 “diversityresult” from BiodiversityR package (Kindt; Coe 2005). Richness
 207 calculated values were distributed on a scatterplot with latitudes on the X axis, with
 208 “ggplot” function from “ggplot2” package (Wickham 2009) within an R environment (R
 209 Core Team, 2020). The “A/SE” indexes were summed for each georeferenced site and plotted
 210 on a scatterplot in the same way as for richness.

211 5.2.5. Representativeness of state and federal Protected Areas for bat species in the Atlantic
 212 Forest

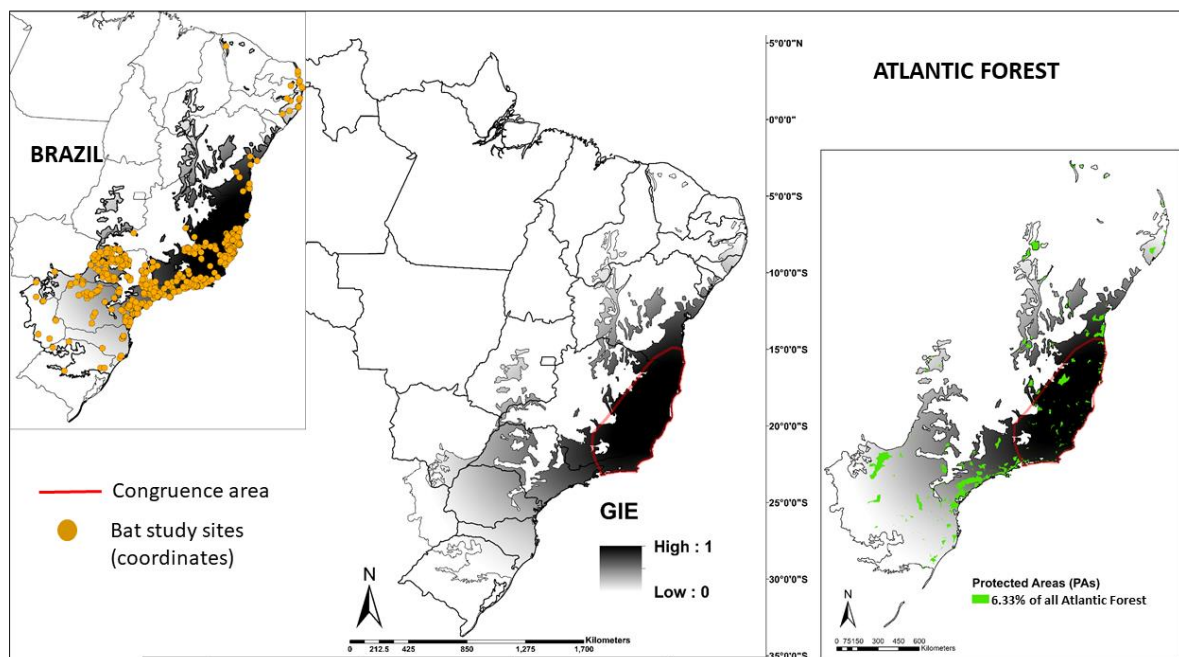
213 State and federal Protected Areas (PAs) were plotted (shape file available at
 214 <https://www.mma.gov.br/areas-protegidas/cadastro-nacional-de-ucs/mapas.html>) over the
 215 Atlantic Forest (shape file available at https://www.mma.gov.br/biomas/mata-atl%C3%A2ntica_emdesenvolvimento/mapas-da-mata-atl%C3%A2ntica) and the percentage
 216 of PAs was calculated within the congruence area indicated by GIE, using “calculate
 217 geometry” tool of the table attribute in ArcGis 10.5. We transformed Bat Richness sheet
 218 (occurrence by coordinates) into layers and selected only the occurrence sites within PAs. We
 219 then determined the number of species within PAs in the Atlantic forest and within the
 220 congruence area. Analysis were also performed in ArcGis 10.5 of ESRI (ArcGIS trial
 221 license).

223 5.3. Results

224 The GIE analysis detected a single congruence area (Fig. 2) for the distribution of 58
 225 out of 59 bat species in the Atlantic Forest, based on the presence of five restricted species
 226 (*Micronycteris hirsuta*, *Micronycteris schmidtorum*, *Dryadonycteris capixaba*,
 227 *Glyphonycteris daviesi* and *Dermanura gnomia*) (Table 1A in pag. 148). Only *Noctilio*
 228 *albiventris* did not occur within the congruence area. Rarefaction curve and extrapolation
 229 showed that, if sampling effort were doubled (1070 sites), the final list would contain a total
 230 of 59 species (not different from the observed richness), showing that our database for the

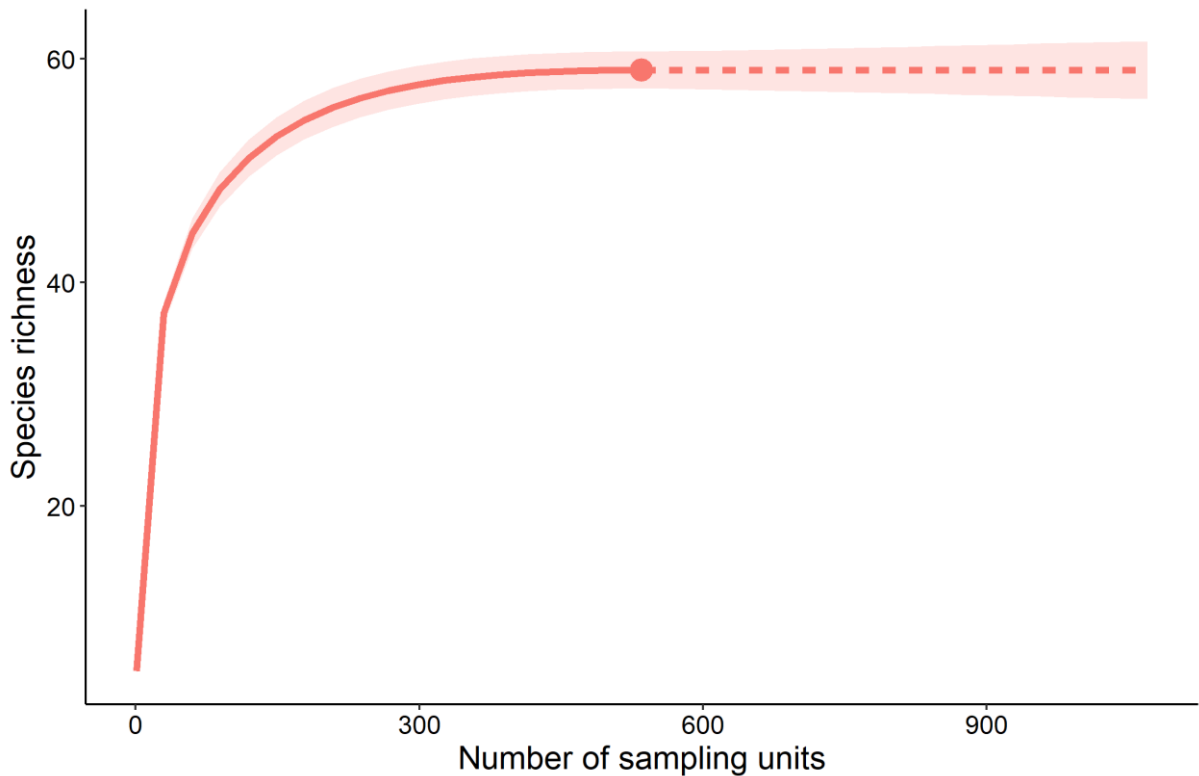
231 Atlantic Forest represents a reliable sample of bat species from Phyllostomidae,
 232 Noctilionidae, Furipteridae, Thyropteridae and Natalidae families, frequently captured by mist
 233 nets (Fig. 3). Analysis on sampling efforts revealed that the Serra do Mar region on the south
 234 coast of the Atlantic Forest, the northwest of the São Paulo state and the Biome's central
 235 region are sampling hotspots, or rather, that current results are not merely a product of the
 236 sampling effort (Fig. 4).

237 Highest species richness lies within latitudes -25 and -19, whilst the greatest "A/SE"
 238 indexes are between latitudes -22 and -19. Richness and abundance peaks corroborate with
 239 the congruence area detected by GIE (Fig. 5). Although 58 out of 59 bat species have
 240 occurrence records within PAs throughout the Atlantic Forest, only 14 species had more than
 241 50% of their records within these areas. If the congruence area alone is considered, five of the
 242 58 species fails to occur in PAs (*Macrophyllum macrophyllum*, *Phylloderma stenops*,
 243 *Vampyrodes carccioli*, *Thyroptera tricolor*, and *Lonchorhina aurita* which is categorized as
 244 "Vulnerable" by "Lista Vermelha da Fauna Ameaçada", ICMBio/MMA, 2018) and 34
 245 species shows less than 50% of occurrence records in PAs (Table 1A in pag. 148). The map
 246 overlapping PAs and Atlantic Forest (Fig. 2) indicates that about 6.33% of the forest cover of
 247 the Atlantic Forest is legally protected, and of the total congruence area, 3.9% is legally
 248 protected.

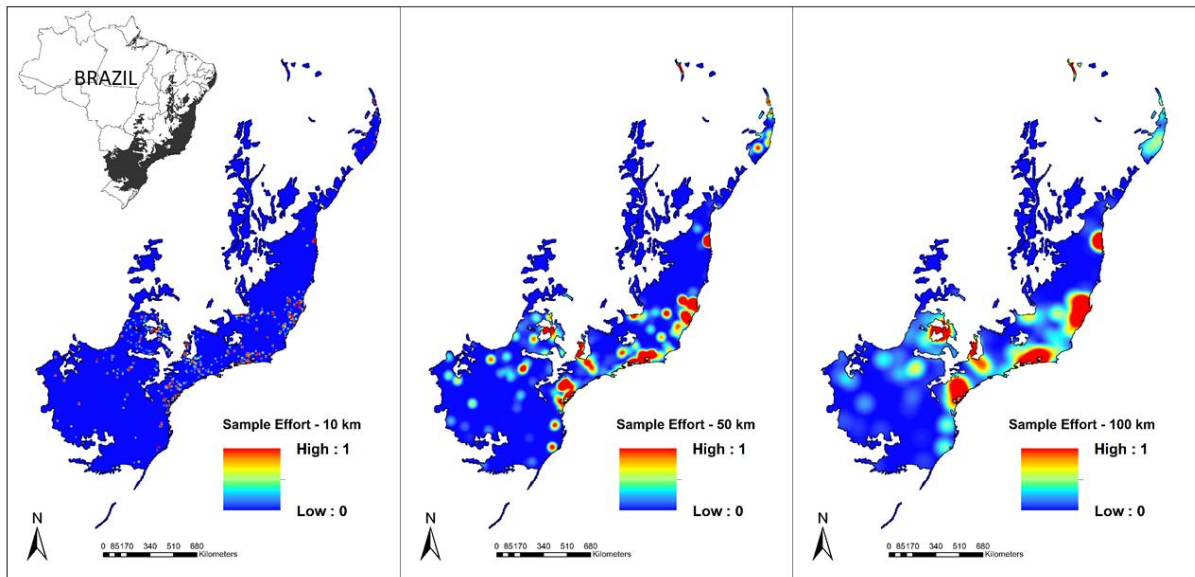


249
 250 **Fig. 2.** Consensus map of the endemism area resulting from the Geographic Interpolation of
 251 Endemism (GIE) for 59 bat species of the Atlantic Forest, of which 58 are present in the congruence

252 area .The GIE index ranges from 0 (area with low or no congruence of species distribution and without
 253 restricted species) to 1 (area with high congruence of species distribution and with restricted species);
 254 and overlapping map of Protected Areas (PAs). About 6.33% of the forest cover of the Atlantic Forest
 255 is legally protected, and the congruence area features 3.9% of the legally protected area.

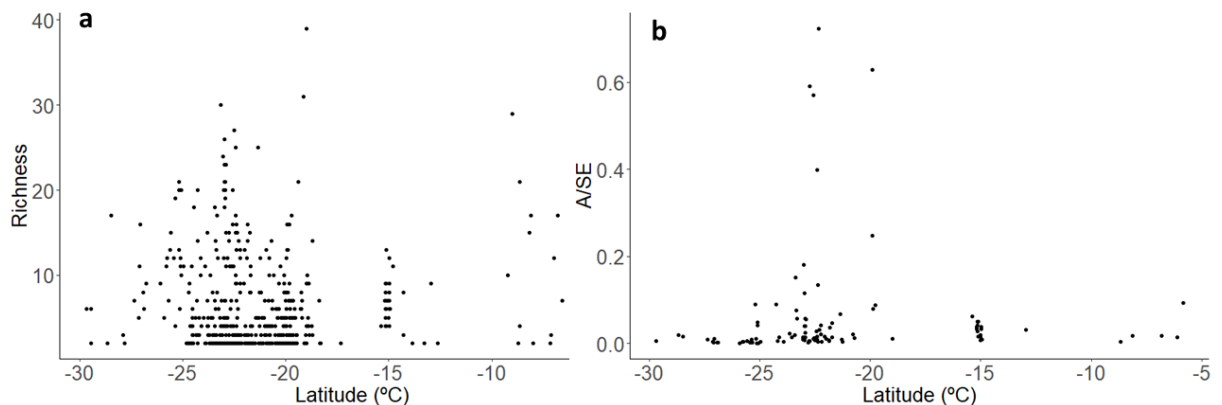


256 **Fig. 3.** Observed richness of 59 bat species (continuous line) as a function of the total number of
 257 sampling units (georeferenced locations) in the Atlantic Forest, with a 95% confidence interval for the
 258 Hill series. The dotted line represents the extrapolation of species richness in the study area if the
 259 sampling effort was doubled (the estimated final list would contain a total of 59 species - not different
 260 from the observed richness), showing that our database for the Atlantic Forest represents a reliable
 261 sample of bat species from Phyllostomidae, Noctilionidae, Furipteridae, Thyropteridae and Natalidae
 262 families, frequently captured by mist nets
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Fig. 4. Maps resulting from the sampling effort analysis for bats in the Atlantic Forest, with influence areas of 10 km², 50 km² and 100 km². The sampling effort index ranges from 0 (no effort employed on the spot) to 1 (locations with the highest employed sampling effort).



268

Fig. 5. Distribution of richness (a) and abundance (b; represented by the sum of the abundance index controlled by the sampling effort for each georeferenced location “A / SE sum”) of bats in the Atlantic Forest by latitude (georeferenced occurrence locations). The richness was calculated based on the total matrix of 59 species in 535 georeferenced sites. The abundance is represented by the sum of the abundance index controlled by the sampling effort for each georeferenced location “A / SE sum”, and was calculated from the matrix with data of abundance controlled by the sampling effort (“A / SE”), with a total of 58 species in 133 georeferenced sites.

276 5.4. Discussion

277

This research is the first to evaluate bat distribution for the Atlantic Forest and it has

278 demonstrated that the biome's central region is the only congruence area for the distribution
279 of 98.3% of species included in this study. In fact, it represents 60.20% of bat species
280 occurring in the Atlantic Forest and listed by Muylaert et al., (2017), and 32,41% of bat
281 species known in Brazil (based on the list of Nogueira et al., 2018). Although five bat species
282 with restricted range may be found in other biomes (Table 1A in pag. 148), in the Atlantic
283 Forest they are concentrated within this congruence area. Two species belong to the same
284 genus (*Micronycteris*) and *Dryadonycteris capixaba* (Nogueira, Lima, Peracchi & Simmons,
285 2012). *Dryadonycteris capixaba* is an endemic species to Brazil, it has been recently
286 described and its conservation status has not yet been accessed (Global assessment - IUCN,
287 2020; National assessment - ICMBio/MMA, 2018).

288 Despite the Atlantic Forest not being sufficiently sampled for bats, our analysis on
289 sampling effort (Fig. 4) and corroboration of congruence area by other studies prove that our
290 result is not a mere outcome of the sampling effort. The Atlantic Forest center, covered by
291 Bahia's coastal and interior ecoregions (Olson et al., 2001) is pinpointed as an endemic and
292 priority area for the conservation of different taxa, such as amphibians, reptiles, arthropods
293 (Sigrist; Carvalho, 2008; Carnaval et al., 2014; Campos; Lourenço-De-Moraes, 2017), birds
294 and mammals (Ceballos; Ehrlich, 2006), besides being acknowledged as a forest refuge
295 (Carnaval; Moritz, 2008; Costa et al., 2018). However, there are other areas throughout the
296 Atlantic Forest for these taxa, especially in the state of Pernambuco and in the Serra do Mar.
297 Models of bat distribution developed for the entire Brazilian area (Delgado-Jaramillo et al.,
298 2020) indicates that Pernambuco area has the greatest richness, whereas Serra do Mar region
299 has the greatest endemism. We emphasize that this area is the only one where distribution
300 bands overlaps for more than half of the bat species of the entire Atlantic Forest (Muylaert et
301 al., 2017). Either for species with their distribution centers coinciding with congruence areas
302 as for species whose congruence area represents one of its distribution limits which extend
303 towards the far south or north of the Atlantic Forest (see distribution maps of species in pag.
304 127).

305 This central trend for greater richness and species restriction in the middle latitudes
306 has already been demonstrated at a larger scale for New World bats, where the inter-tropical
307 zone proved to be one of the richest and most endemic regions (Arita et al., 2014). The
308 Atlantic Forest is a biome with a broad range in latitude (featuring a 30-degree interval) and it
309 is known that latitude predicts environmental (climatic) gradients (Willig et al., 2003;
310 Soininen et al., 2018). These gradients affect bat distribution patterns with the Bahia's eco-

311 regions (the biome's middle latitudes), especially its coastal ecoregion, ~~to be one of being~~ the
312 richest in all of the Atlantic Forest biome regarding bat species (families Phyllostomidae,
313 Noctilionidae, Furipteridae, Thyropteridae and Natalidae) (Batista et al., 2020). Our approach
314 demonstrates that this central region is not only the richest zone but also represents a
315 congruence area with the greatest diversity and number of restricted species. Although not all
316 species have the same behavior (see maps on species distribution in Supplementary Material
317 1), the general distribution of abundance follows the same richness pattern. The "A/SE" index
318 decreases from the center of the Atlantic Forest, mainly towards the biome's far southern
319 region (Fig. 4), which is highly nested to the central and the northern segments of the Atlantic
320 Forest, regarding bat richness (Batista et al., 2020). Consequently, bat species respond to
321 favorable environmental conditions which, for most of them, decrease from the center to the
322 biome's extremes, especially the far southern segment where low temperature climates are
323 more hostile and species ecological requirements are harder to attain (Batista et al., 2020).

324 Besides taxonomic diversity, the congruence area analyzed comprises species with
325 diversified life history traits including all feeding-habit guilds of bats (frugivore, nectarivore,
326 insectivore, carnivore, omnivore, piscivore and hematophagous species) (Table 1A in pag.
327 148). This suggests an area rich in biotic resources, with interaction networks between
328 different *taxa*, suggesting high diversity of functional features and ecosystem services.
329 Although it is still necessary to test the phylogenetic diversity of bats for the Atlantic Forest,
330 our congruence area corroborates with one of the refuge areas of the Atlantic Forest
331 (Carnaval; Moritz, 2008; Carnaval et al., 2009; Costa et al., 2018). A refuge area is a zone of
332 climatic stability which permitted, during interglacial periods, an accumulation of richness,
333 endemism and genetic diversity (Carnaval; Moritz, 2008; Carnaval et al., 2009).

334 In contradiction with the high bat biodiversity represented by the congruence area, this
335 area is currently one of the regions with the lowest protected area coverage rates for the biome
336 (Fig. 2). Most PAs in the Atlantic Forest are concentrated in the south region, especially in the
337 Serra do Mar (Sobral-Souza et al., 2018). Although 58 out of 59 bat species from this study
338 occur in PAs throughout the Atlantic Forest, these species scantily occur in PAs within the
339 congruence area. It should be emphasized that five bat species (*Lonchorhina aurita*,
340 *Macrophyllum macrophyllum*, *Phylloderma stenops*, *Thyroptera tricolor* and *Vampyroides*
341 *caraccioli*) failed to have any record in legally PAs within the congruence area, and *L. aurita*
342 is threatened in Brazil (ICMBio/MMA, 2018). Furthermore, the congruence area in this study
343 lies within one of the regions with the smallest number of forest remnants and with the

344 highest anthropogenic impacts, precisely within the Brazilian economic hub (Ribeiro et al.,
345 2009). This region is being strongly impacted by mining (Garcia et al., 2016), by the lack of
346 proper evaluation of long-term impacts (Fernandes et al. 2016) and by precarious regulations
347 (through legislation) of mining activities. For example, the rupture of iron residue dams in
348 Mariana (Neves et al. 2016) and Brumadinho (Cionek et al. 2019), in the state of Minas
349 Gerais, has been perhaps, one of the worst environmental disasters in the history of Brazil. It
350 was estimated that environmental and material losses caused by the Mariana disaster alone
351 sums more than 20 billion dollars (Fernandes et al. 2016).

352 Therefore, under conservation perspective, the area analyzed in this study indicates
353 one of the key-regions for the establishment of PAs for the protection of diversity and
354 ecosystems services of the Atlantic Forest. It represents an area with critical sites for the
355 conservation of different taxa; predictions on how bats would respond to climatic changes,
356 with loss of adequate environmental areas within the Brazilian savanna suggests that species
357 would mainly move towards the southeastern region, to the state of Minas Gerais (Aguiar et
358 al., 2016) (this is precisely the ecotone region between savanna and the Atlantic Forest where
359 the Bahia ecoregions and the reported congruence areas are found); and finally, it is a poorly
360 protected area and highly impacted by human activities. New PAs should be established
361 together with the expansion of existing Pas, where investments should applied on
362 reforestation techniques towards a more “biodiversity friendly” landscape by connecting
363 secondary forest fragments to PAs (Melo et al., 2013). It is important to note that our area of
364 congruence is one of the areas that deserve special attention when planning the zoning of
365 prioritary areas for conservation, but not the only one. Other sites along the biome are also
366 critical for conserving the diversity of other taxa (Ceballos; Ehrlich, 2006; Sigrist; Carvalho,
367 2008; Carnaval et al., 2014; Campos; Lourenço-De-Moraes, 2017); however, the biome center
368 is always indicated as one of those sites.

369 Since bats are a highly diversified group, using them as a model in the search for areas
370 with high taxonomic diversity and ecosystem services may be useful to pinpoint areas where
371 conservation efforts should be primarily focused upon. However, our results omit other
372 important areas for conservation, both for bats and for other taxa, and this may be due to the
373 fact that areas of endemism are analysis scale dependent (Daru et al., 2020). *Noctilio*
374 *albiventris*, for example, is not present in our congruence area, since its distribution records
375 for the Atlantic Forest are restricted to the south of the biome. This makes it necessary for the
376 analysis to be repeated at other scales, for example, by subdividing (north, center and south)

377 the Atlantic Forest. This should allow a more precise answer to the question about using bats
 378 and endemic area models to indicate PAs. Nevertheless, our study was able to find an
 379 important critical area presenting a deficit in the number and size of PAs. Therefore, we
 380 emphasize the urgency for the establishment of PAs within coastal and interior regions of the
 381 state of Bahia. As we stand in front of an economic crisis scene in a post-Covid-19 pandemic,
 382 where conservation financing will not be a priority, we should shed light in the opportunity
 383 brought by this pandemic, highlighting the relationships between conservation and
 384 environmental health, coupled to human health and well-being (Corlett et al 2020). It is
 385 primordial to increase pressure on public managers and decision-makers to strengthen
 386 conservation regulations, providing support for PAs (Corlett et al., 2020). Our analysis,
 387 therefore, results in an efficient strategy for the ecological-economic planning of the Atlantic
 388 Forest PAs, implying that efforts for the conservation of biodiversity and for the maintenance
 389 of ecosystem services will be more efficient in an increase in the cost-benefit relationship in
 390 the establishment of PAs.

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611

1 **6. CONCLUSÃO GERAL**

2 Esta tese foi desenvolvida com base em revisões e análises de dados
3 cuidadosas e criteriosas, permitindo o uso de seus resultados em trabalhos futuros
4 ou para tomas de decisões em conservação. O capítulo 1, fornece um confiável
5 banco de dados de ocorrência de morcegos na região de Londrina/PR e aponta as
6 possíveis lacunas de amostragem. Dessa forma, pode servir de base para estudos
7 ecológicos mais complexos e como orientação para que futuros projetos
8 desenvolvidos na região direcionem novos esforços de amostragem onde ainda
9 existem lacunas de conhecimento. Ainda, demonstra a importância da conservação
10 de vários fragmentos florestais de diferentes tipos (ex. UC's, reservas legais e áreas
11 verdes urbanas) na paisagem para manter a diversidade de morcegos na região de
12 Londrina. O capítulo 2, conseguiu explicar como ocorre a variação na composição
13 das comunidades de morcegos em toda a Mata Atlântica, indicando a relação da
14 riqueza com fatores climáticos e distância geográfica, associados aos gradientes
15 latitudinais e longitudinais do bioma. Essa complexa variação demonstrou a
16 importância de particionar a diversidade beta (substituição e aninhamento) e de
17 analisá-la em diferentes escalas para entender como a variação ambiental no
18 espaço geográfico influencia os padrões de distribuição das espécies. Os resultados
19 desse capítulo têm implicações práticas importantes pois ajuda a prever como as
20 mudanças ambientais podem interferir na distribuição das espécies, uma vez que
21 indicam a importância de fatores ambientais como preditores da composição das
22 espécies de morcegos. Sendo que essas preditores ambientais devem ser levados
23 em consideração nos estudos biogeográficos, ecológicos e de conservação,
24 inclusive na conservação prática. O capítulo 3, mostrou que o uso da análise de área
25 de endemismo com morcegos como modelo, pode possibilitar encontrar uma região
26 chave para investir em conservação e salvaguardar os atributos de diversidade e
27 serviços ecossistêmicos da Mata Atlântica. Além disso, o fato de a área de
28 congruência (de 58 das 59 espécies de morcegos utilizadas na análise) estar
29 localizada em uma zona de estabilidade climática, mas com alto impacto antrópico e
30 pouca proteção eficiente, indica a urgência na criação de UCs dentro das regiões da
31 costa e interior da Bahia. Dessa forma, a análise desse capítulo resulta em uma
32 estratégia mais eficaz no planejamento de zoneamento ecológico-econômico das
33 áreas prioritárias na Mata Atlântica, implicando que os esforços para a conservação

34 da biodiversidade e manutenção dos serviços ecossistêmicos serão mais eficientes
35 em cenários climáticos futuros, aumentando a relação custo-benefício na criação de
36 UCs.

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APÊNDICES

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APÊNDICE A

Material suplementar do Capítulo 1. O material encontra-se nas normas do periódico "Neotropical Biology and Conservation".

50 SUPPORTING INFORMATION

Table 1. References used to provide data on bat occurrence records for the Londrina region, Brazil, with their respective period of capture. Captured sites were: Parque Estadual Mata dos Godoy (PEMG); Parque Municipal Arthur Thomas (PMAT); campus of the Universidade Estadual de Londrina (UEL) (including Horto Florestal); Jardim Botânico; Legal Reserves within private farmland (Mata do Bule, Fazenda Imbaúva, Chácara Kanashiro, Chácara Vitória and Fazenda Regina); and the metropolitan region of Londrina (including banks of the Igapó Lake; and other green areas in the municipality and human building, both sites without georeferenced exact indication).

Publications	Place of capture	Period of capture	Source
Sekiana (1992)	PEMG; PMAT; UEL	1982 – 1993; 1989 – 1992.	Undergraduate dissertation.
Muller and Reis (1992)	PEMG; PMAT	1989 - 1992	Published article; data were retrieved from field notes since the publication does not give a complete list of captured species.
Reis <i>et al.</i> (1993)	PEMG; PMAT; UEL; Fazenda Imbaúva; Chácara Kanashiro	1982 a 1983	Published article.
Lima (1994)	UEL	1993	Undergraduate dissertation and field notebook to supplement information.

Reis and Muller (1995)	PEMG; PMAT; UEL; Fazenda Regina	1989 - 1992	Published article.
Sekiana (1996)	PEMG	1993 - 1994	Masters Dissertation.
Lima, I. P.	UEL	1998	Non-published data; field notes by researcher from the mammal ecology laboratory of the UEL.
Reis <i>et al.</i> (2000)	PEMG	1993 - 1994	Published article.
Félix <i>et al.</i> (2001)	PMAT; UEL	1999 - 2000	Published article: data retrieved from field notebook to supplement information.
Reis <i>et al.</i> (2006)	UEL and other sites in metropolitan Londrina region	1998 - 1999	Published article; however, data were retrieved from field notebook. The article failed to give the exact site of capture of each species.
Reis <i>et al.</i> (2012a)	PEMG; PMAT; UEL; Fazenda Regina; Chácara Vitória; and other sites in metropolitan Londrina region	1982 - 2012	Published article.

Reis <i>et al.</i> (2012b)	UEL	2008 - 2009	Published article.
Fregonezi (2012)	PEMG	2011	Master Dissertation.
Batista, C. B.	PEMG	2013 - 2014	Master Dissertation data, retrieved from field notebook.
Batista, C. B.	PEMG; UEL;	2015 - 2016	Non-published data; field notes by researcher from the Laboratório de Ecologia de Mamíferos of the UEL.
Batista <i>et al.</i> , 2018	Mata do Bule and Jardim Botânico de Londrina	2015 - 2016	Published article.

Table 2. List of species reported for the region of Londrina, PR, Brazil, in published articles and unpublished data included in this study. Preferential food: C- Carnivorous; I – Insectivorous; H – Hematophagous; O - Omnivorous; N – Nectarivorous; F – Frugivorous; P - Piscivorous.. The capture sites are: two Protected Areas in the region – Parque Estadual Mata dos Godoy (PEMG) and Parque Municipal Arthur Thomas (PMAT) -; the Legal Reserves (Mata do Bule, Fazenda Imbaúva, Chácara Kanashiro, Chácara Vitória and Fazenda Regina); and the metropolitan region of Londrina, which includes the campus of the Universidade Estadual de Londrina (UEL) and its Horto Florestal; the Jardim Botânico de Londrina; banks of the Igapó Lake; other green areas in the municipality (non-identified valleys) and human building (sites without exact indication). The threat categories were based on a regional (RE) (Margarido and Braga, 2004) National (NA) (ICMBio, 2014) and international (IN) (IUCN) list and are: LC- Least concern; VU - Vulnerable; CR - Critically Endangered; DD - Data deficient; and NT - Near threatened.

Specie	Guild	Conservation status	Protected Areas		Other sites
			PEMG PMAT		
Family Phyllostomidae					
Subfamily Micronycterinae					
<i>Micronycteris microtis</i> Miller, 1898	I	LC (NA; IN)	X	X	Fazenda Regina; Jardim Botânico
Subfamily Desmodontinae					

<i>Desmodus rotundus</i> (É Geoffroy, 1810)	H	LC (NA; IN)	X	Fazenda Regina
<i>Diaemus youngii</i> (Jentink, 1893)	H	LC (NA; IN); CR (RE)	X	
<i>Diphylla ecaudata</i> (Spix, 1823)	H	LC (NA; IN); VU (RE)	X	
Subfamily Phyllostominae				
<i>Chrotopterus auritus</i> (Peters, 1856)	C	LC (NA; IN); VU (RE)	X X	Fazenda Regina;
<i>Phyllostomus discolor</i> (Wagner, 1843)	O	LC (NA; IN)		Sites without exact indication, but sampling included UEL, Chácara Vitória, banks of the Igapó Lake and non-identified valleys
<i>Phyllostomus hastatus</i> (Pallas, 1767)	O	LC (NA; IN)	X	UEL; Jardim Botânico
Subfamily Glossophaginae				
<i>Anoura caudifer</i> (É. Geoffroy, 1818)	N	LC (NA; IN)	X X	Jardim Botânico
<i>Glossophaga soricina</i> (Pallas, 1766)	N	LC (NA; IN)	X	UEL

Subfamily Carrolliinae					
<i>Carollia perspicillata</i> (Linnaeus, 1758)	F	LC (NA; IN)	X	X	Fazenda Regina; Mata do Bule; UEL; Jardim Botânico and other sites without exact indication, but sampling included Chácara Vitória, banks of the Igapó Lake and non-identified valleys
Subfamily Stenodermatinae					
<i>Artibeus fimbriatus</i> Gray, 1838	F	LC (NA; IN)	X	X	Mata do Bule; UEL; Jardim Botânico and other sites without exact indication, but sampling included Chácara Vitória, banks of the Igapó Lake and non-identified valleys
<i>Artibeus lituratus</i> (Olfers, 1818)	F	LC (NA; IN)	X	X	Fazenda Imbaúva; Fazenda Regina; Chácara Kanashiro; Mata do Bule; UEL; Jardim Botânico; non-identified valleys ; Human buildings (includes roof for houses and buildings, manholes, crevices and others)

<i>Artibeus obscurus</i> (Schinz, 1821)	F	LC (NA; IN)	X	Sites without exact indication, but sampling included UEL, Chácara Vitória, banks of the Igapó Lake and non-identified valleys
<i>Artibeus planirostris</i> (Spix, 1823)	F	LC (NA; IN)	X X	Mata do Bule
<i>Chiroderma doriae</i> Thomas, 1891	F	LC (NA; IN); VU (RE)		Fazenda Regina
<i>Chiroderma villosum</i> Peters, 1860	F	LC (NA; IN); DD (RE)		Sites without exact indication, but sampling included UEL, Chácara Vitória, banks of the Igapó Lake and non-identified valleys
<i>Platyrrhinus lineatus</i> (É. Geoffroy, 1810)	F	LC (NA; IN)	X X	Fazenda Regina; UEL; Jardim Botânico and other sites without exact indication, but sampling included UEL, Chácara Vitória, banks of the Igapó Lake and non-identified valleys; Human buildings (includes roof for houses and buildings, manholes, crevices and others)

<i>Pygoderma bilabiatum</i> (Wagner, 1843)	F	LC (NA; IN)	X	X	Fazenda Regina; Mata do Bule; UEL; Jardim Botânico and other sites without exact indication, but sampling included UEL, Chácara Vitória, banks of the Igapó Lake and non-identified valleys
<i>Sturnira lilium</i> (É. Geoffroy, 1810)	F	LC (NA; IN)	X	X	Fazenda Regina; Chácara Kanashiro; Mata do Bule; UEL; Jardim Botânico and other sites without exact indication, but sampling included Chácara Vitória, banks of the Igapó Lake and non-identified valleys
<i>Vampyressa pusilla</i> (Wagner, 1843)	F	LC (NA); DD (IN)	X	X	UEL
Family Noctilionidae					
<i>Noctilio albiventris</i> Desmarest, 1818	I	LC (NA; IN)			banks of the Igapó Lake 23°19'34.24"S 51°10'27.43"O

<i>Noctilio leporinus</i> (Linnaeus, 1758)	P	LC (NA; IN)		Sites without exact indication, but sampling included UEL, Chácara Vitória, banks of the Igapó Lake and non-identified valleys
Family Molossidae				
Subfamily Molossinae				
<i>Eumops glaucinus</i> (Wagner, 1843)	I	LC (NA; IN)		Human buildings (includes roof for houses and buildings, manholes, crevices and others)
<i>Eumops perotis</i> (Schinz, 1821)	I	LC (NA; IN)		Metropolitan region of Londrina, without exact site indication.
<i>Cynomops abrasus</i> (Temnick, 1827)	I	LC (NA); DD (IN)	X	UEL and other sites without exact indication, but sampling included, Chácara Vitória, banks of the Igapó Lake and non-identified valleys

<i>Molossus molossus</i> (Pallas, 1766)	I	LC (NA; IN)	X	UEL; Jardim botânico and other sites without exact indication, but sampling included, Chácara Vitória, banks of the Igapó Lake and non-identified valleys; Human buildings (includes roof for houses and buildings, manholes, crevices and others)
<i>Molossus rufus</i> É. Geoffroy, 1805	I	LC (NA; IN)	X	UEL; Jardim botânico; Human buildings (includes roof for houses and buildings, manholes, crevices and others)
<i>Nyctinomops laticaudatus</i> (É. Geoffroy, 1805)	I	LC (NA; IN)	X	UEL; Human buildings (includes roof for houses and buildings, manholes, crevices and others)
<i>Nyctinomops macrotis</i> (Gray, 1840)	I	LC (NA; IN)		Human buildings (includes roof for houses and buildings, manholes, crevices and others)

<p><i>Tadarida brasiliensis</i> (I. Geoffroy, 1824)</p>	I	LC (NA; IN)		<p>UEL and other sites without exact indication, but sampling included, Chácara Vitória, banks of the Igapó Lake and non-identified valleys; Human buildings (includes roof for houses and buildings, manholes, crevices and others)</p>
<p>Family Vespertilionidae</p> <p>Subfamily Vespertilioninae</p>				
<p><i>Eptesicus brasiliensis</i> (Desmarest, 1819)</p>	I	LC (NA; IN)	X X	<p>Fazenda Imbaúva; other sites without exact indication, but sampling included UEL, Chácara Vitória, banks of the Igapó Lake and non-identified valleys; Human buildings (includes roof for houses and buildings, manholes, crevices and others)</p>

<i>Eptesicus diminutus</i> Osgood, 1915	I	LC (NA); DD (IN)	X X	Mata do Bule; Jardim Botânico; UEL and other sites without exact indication, but sampling included Chácara Vitória, banks of the Igapó Lake and non-identified valleys
<i>Eptesicus furinalis</i> (d'Orbigny & Gervais, 1847)	I	LC (NA; IN)	X	Mata do Bule; Jardim Botânico; UEL and other sites without exact indication, but sampling included Chácara Vitória, banks of the Igapó Lake and non-identified valleys
<i>Epitesicus taddei</i> Miranda, Bernardi & Passos, 2006	I	VU (NA); DD (IN)		Sites without exact indication, but sampling included Chácara Vitória, banks of the Igapó Lake and non-identified valleys
<i>Rogheessa io</i> Thomas, 1903	I	DD (NA); LC (IN)		Fazenda Regina
<i>Histiopus velatus</i> (I. Geoffroy, 1824)	I	LC (NA); DD (IN)	X	Human buildings (includes roof for houses and buildings, manholes, crevices and others)

<i>Lasiurus blossevillii</i> ([Lesson, 1826])	I	LC (NA; IN)	X	X	UEL and other sites without exact indication, but sampling included Chácara Vitória, banks of the Igapó Lake and non-identified valleys
<i>Lasiurus ega</i> (Gervais, 1856)	I	LC (NA; IN)	X		Human buildings (includes roof for houses and buildings, manholes, crevices and others)
Subfamily Myotinae					
<i>Myotis levis</i> (I. Geoffroy, 1824)	I	LC (NA; IN)			Sites without exact indication, but sampling included UEL, Chácara Vitória, banks of the Igapó Lake and non-identified valleys
<i>Myotis nigricans</i> (Schinz, 1821)	I	LC (NA; IN)	X	X	Fazenda Regina; UEL and other sites without exact indication, but sampling included Chácara Vitória, banks of the Igapó Lake and non-identified valleys

<i>Myotis ruber</i> (É. Geoffroy, 1806)	I	LC (NA); NT -IN; DD - RE	X	X	Mata do Bule and other sites without exact indication, but sampling included Chácara Vitória, banks of the Igapó Lake and non-identified valleys
RICHNESS			28	17	39
Total of species in Londrina region, Paraná state			41		

Table 3. Information on capture environments and number of nights sampled at each site in the Londrina region (Paraná, Brazil), studied for bats since 1982. PEMG – Parque Estadual Mata dos Godoy; PMAT – Parque Municipal Arthur Thomas; UEL – Campus da Universidade Estadual de Londrina. The asterisk highlight that for the PEMG information about the capture environments is not published in any of the related articles. This information was obtained directly from researchers at the Laboratório de Ecologia de Mamíferos of the Universidade Estadual de Londrina. The references that were used to compose the species list (See in Table 1 of the manuscript) and are not in this table is because they do not have information about the capture environment within the fragment or about the sampling effort.

Site	Capture environments	Number of nights sampled	Source
PEMG*	Visitor trails: Trilha das Perobas and Trilha dos Catetos	12	Sekiama (1992)
	Visitor trails: Trilha das Perobas and Trilha dos Catetos	23	Muller & Reis (1992)

	Visitor trails: Trilha das Perobas and Trilha dos Catetos	23	Reis <i>et al.</i> (1993)
	Visitor trails: Trilha das Perobas and Trilha dos Catetos	24	Reis & Muller (1995)
	Visitor trails: Trilha das Perobas and Trilha dos Catetos	36	Sekiana (1996)
	Visitor trails: Trilha das Perobas and Trilha dos Catetos	12	Reis <i>et al.</i> (2000)
	Visitor trails: Trilha das Perobas and Trilha dos Catetos	48	Fregonezi (2012)
	Visitor trails: Trilha das Perobas and Trilha dos Catetos	60	Batista, C. B. (Dados não publicados da dissertação e da tese. Coletados entre 2013 e 2016)
Total capture nights	238		
PMAT	No informations	12	Sekiana (1992)
	No informations	23	Muller & Reis (1992)
	No informations	23	Reis <i>et al.</i> (1993)
	No informations	24	Reis & Muller (1995)
	The mist nets were spread throughout the Park; installed on visitor trails at	24	Félix <i>et al.</i> (2001)

	the edge of the fragment; on the trails inside the fragment; near the waterfall and around the lake.		
Total capture nights	106		
UEL	No informations	12	Sekiana (1992)
	No informations	23	Reis <i>et al.</i> (1993)
	The mist nets were spread throughout the campus da Universidade Estadual de Londrina; placed on the streets and in the groves of fruit plants.	12	Lima (1994)
	No informations	24	Reis & Muller (1995)
	The mist nets were spread throughout the campus da Universidade Estadual de Londrina; placed on the streets and in the groves of fruit plants.	2	Lima, I. P. (unpublished data)
	The mist nets were spread throughout the campus da Universidade Estadual de Londrina; placed on the streets and in the groves of fruit plants frutíferas and in the Horto Florestal.	24	Félix <i>et al.</i> (2001)
	The mist nets were spread throughout the campus da Universidade Estadual de Londrina; captures were	12	Reis <i>et al.</i> (2006)

	also made in buildings of the Center for Biological Sciences (CCB)		
	Horto Florestal	12	Reis <i>et al.</i> (2012b)
	Horto Florestal	12	Batista, C. B. (unpublished data)
Total capture nights	133		
Jardim Botânico	The mist nets were installed in the eastern portion of the Jardim Botânico: walking visitor's tracks; trails within the forest fragment; around the artificial lake; in the area of bamboo and fruit groves.	16	Batista <i>et al.</i> , 2018
Total capture nights	16		
Chácara Vitória	No informations	No informations	Reis <i>et al.</i> (2012a)
Total capture nights	-		
Chácara Kaneshiro	No informations	23	Reis <i>et al.</i> (1993)
Total capture nights	23		
Fazenda Imbaúva	No informations	23	Reis <i>et al.</i> (1993)
Total capture nights	23		
Fazenda Regina	No informations	24	Reis & Muller (1995)
Total capture nights	24		

Mata do Bule	The captures were made at random points to cover the northern, southern, eastern and western regions of the fragment. The mist-nets were installed inside trails, clearings, at the edges of the fragment and around a dam.	14	Batista <i>et al.</i> , 2018
Total capture nights	14		
Total capture nights to Londrina region	577		

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APÊNDICE B

Material suplementar do Capítulo 2. O material encontra-se nas normas do periódico “Journal of Biogeography”.

SUPPORTING INFORMATION

8

Beta diversity patterns of bats in the Atlantic Forest: How does the scale of analysis affect the importance of spatial and environmental factors?

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Short running title: Beta diversity of bats in the Atlantic Forest

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19 APPENDIX S1

20 **Table S1.1** List of bat species and occurrence sites (coordinates) for the Atlantic Forest ecoregions. The number represent the total number of records
 21 (occurrence site) of each specie in a specific ecoregion. The table also shows other information for the species: preferential food, occurrence biomes (Reis et
 22 al., 2017) and conservation status (IUCN, 2016). Preferential food: O/C - Omnivorous/Carnivorous; I – Insectivorous; H – Hematophagous; O - Omnivorous;
 23 N – Nectarivorous; F – Frugivorous; P - Piscivorous. Occurrence in other biomes: A – Amazon; C – Cerrado; Ca – Caatinga; P – Pantanal; Pa – Pampas.
 24 Conservation status: LC – Least Concern; DD – Data Deficient and NT – Near Threatened

Species	Occurrence in other biomes	Preferential food	Conservation status	Occurrence sites in ecoregions										Total occurrence sites
				Alto Parana	Araucaria	Restinga Coast	Bahia Coastal	Bahia Interior	Campos	Rupestres	remambuco Coastal	remambuco Interior	Serra do Mar	
Family Phyllostomidae														
Subfamily Micronycterinae														
<i>Lamproncyteris brachyotis</i> (Dobson, 1879)	C and A	O/C	LC	0	0	0	2	0	0	0	0	2	4	
<i>Micronycteris hirsuta</i> (Peters, 1869)	A	I	LC	0	0	0	6	1	0	0	0	1	8	
<i>Micronycteris megalotis</i> (Gray, 1842)	A, C, Ca and P	I	LC	17	0	0	13	5	0	1	0	20	56	
<i>Micronycteris microtis</i> Miller, 1898	A and C	I	LC	5	0	0	3	1	2	0	0	8	19	

<i>Mimon bennettii</i> (Gray, 1838)	A, Ca, C and P	I	LC	5	2	0	4	2	1	0	0	14	28
<i>Mimon crenulatum</i> (É. Geoffroy, 1803)	A, Ca, C and P	I	LC	1	0	0	3	2	0	0	0	2	8
<i>Phylloderma stenops</i> (Peters, 1865)	A, C and P	O	LC	1	0	0	0	0	0	1	0	2	4
<i>Phyllostomus discolor</i> (Wagner, 1843)	A, Ca, C and P	O	LC	9	0	1	11	7	0	2	1	2	33
<i>Phyllostomus elongatus</i> (É. Geoffroy, 1810)	A, C and P	O	LC	0	0	0	2	0	0	2	0	0	4
<i>Phyllostomus hastatus</i> (Pallas, 1767)	A, Ca, C and P	O	LC	17	0	1	18	8	0	2	1	17	65
<i>Tonatia bidens</i> (Spix, 1823)	P, C and Ca	O	DD	5	0	0	3	2	1	1	0	25	37
<i>Tonatia saurophila</i> Koopman & Williams, 1951	Ca, C and A	O	LC	0	0	0	2	0	0	1	1	1	5
<i>Trachops cirrhosus</i> (Spix, 1823)	A, C, Ca and P	O	LC	1	1	1	17	2	0	4	2	16	44
Subfamily Glossophaginae													
<i>Anoura caudifer</i> (É. Geoffroy, 1818)	A, C, P and Pa	N	LC	19	2	1	1	11	2	1	0	40	77
<i>Anoura geoffroyi</i> Gray, 1838	A, Ca, C, P and Pa	N	LC	4	2	0	2	4	1	1	0	29	43
<i>Dryadonycteris capixaba</i> Nogueira, Lima, Peracchi & Simmons, 2012	Ca and C	N	DD	0	0	0	3	2	0	0	0	0	5

<i>Glossophaga soricina</i> (Pallas, 1766)	A, Ca, C, P and Pa	N	LC	91	1	4	30	15	1	3	1	59	206
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<i>Lichonycteris degener</i> Miller, 1931	A	N	LC	0	0	0	1	0	0	1	0	0	2
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Subfamily Lonchophyllinae

<i>Lonchophylla mordax</i> Thomas, 1903	Ca	N	LC	0	0	0	10	0	0	0	0	7	17
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<i>Lonchophylla peracchii</i> Dias, Esbérard & Moratelli, 2013	Ca	N	NC	0	0	0	1	2	0	0	0	16	19
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Subfamily Carollinae

<i>Carollia brevicauda</i> (Schinz, 1821)	A, Ca, C, P	F	LC	0	0	0	11	0	0	0	0	2	13
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<i>Carollia perspicillata</i> (Linnaeus, 1758)	A, Ca, C, P and Pa	F	LC	80	6	3	88	23	2	4	1	79	287
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Subfamily Glyphonycterinae

<i>Glyphonycteris daviesi</i> (Hill, 1964)	A	I	LC	0	0	0	1	0	0	0	0	0	1
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<i>Glyphonycteris sylvestris</i> Thomas, 1896	A	I	LC	2	0	0	2	0	0	0	0	0	4
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<i>Trinycteris nicefori</i> (Sanborn, 1949)	A	I	LC	0	0	0	2	0	0	1	0	0	3
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Subfamily Rhinophyllinae

<i>Rhinophylla pumilio</i> Peters, 1865	A; C and Ca	F	LC	0	0	0	49	2	0	2	0	0	53
Subfamily Stenodermatinae													
<i>Dermanura cinerea</i> Gervais, 1856	A, C and Ca	F	LC	0	0	0	24	3	0	3	1	13	45
<i>Artibeus fimbriatus</i> Gray, 1838	A, Ca, C, P and Pa	F	LC	5	2	1	34	5	1	0	0	30	78
<i>Dermanura gnoma</i> (Handley, 1987)	A, C and Ca	F	LC	0	0	0	5	0	0	0	0	0	5
<i>Artibeus lituratus</i> (Olfers, 1818)	A, Ca, C, P and Pa	F	LC	46	4	4	55	16	2	3	1	61	192
<i>Artibeus obscurus</i> (Schinz, 1821)	A, Ca, C, P and Pa	F	LC	8	1	1	39	5	1	3	1	41	100
<i>Artibeus planirostris</i> (Spix, 1823)	A, Ca, C, P and Pa	F	LC	31	0	2	9	4	0	3	1	32	82
<i>Chiroderma doriae</i> Thomas, 1891	C and P	F	LC	24	0	0	1	1	0	2	1	30	59
<i>Chiroderma villosum</i> Peters, 1860	A, C and P	F	LC	9	0	2	7	2	0	1	0	15	36
<i>Platyrrhinus incarum</i> (Thomas, 1912)	A	F	LC	0	0	0	1	0	0	1	0	0	2
<i>Platyrrhinus lineatus</i> (É. Geoffroy, 1810)	A, Ca, C, P and Pa	F	LC	67	1	3	28	15	0	4	1	40	159
<i>Platyrrhinus recifinus</i> (Thomas, 1901)	C and Ca	F	LC	5	1	1	13	4	2	0	0	28	54
<i>Pygoderma bilabiatum</i> (Wagner, 1843)	C, Ca and P	F	LC	23	6	0	6	7	0	1	0	28	71

<i>Sturnira lilium</i> (É. Geoffroy, 1810)	A, Ca, C, P and Pa	F	LC	50	11	2	27	19	3	3	1	55	171
<i>Sturnira tildae</i> de la Torre, 1959	A, Ca, C and P	F	LC	9	3	0	3	0	1	0	0	14	30
<i>Uroderma bilobatum</i> Peters, 1866	A, Ca, C, P and Pa	F	LC	0	0	0	4	1	0	0	0	0	5
<i>Uroderma magnirostrum</i> Davis, 1968	A, Ca, C, P and Pa	F	LC	0	0	1	3	1	0	0	0	1	6
<i>Vampyressa pusilla</i> (Wagner, 1843)	C	F	DD	17	0	0	4	7	0	1	0	33	62
<i>Vampyrodes caraccioli</i> (Thomas, 1889)	A	F	LC	0	0	0	1	0	0	0	0	3	4

Family Noctilionidae

<i>Noctilio albiventris</i> Desmarest, 1818	A, Ca, C and P	I	LC	16	0	0	0	0	0	0	0	1	17
<i>Noctilio leporinus</i> (Linnaeus, 1758)	A, Ca, C and P	P	LC	7	0	1	6	5	0	0	1	13	33

Family Furipteridae

<i>Furipterus horrens</i> (Cuvier, 1828)	A, C and Ca	I	LC	3	0	0	1	0	0	0	0	4	8
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Family Thyropteridae

<i>Thyroptera tricolor</i> Spix, 1823	A	I	LC	0	0	0	0	1	0	0	0	3	4
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Family Natalidae

<i>Natalus macrourus</i> (Gervais, 1856)	A, C and Ca	I	NT	1	0	0	0	1	0	0	0	0	2
<i>Total of Species in ecoregions</i>				37	17	19	53	38	15	32	18	45	59

APPENDIX S2

Rasters of environmental variables and graphic representations of splines that were generated based on environmental variables included in the GDM of bat species composition (β_{sim}) in the Atlantic Forest. The x axis is the predicted ecological distance and y axis is the observed distance. The height of spline indicates the magnitude of biological change along that gradient and represents the relative importance of each predictor contributing to the turnover component. The shape of the spline indicates where, along each gradient, the changes are most pronounced.

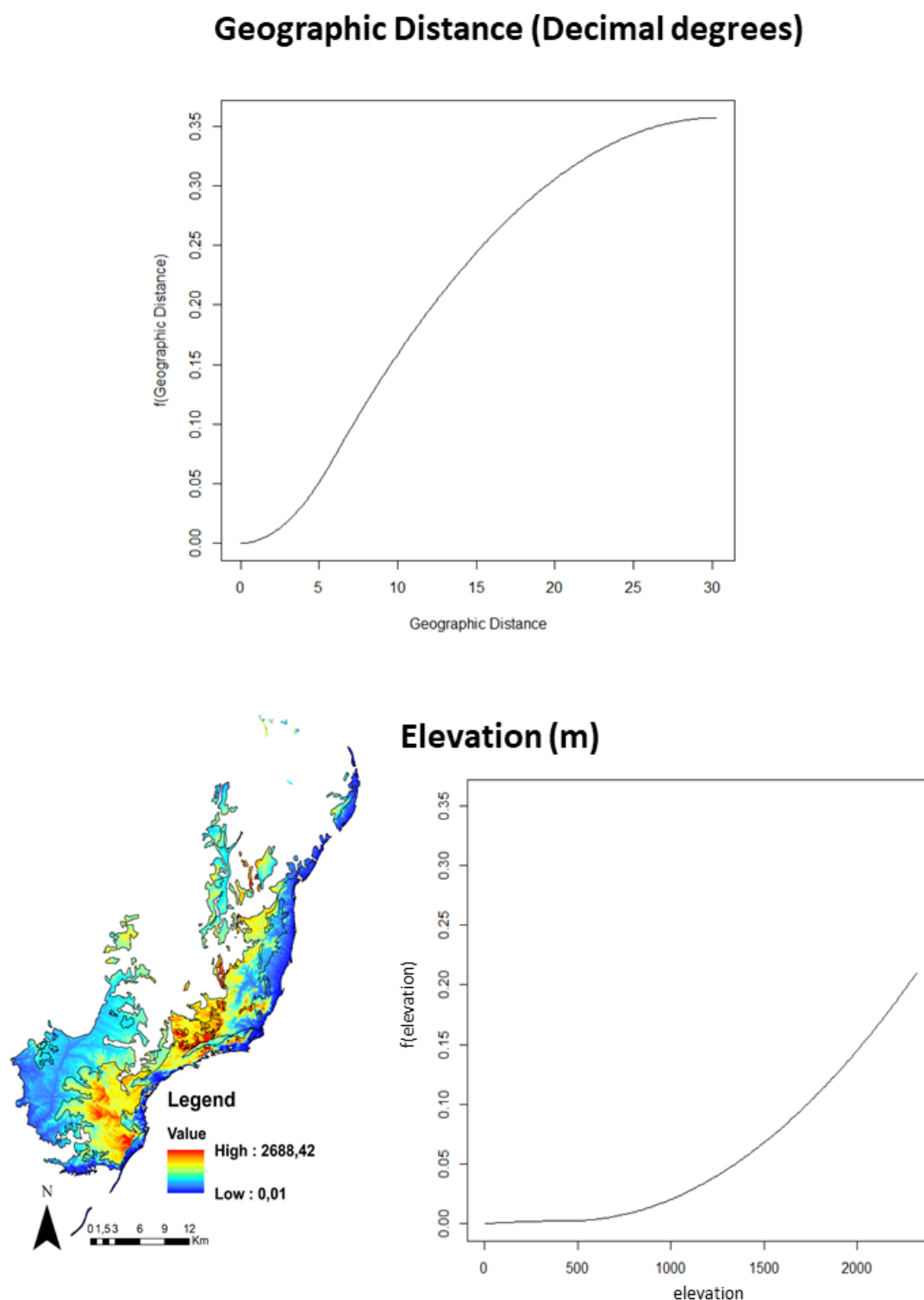
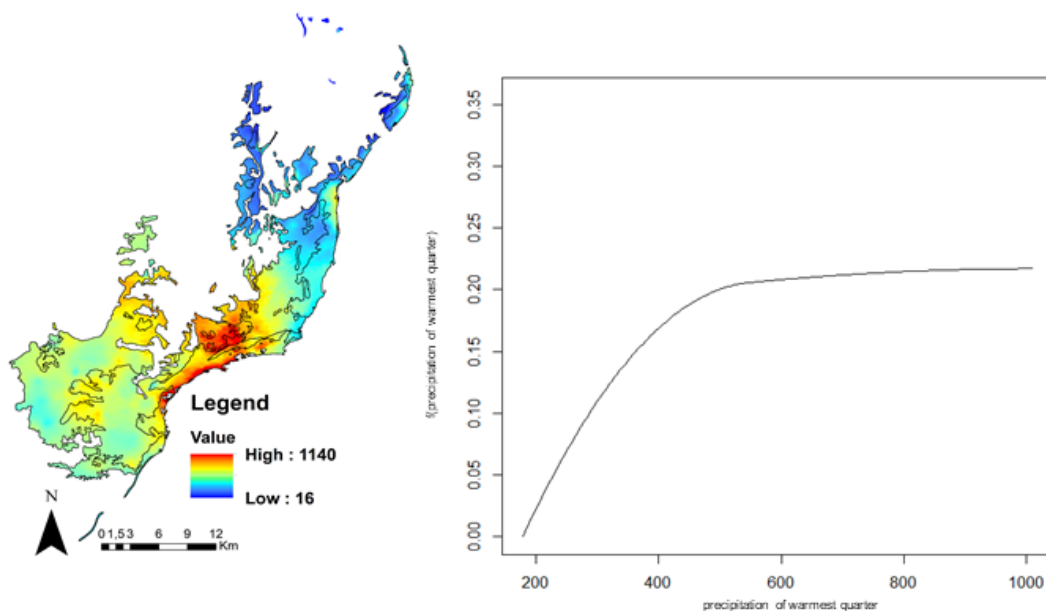


Figure S1.2 Rasters of geographical distance and altitude, and graphic representations of splines that were generated based on environmental variables included in the GDM of bat species composition (β_{sim}) in the Atlantic Forest. The x axis is the predicted ecological distance and y axis is the observed distance. The height of spline indicates the magnitude of biological change along that gradient and represents the relative importance of each predictor contributing to the turnover component. The shape of the spline indicates where, along each

gradient, the changes are most pronounced.

Precipitation of warmest quarter (mm)



Precipitation of coldest quarter (mm)

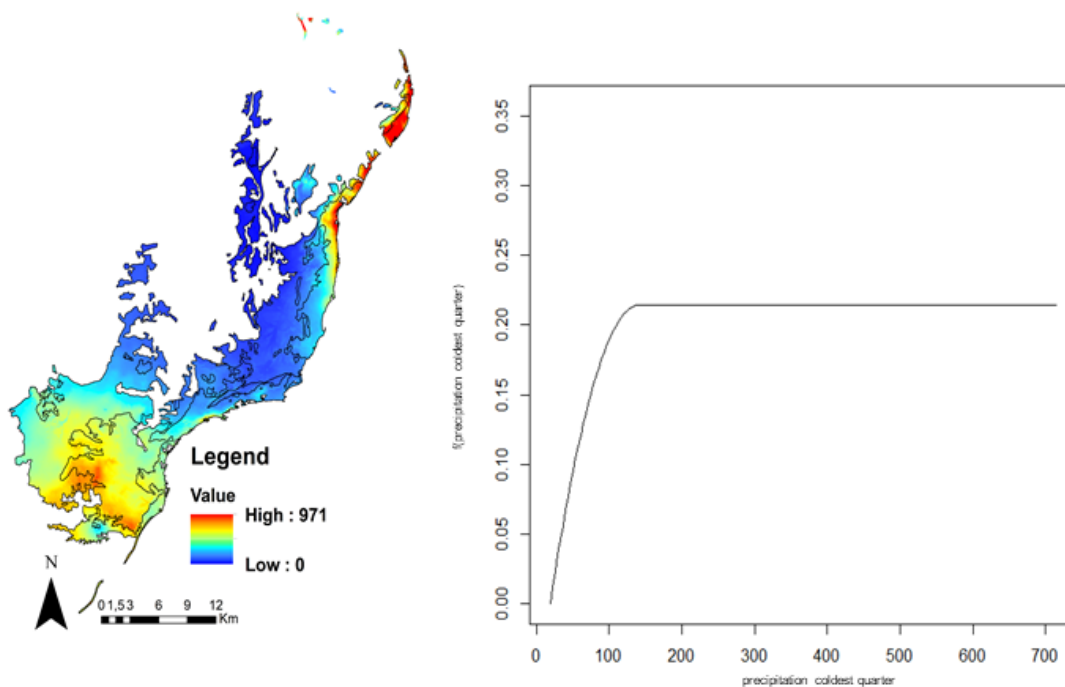
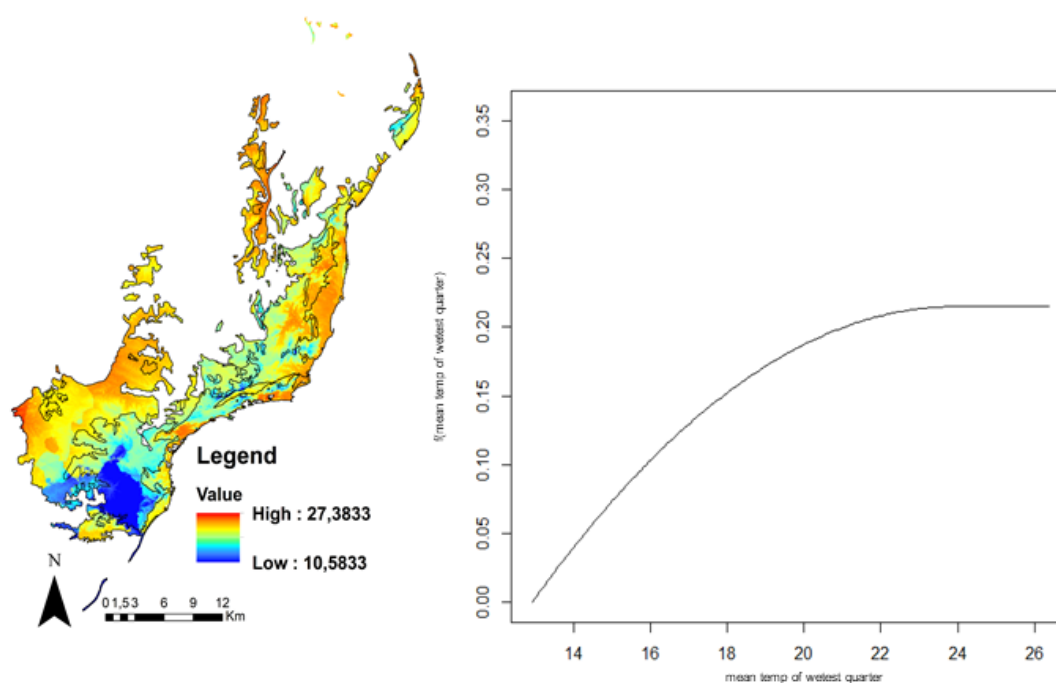


Figure S2.2 Rasters of precipitation of warmest quarter and precipitation of coldest quarter, and graphic representations of splines that were generated based on environmental variables included in the GDM of bat species composition (β_{sim}) in the Atlantic Forest. The x axis is the predicted ecological distance and y axis is the observed distance. The height of spline indicates the magnitude of biological change along that gradient and represents the relative

importance of each predictor contributing to the turnover component. The shape of the spline indicates where, along each gradient, the changes are most pronounced.

Mean temperature of wettest quarter (°C)



Minimum temperature of coldest month (°C)

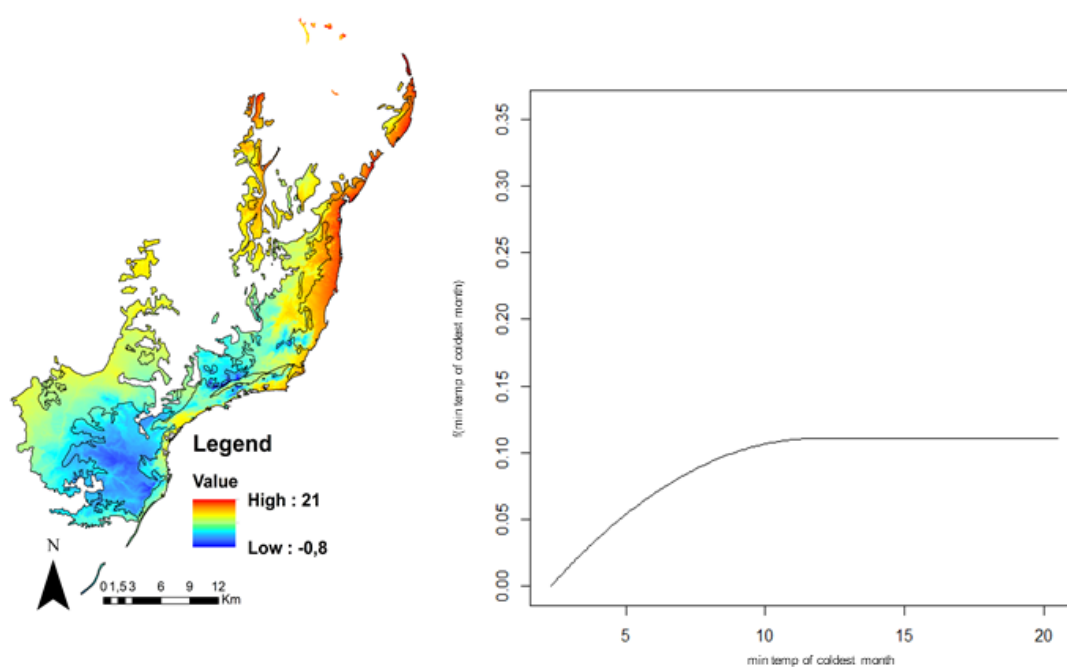


Figure S3.2 Rasters of mean temperature of wettest quarter and minimum temperature of coldest month, and graphic representations of splines that were generated based on environmental variables included in the GDM of bat species composition (β_{sim}) in the Atlantic Forest. The x axis is the predicted ecological distance and y axis is the observed

distance. The height of spline indicates the magnitude of biological change along that gradient and represents the relative importance of each predictor contributing to the turnover component. The shape of the spline indicates where, along each gradient, the changes are most pronounced.

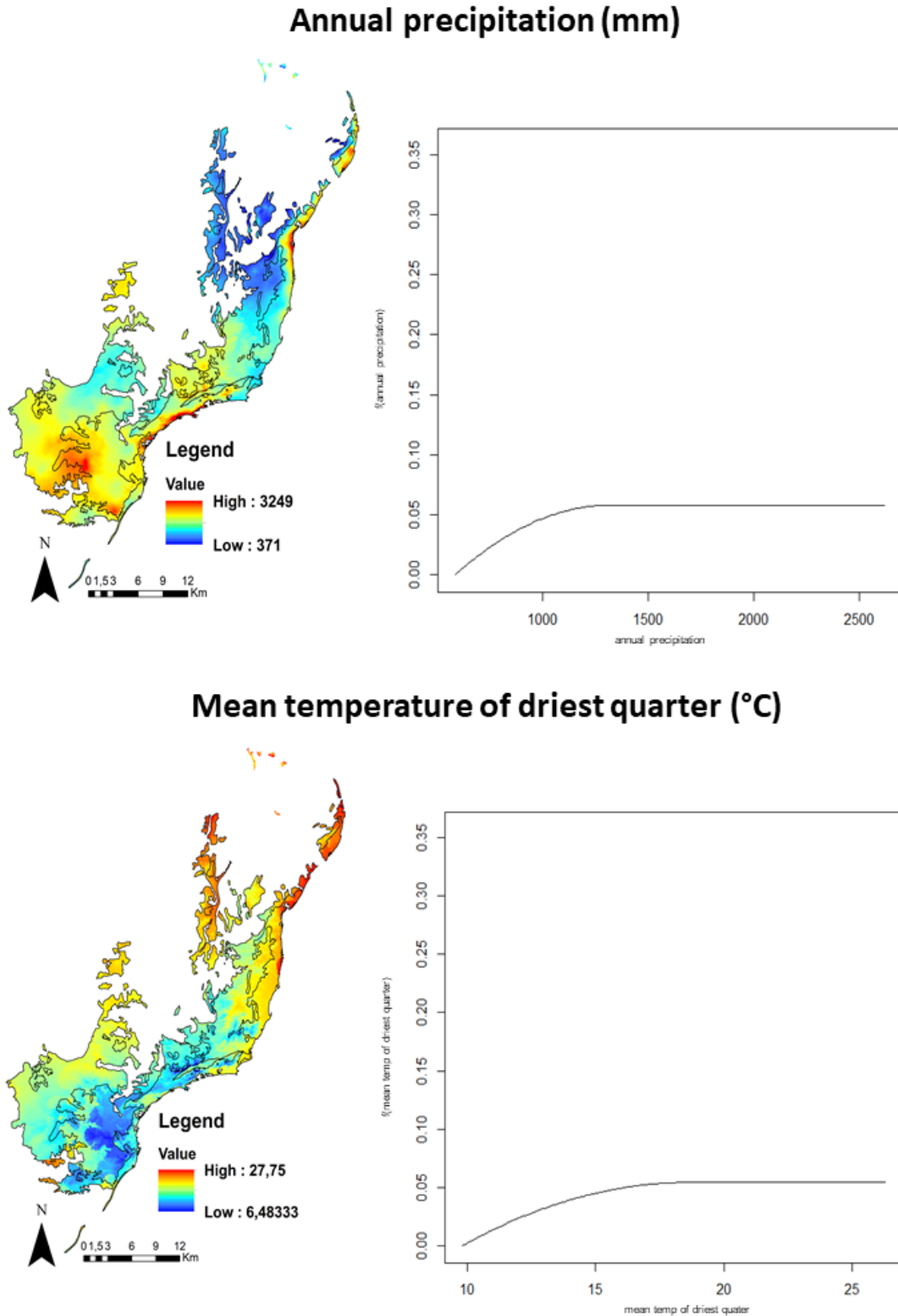
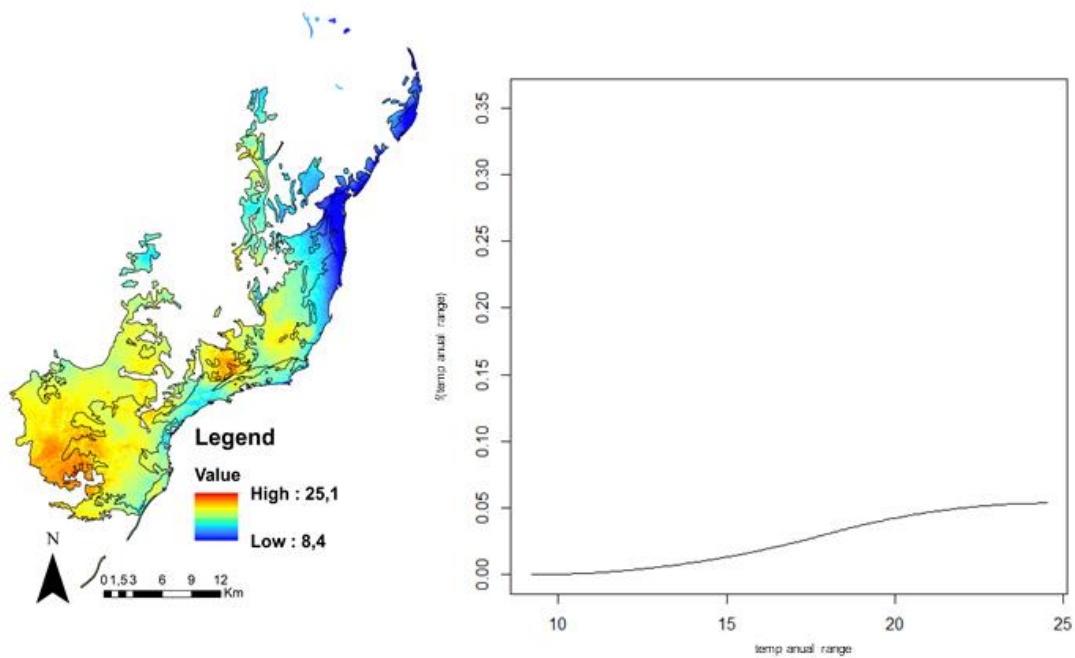


Figure S4.2 Rasters of annual precipitation and mean temperature of driest quarter, and graphic representations of splines that were generated based on environmental variables

included in the GDM of bat species composition (β_{sim}) in the Atlantic Forest. The x axis is the predicted ecological distance and y axis is the observed distance. The height of spline indicates the magnitude of biological change along that gradient and represents the relative importance of each predictor contributing to the turnover component. The shape of the spline indicates where, along each gradient, the changes are most pronounced.

Temperature anual range (°C)



Mean diurnal range (°C)

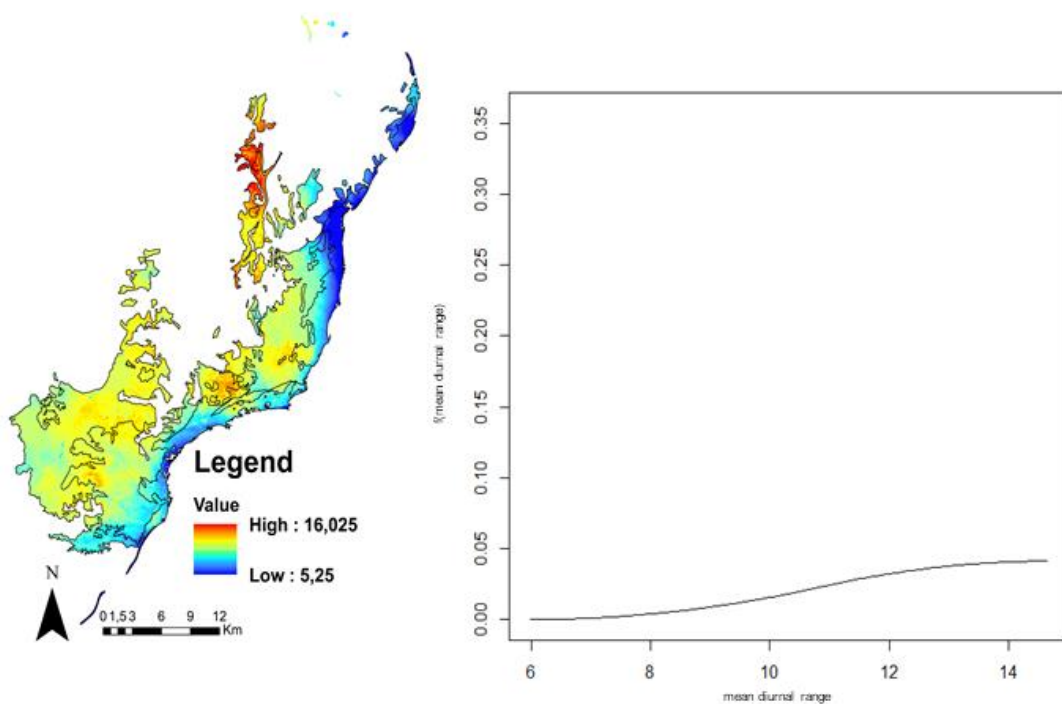


Figure S5.2 Rasters of temperature annual range and mean diurnal range, and graphic

representations of splines that were generated based on environmental variables included in the GDM of bat species composition (β_{sim}) in the Atlantic Forest. The x axis is the predicted ecological distance and y axis is the observed distance. The height of spline indicates the magnitude of biological change along that gradient and represents the relative importance of each predictor contributing to the turnover component. The shape of the spline indicates where, along each gradient, the changes are most pronounced.

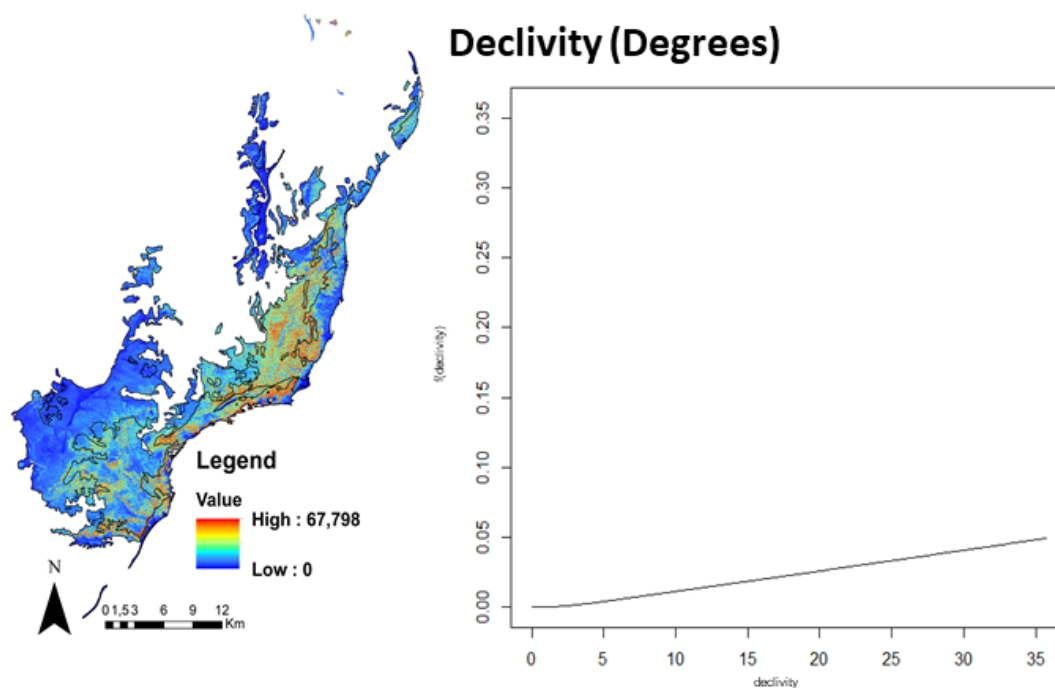


Figure S6.2 Raster of declivity graphic representations of splines that were generated based on environmental variables included in the GDM of bat species composition (β_{sim}) in the Atlantic Forest. The x axis is the predicted ecological distance and y axis is the observed distance. The height of spline indicates the magnitude of biological change along that gradient and represents the relative importance of each predictor contributing to the turnover component. The shape of the spline indicates where, along each gradient, the changes are most pronounced.

APPENDIX S3

Pairwise indexes matrix and clusters of turnover (β_{sim}), nestedness (β_{sne}) and total beta diversity ($\beta_{s\emptyset r}$) of bats species between Atlantic Forest ecoregions and regions. These matrix and clusters are results from Beta diversity analysis (Betapart/R-project) and the index was calculated from a matrix of presence and absence of species in each ecoregion.

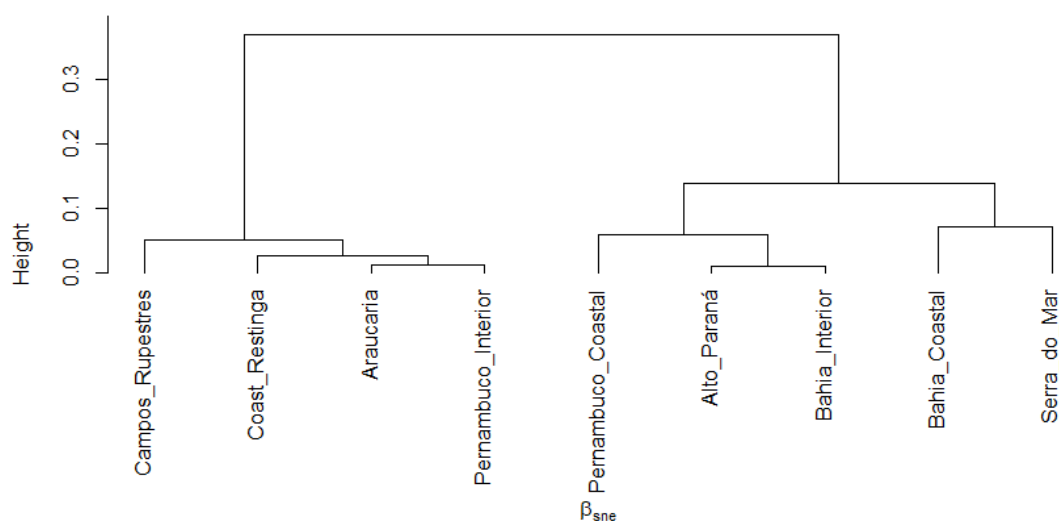


Figure S7.3 Cluster of nestedness (β_{sne}) of bats species between Atlantic Forest ecoregions. Values range from zero (no nestedness between groups) to one (All species are shared between regions). The height varied from 0.01 to 0.55. This cluster is result from Beta diversity analysis (Betapart/R-project) and the index was calculated from a matrix of presence and absence of species in each ecoregion.

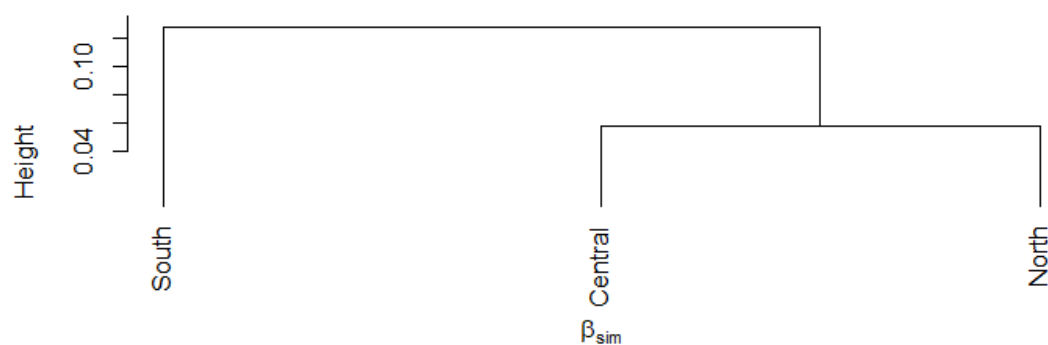


Figure S8.3 Cluster of turnover (β_{sim}) of bats species between Atlantic Forest regions (South, Central and North). South: include the occurrence sites to species in Serra do Mar, Alto Parana, Araucarias and Campos Rupestres ecoregions; Central: include the occurrence sites to species in Bahia Coastal, Bahia Interior and Coast Restinga ecoregions; and the North include the occurrence sites to species in Pernambuco Coastal and Pernambuco Interior

ecoregions. Values range from zero (species composition is the same between both groups) to one (no species are shared between regions). The height varied from 0.04 to 0.10. This cluster is result from Beta diversity analysis (Betapart/R-project) and the index was calculated from a matrix of presence and absence of species in each ecoregion.

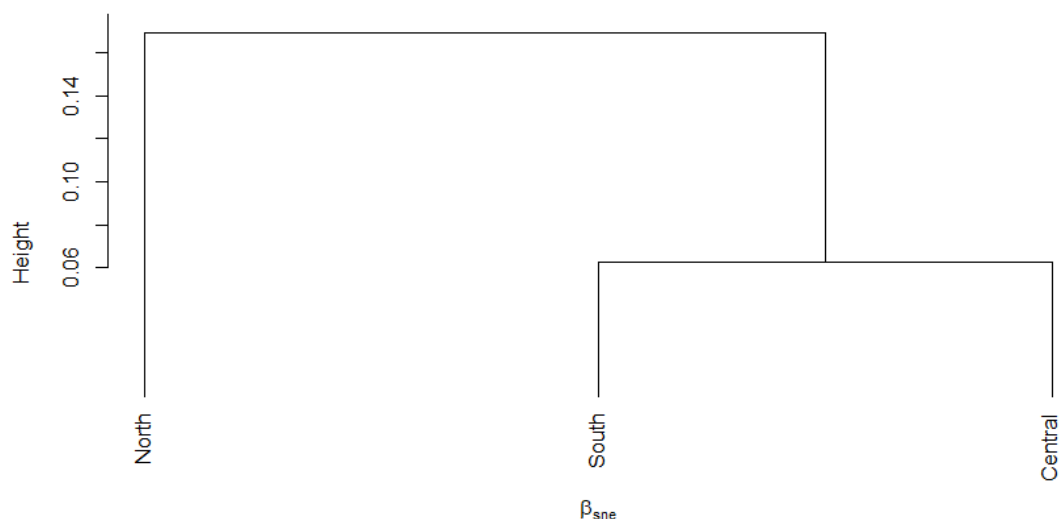


Figure S9.3 Cluster of nestedness (β_{sne}) of bats species between Atlantic Forest regions. South: include the occurrence sites to species in Serra do Mar, Alto Parana, Araucarias and Campos Rupestres ecoregions; Central: include the occurrence sites to species in Bahia Coastal, Bahia Interior and Coast Restinga ecoregions; and the North include the occurrence sites to species in Pernambuco Coastal and Pernambuco Interior ecoregions. Values range from zero (no nestedness between groups) to one (All species are shared between regions). The height varied from 0.06 to 0.20. This cluster is result from Beta diversity analysis (Betapart/R-project) and the index was calculated from a matrix of presence and absence of species in each ecoregion.

Table S2.3 Pairwise indexes matrix of turnover (β_{sim}), nestedness (β_{sne}) and total beta diversity ($\beta_{s\sigma r}$) of bats species between ecoregions in the south of the Atlantic Forest. This matrix is result from Beta diversity analysis (Betapart/R-project) and the index was calculated from a matrix of presence and absence of species in each ecoregion. .

South			
	Alto Paraná	Araucaria	Campos Rupestres
Tunover (β_{sim}) - 0.07042254			

Araucaria	0.00000000		
Campos Rupestres	0.00000000	0.13333333	
Serra do Mar	0.08108108	0.00000000	0.00000000
Nestedness (β_{sne}) - 0.4057679			
Araucaria	0.37037037		
Campos Rupestres	0.42307692	0.05416667	
Serra do Mar	0.08965063	0.45161290	0.50000000
Beta diversity (β_{sOr}) - 0.4761905			
Araucaria	0.3703704		
Campos Rupestres	0.4230769	0.1875000	
Serra do Mar	0.1707317	0.4516129	0.5000000

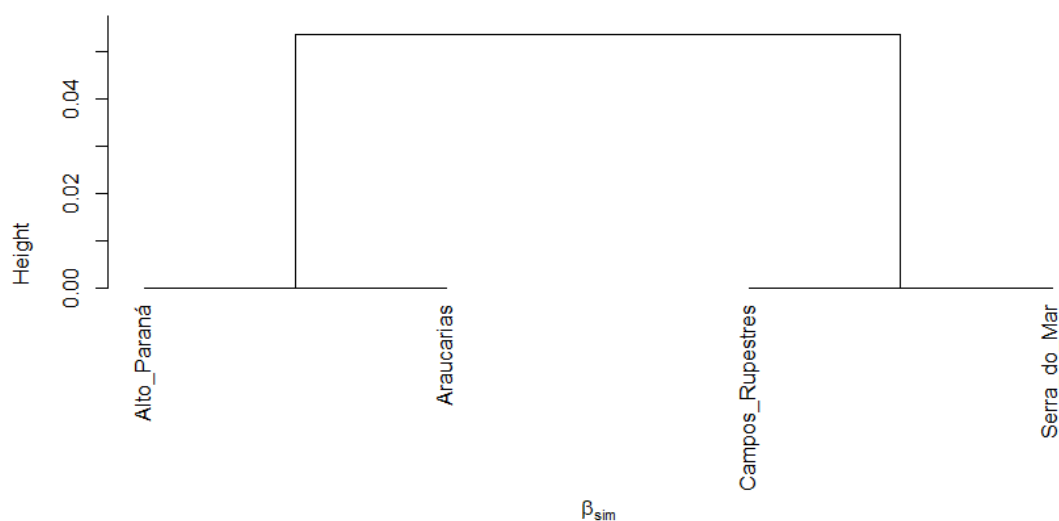


Figure S10.3 Cluster of turnover (β_{sim}) of bats species between the ecoregions located in the south of the Atlantic Forest. Values range from zero (when species composition is the same between both groups) to one (no species are shared between regions). The height varied from 0.00 to 0.05. This cluster is result from Beta diversity analysis (Betapart/R-project) and the index was calculated from a matrix of presence and absence of species in each ecoregion. .

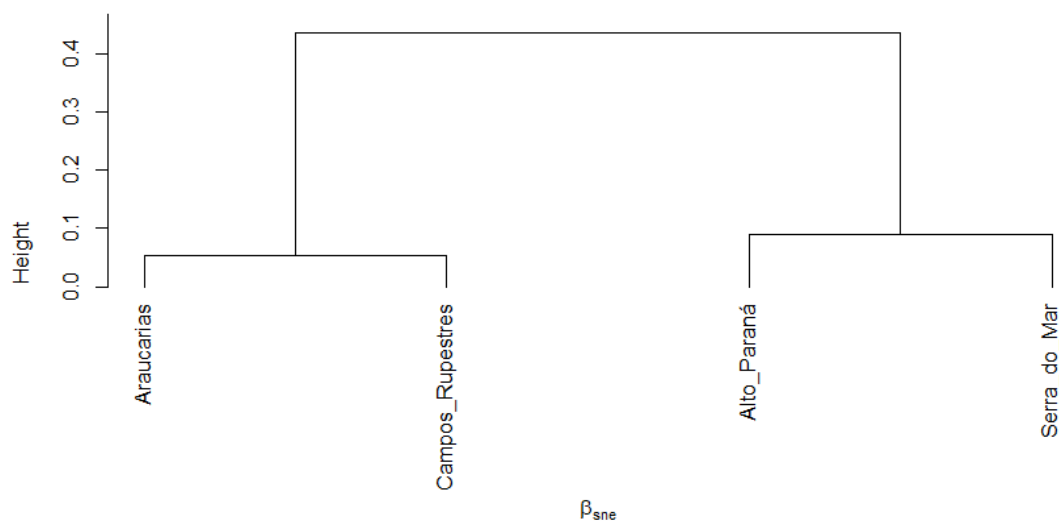


Figure S11.3 Cluster of nestedness (β_{sne}) of bats species between the ecoregions located in the south of the Atlantic Forest. Values range from zero (no nestedness between groups) to one (All species are shared between regions). The height varied from 0.05 to 0.45. This cluster is result from Beta diversity analysis (Betapart/R-project) and the index was calculated from a matrix of presence and absence of species in each ecoregion.

Table S3.3 Pairwise indexes matrix of turnover (β_{sim}), nestedness (β_{sne}) and total beta diversity (β_{sOr}) of bats species between ecoregions in the centre of the Atlantic Forest. This matrix is result from Beta diversity analysis (Betapart/R-project) and the index was calculated from a matrix of presence and absence of species in each ecoregion.

Centre		
	Coast Restinga	Bahia Coastal
Tunover (β_{sim}) - 0.05172414		
Bahia Coastal	0.00000000	
Bahia Interior	0.05263158	0.05263158
Nestedness (β_{sne}) - 0.3504498		
Bahia Coastal	0.4722222	
Bahia Interior	0.3157895	0.1561596
Beta diversity (β_{sOr}) - 0.4021739		
Bahia Coastal	0.4722222	
Bahia Interior	0.3684211	0.2087912

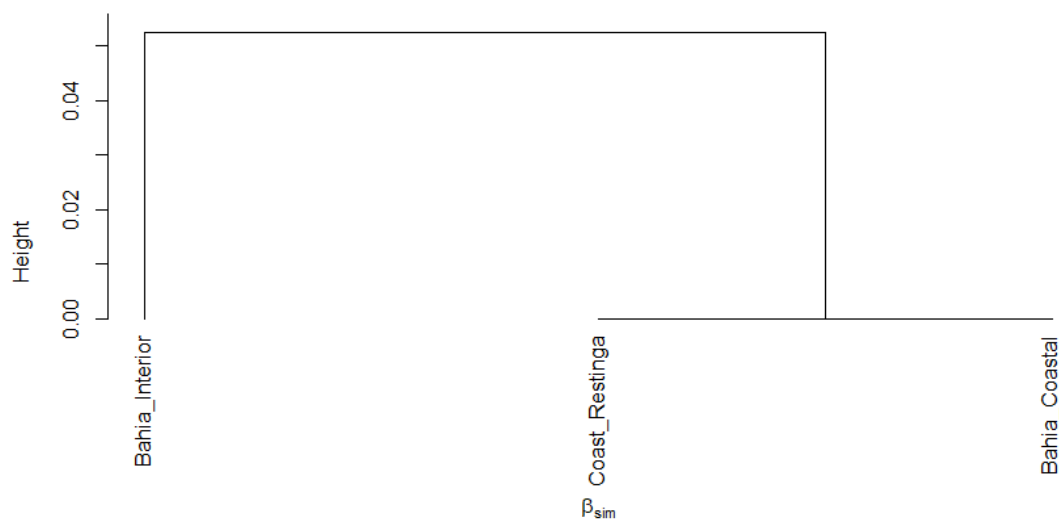


Figure S12.3 Cluster of turnover (β_{sim}) of bats species between the ecoregions located in the center of the Atlantic Forest. Values range from zero (when species composition is the same between both groups) to one (no species are shared between regions). The height varied from 0.00 to 0.05. This cluster is result from Beta diversity analysis (Betapart/R-project) and the index was calculated from a matrix of presence and absence of species in each ecoregion.

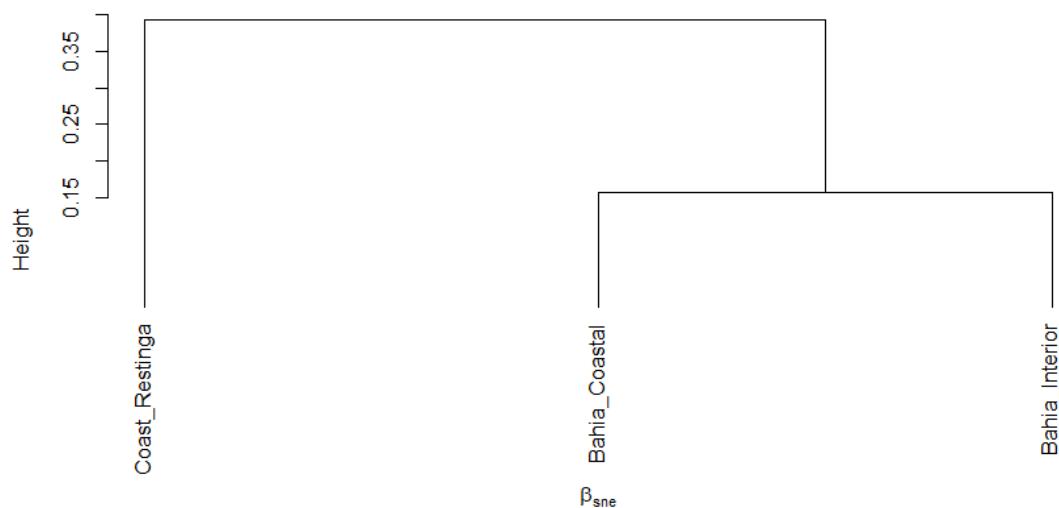


Figure S13.3 Cluster of nestedness (β_{sne}) of bats species between the ecoregions located in the center of the Atlantic Forest. Values range from zero (no nestedness between groups) to one (All species are shared between regions). The height varied from 0.15 to 0.47. This cluster is result from Beta diversity analysis (Betapart/R-project) and the index was calculated from a matrix of presence and absence of species in each ecoregion.

Table S4.3 Pairwise indexes matrix of turnover (β_{sim}), nestedness (β_{sne}) and total beta diversity ($\beta_{s\text{Ø}r}$) of bats species between ecoregions in the north of the Atlantic Forest. This matrix is result from Beta diversity analysis (Betapart/R-project) and the index was calculated from a matrix of presence and absence of species in each ecoregion.

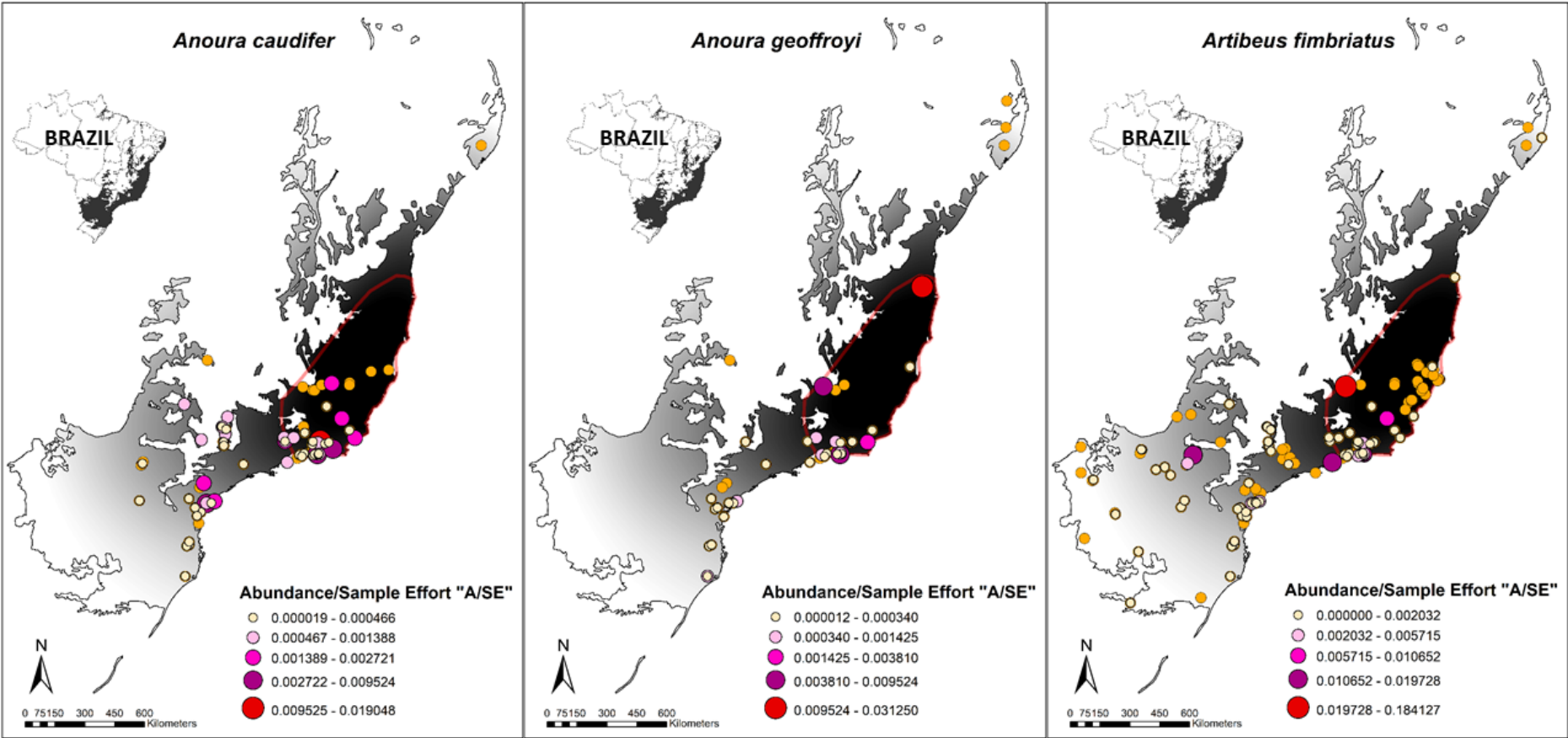
North	
	Pernambuco Coastal
Tunover (β_{sim}) - 0.1666667	
Pernambuco Interior	0.1666667
Nestedness (β_{sne}) - 0.2333333	
Pernambuco Interior	0.2333333
Beta diversity ($\beta_{s\text{Ø}r}$) - 0.4	
Pernambuco Interior	0.4

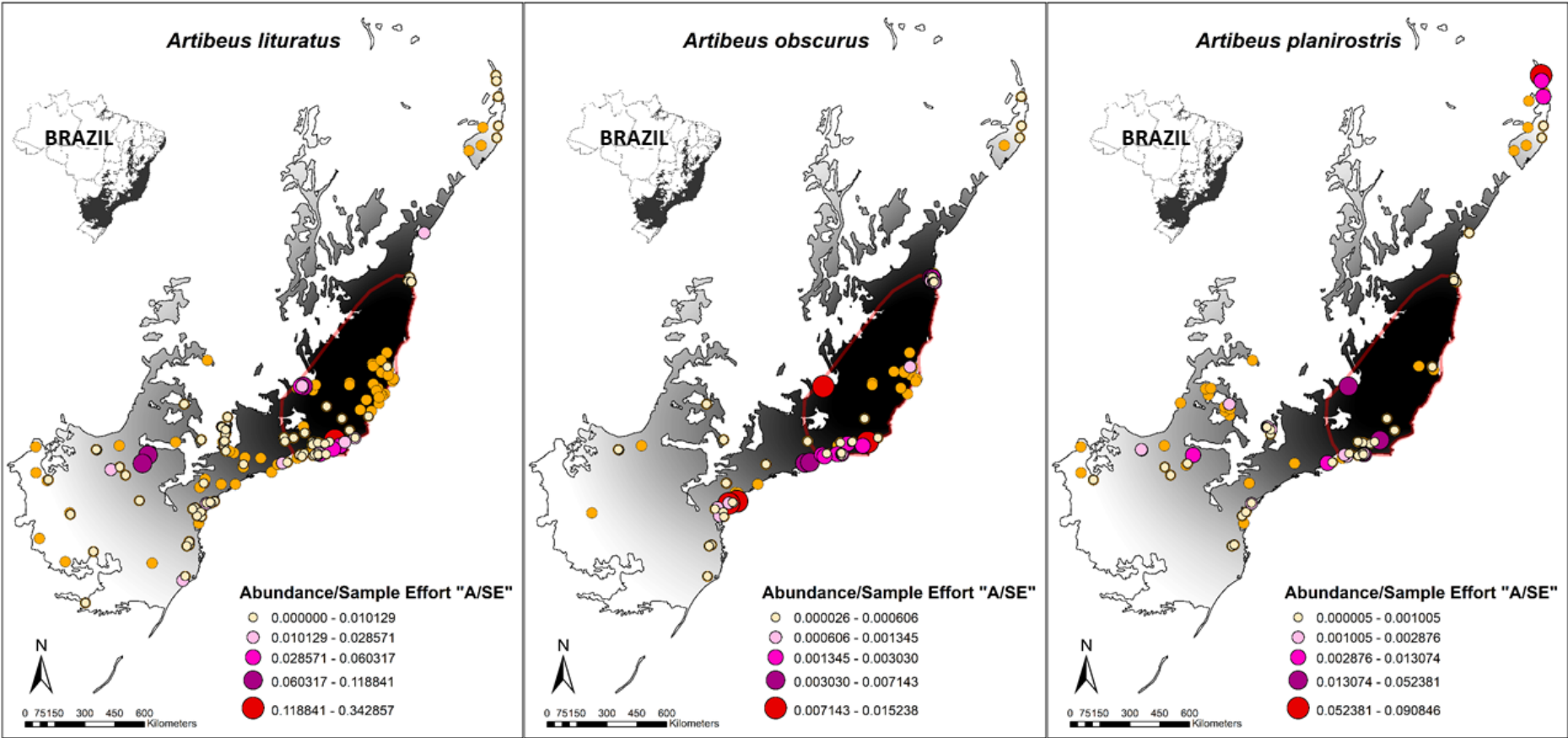
APÊNDICE C

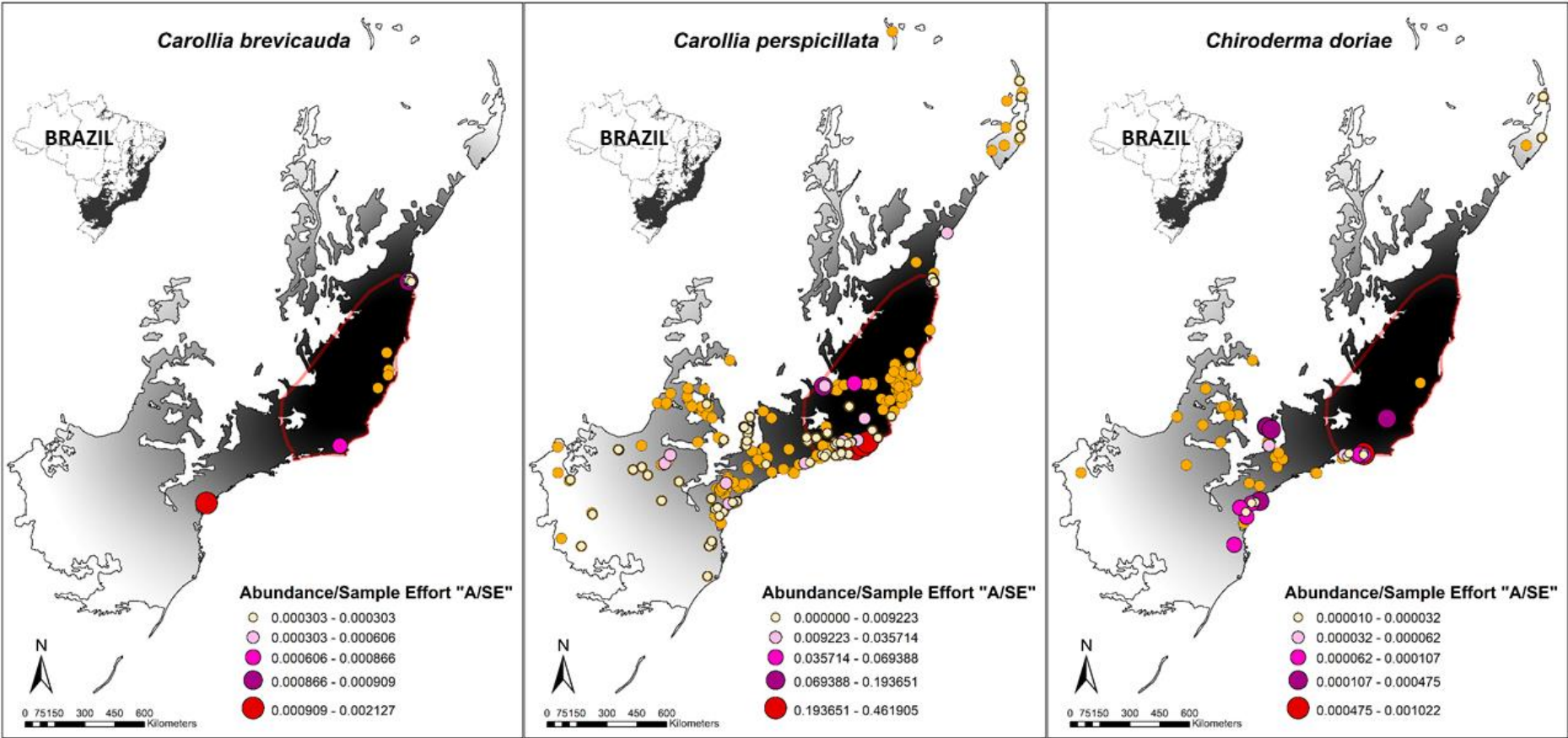
Material suplementar do Capítulo 3. O material encontra-se nas normas do periódico "Biological Conservation".

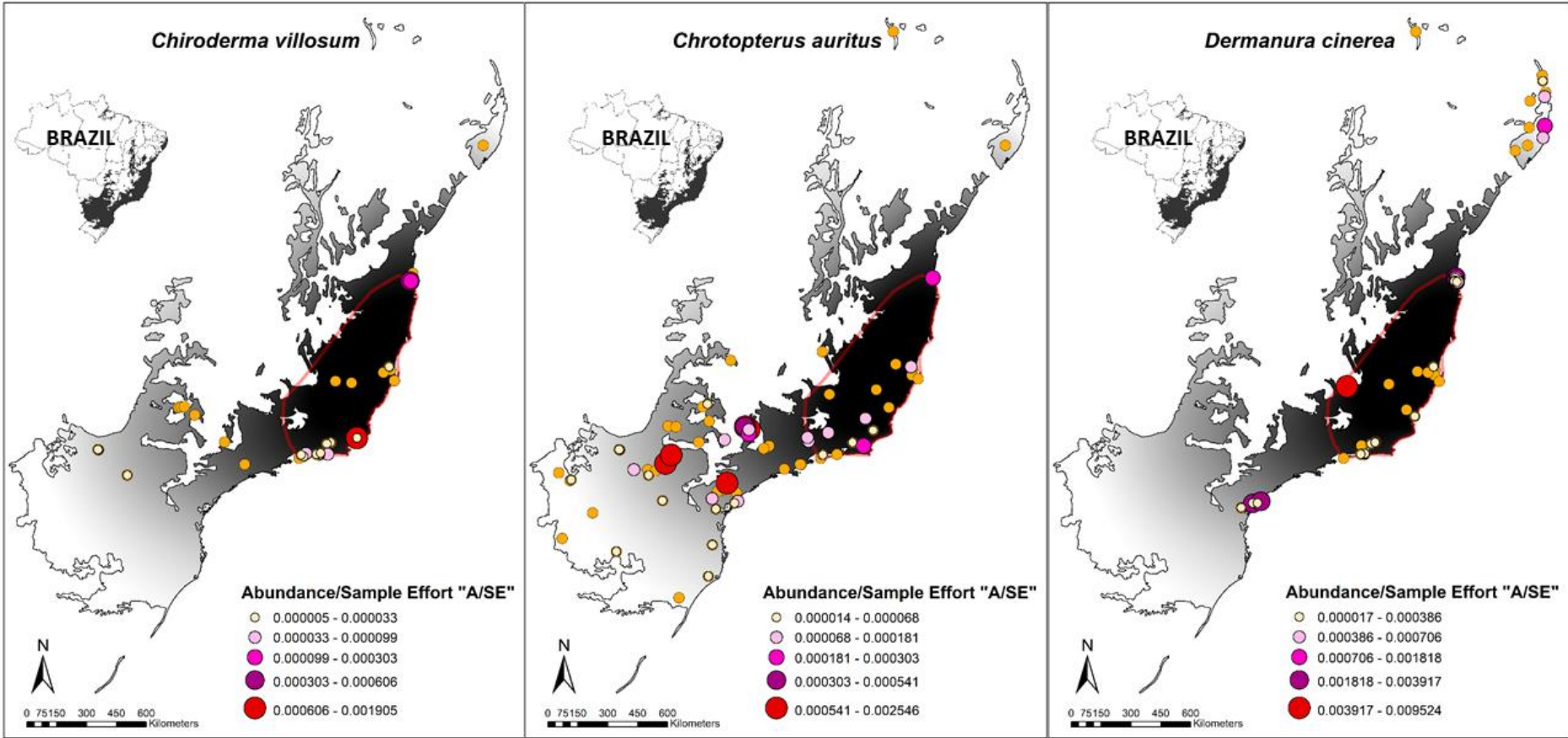
1 Distribution maps of the 59 bat species in the Atlantic Forest included in our study. The orange dots represent georeferenced occurrence sites
2 without sampling effort information, therefore without the Abundance / Effort Sample (“A / SE”) index. The red boundary represents the limits
3 of the congruence area for 58 of the 59 species. The 58 species that had the “A / SE” index calculated have points of colors and sizes that vary
4 according to the index, and are in the caption of each map.

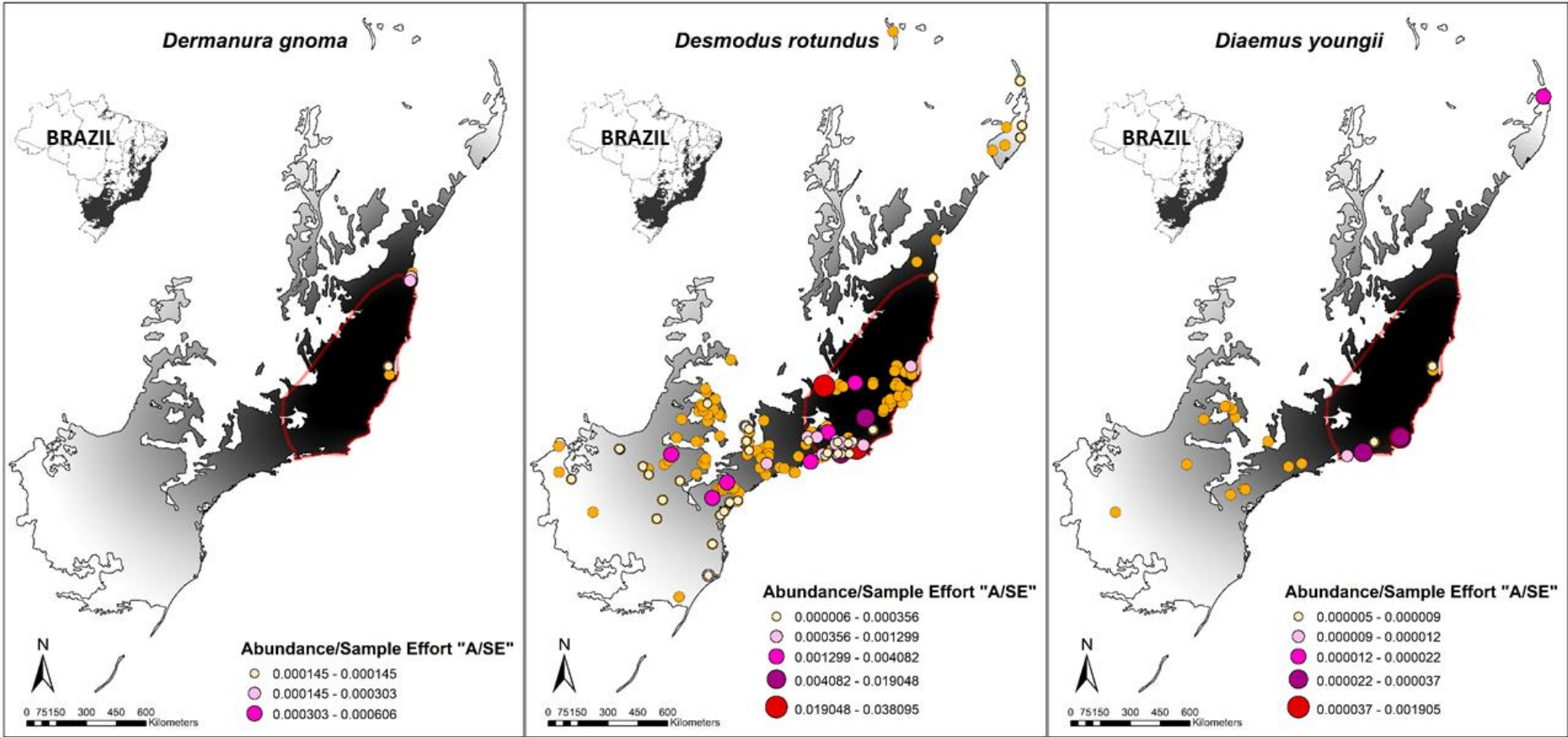
5

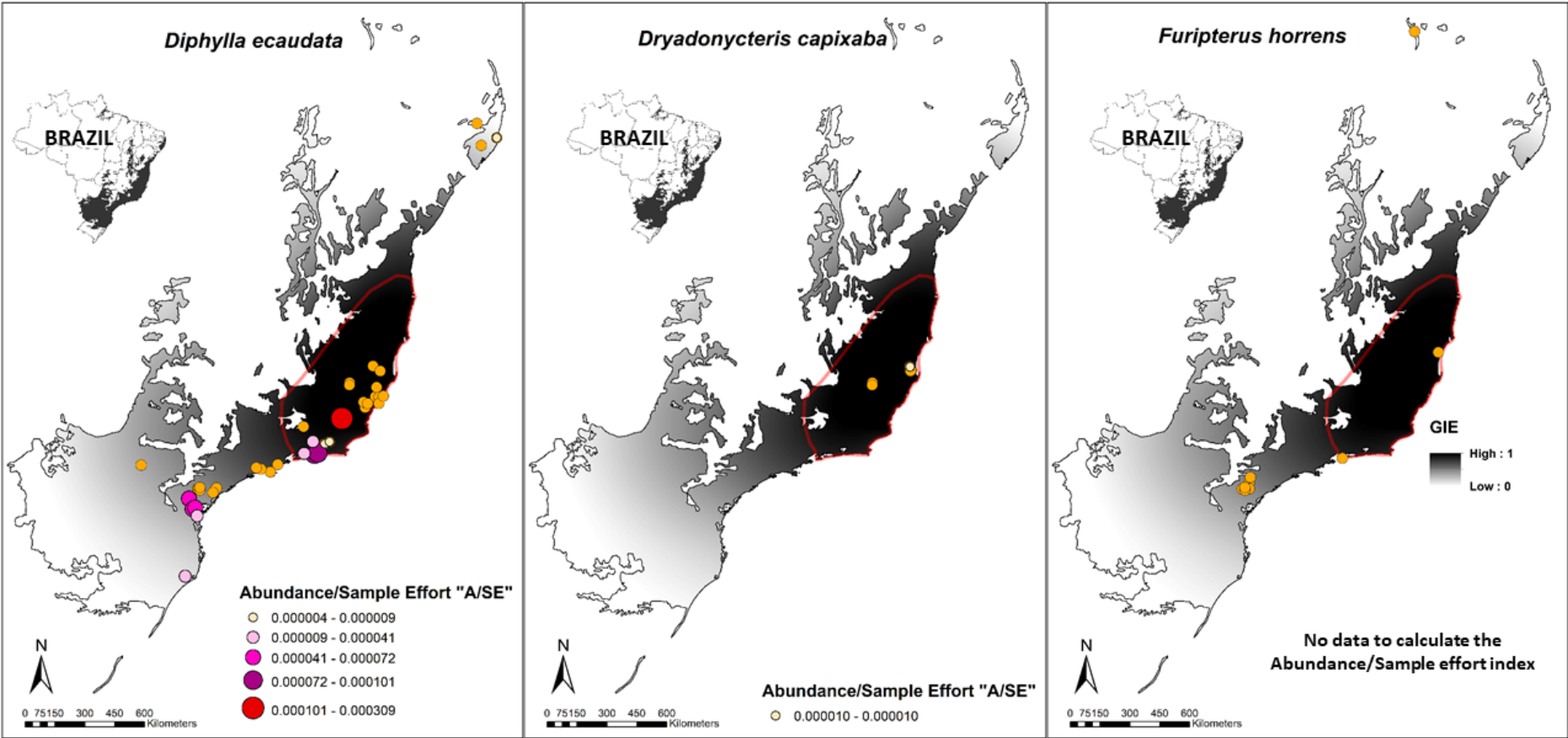


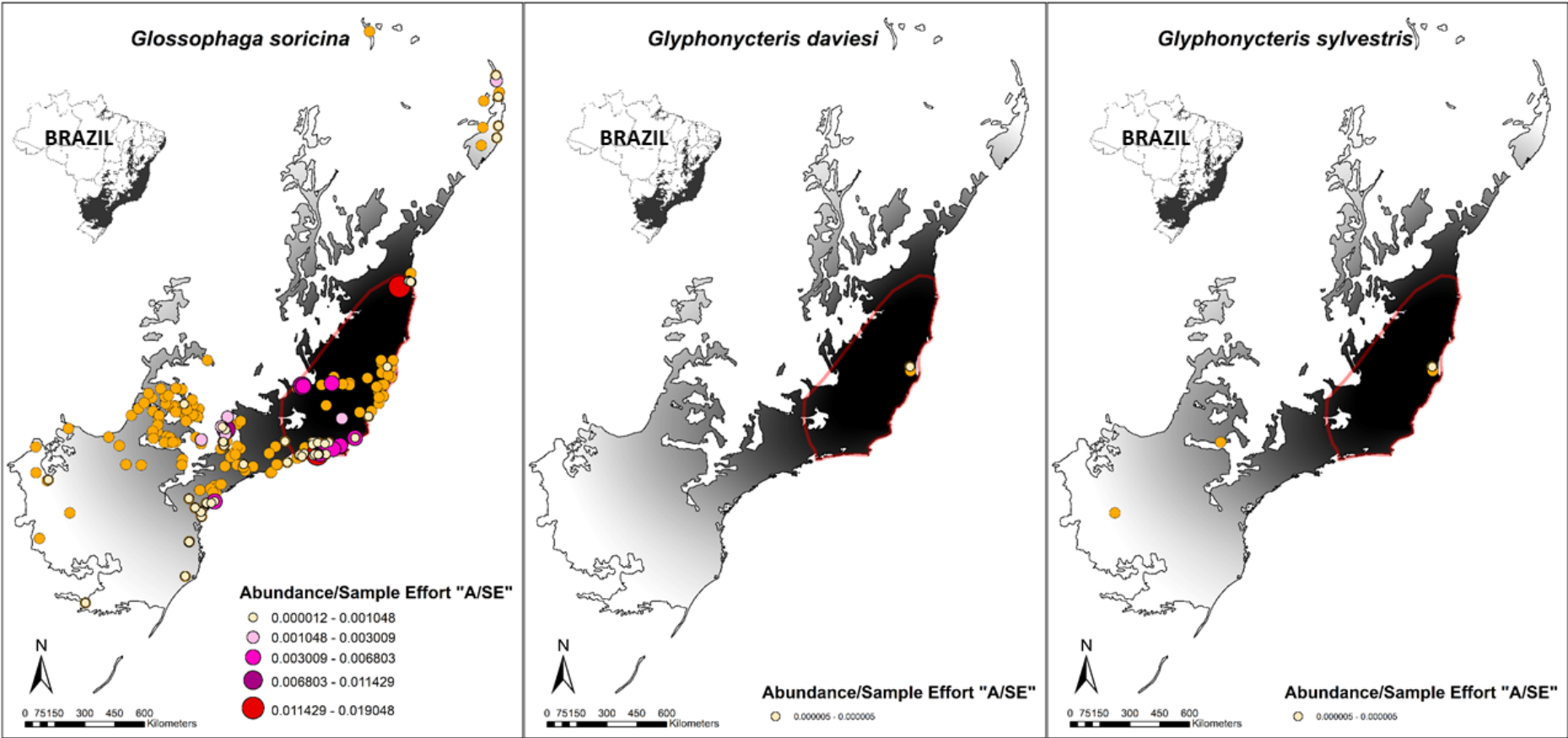


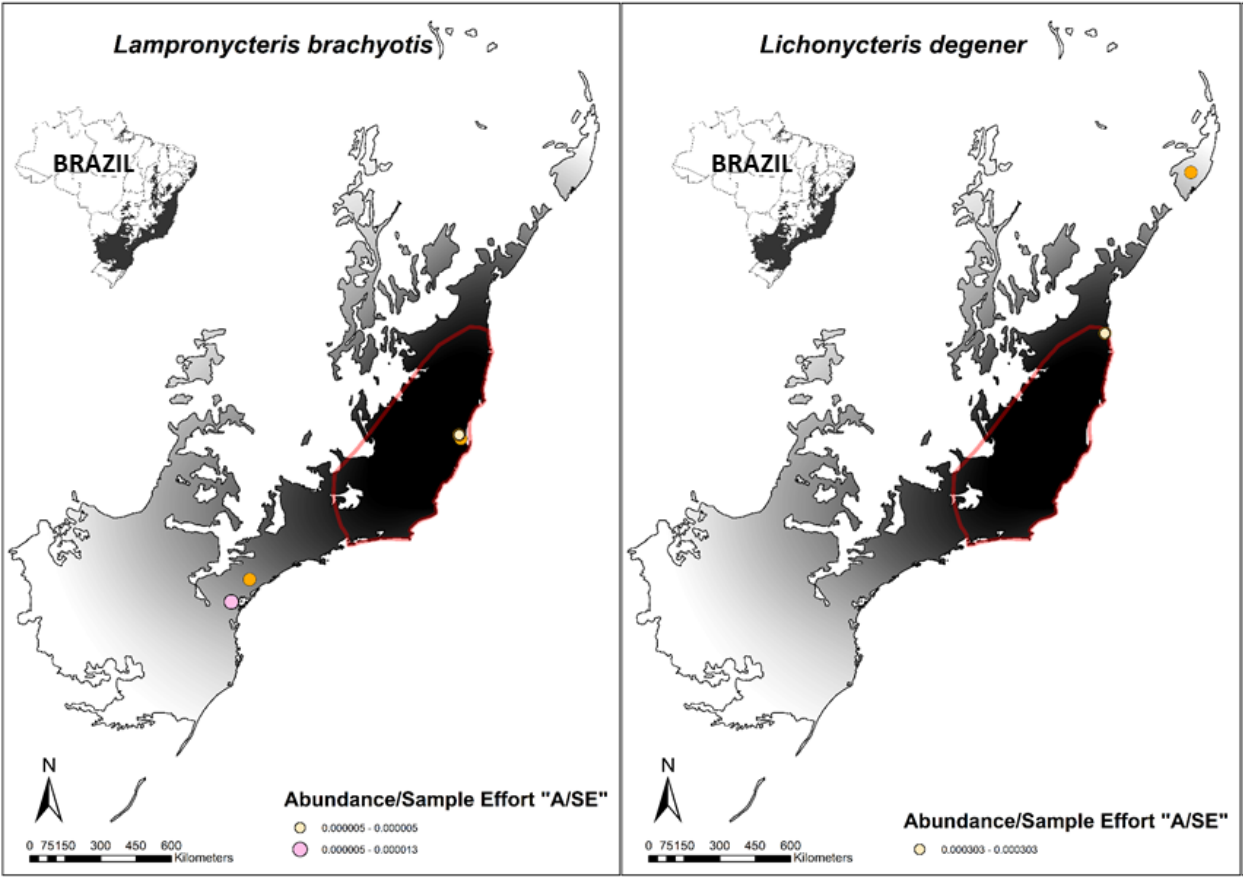


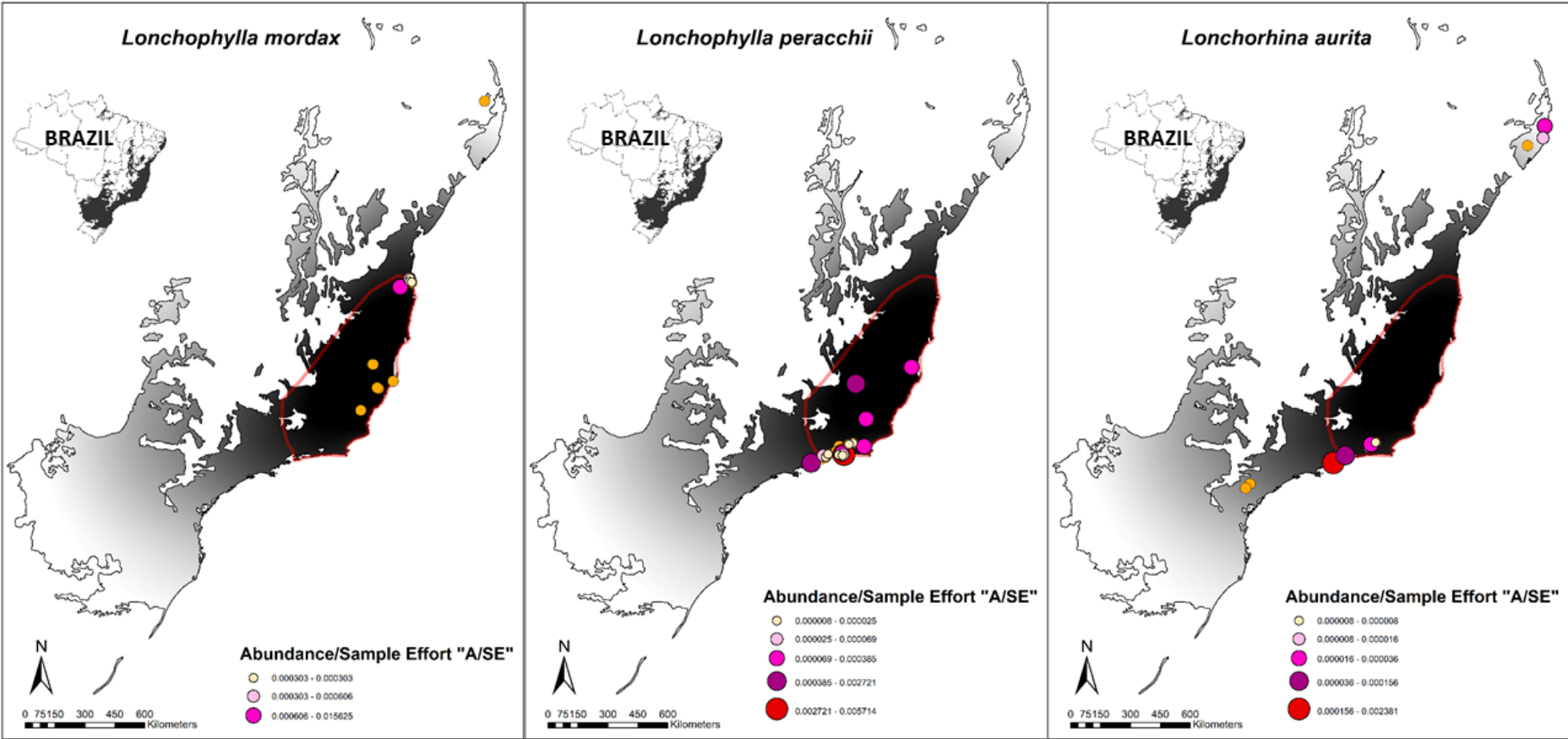


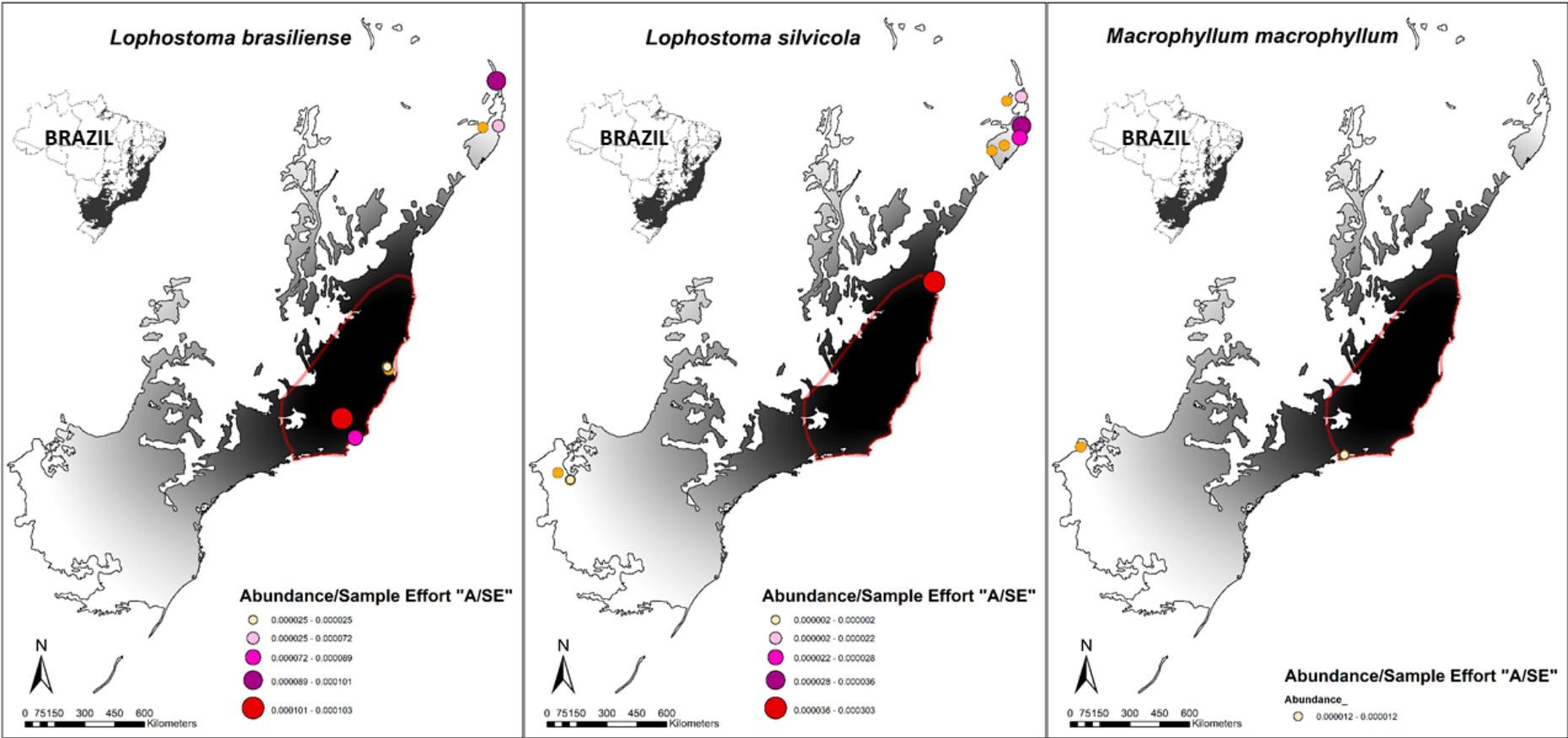


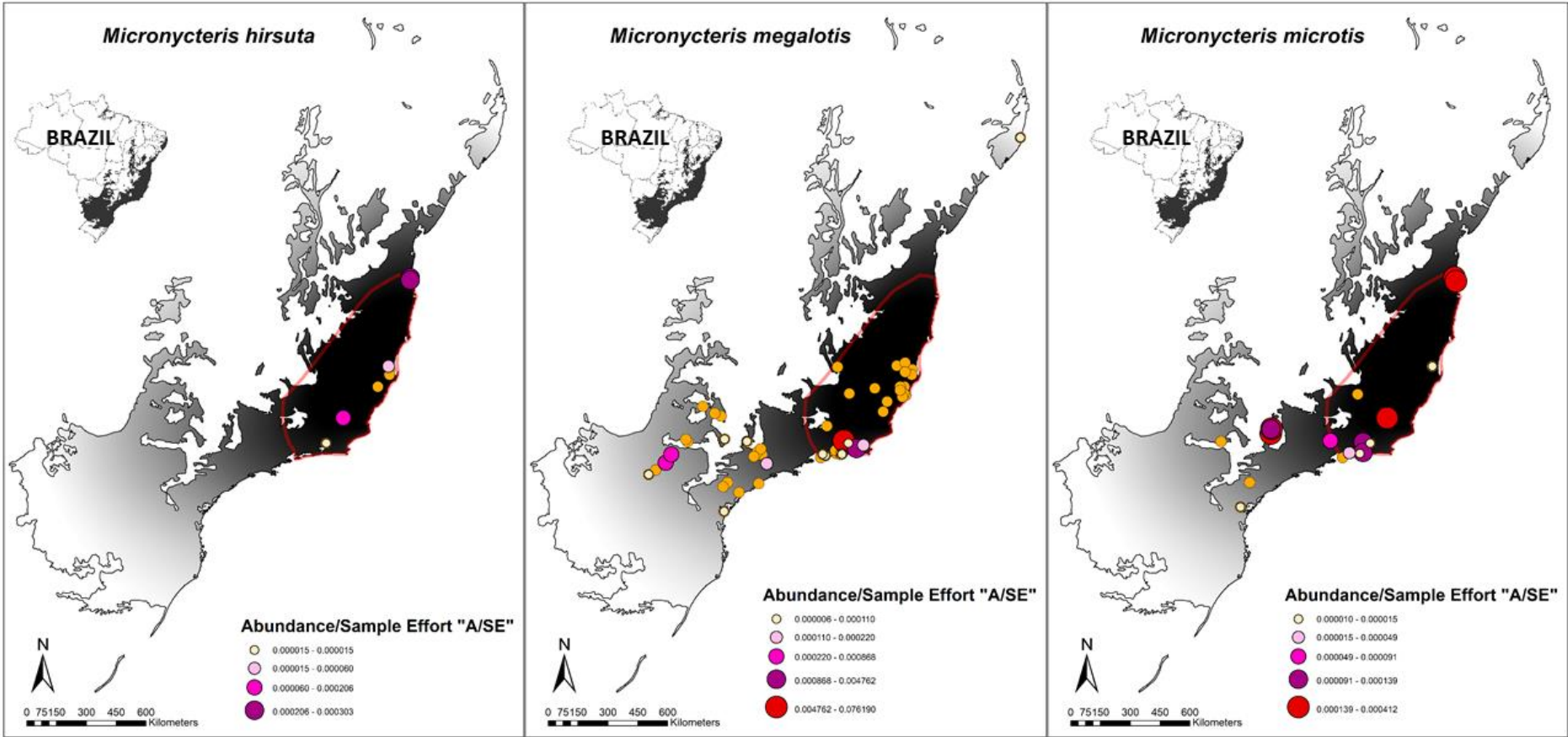


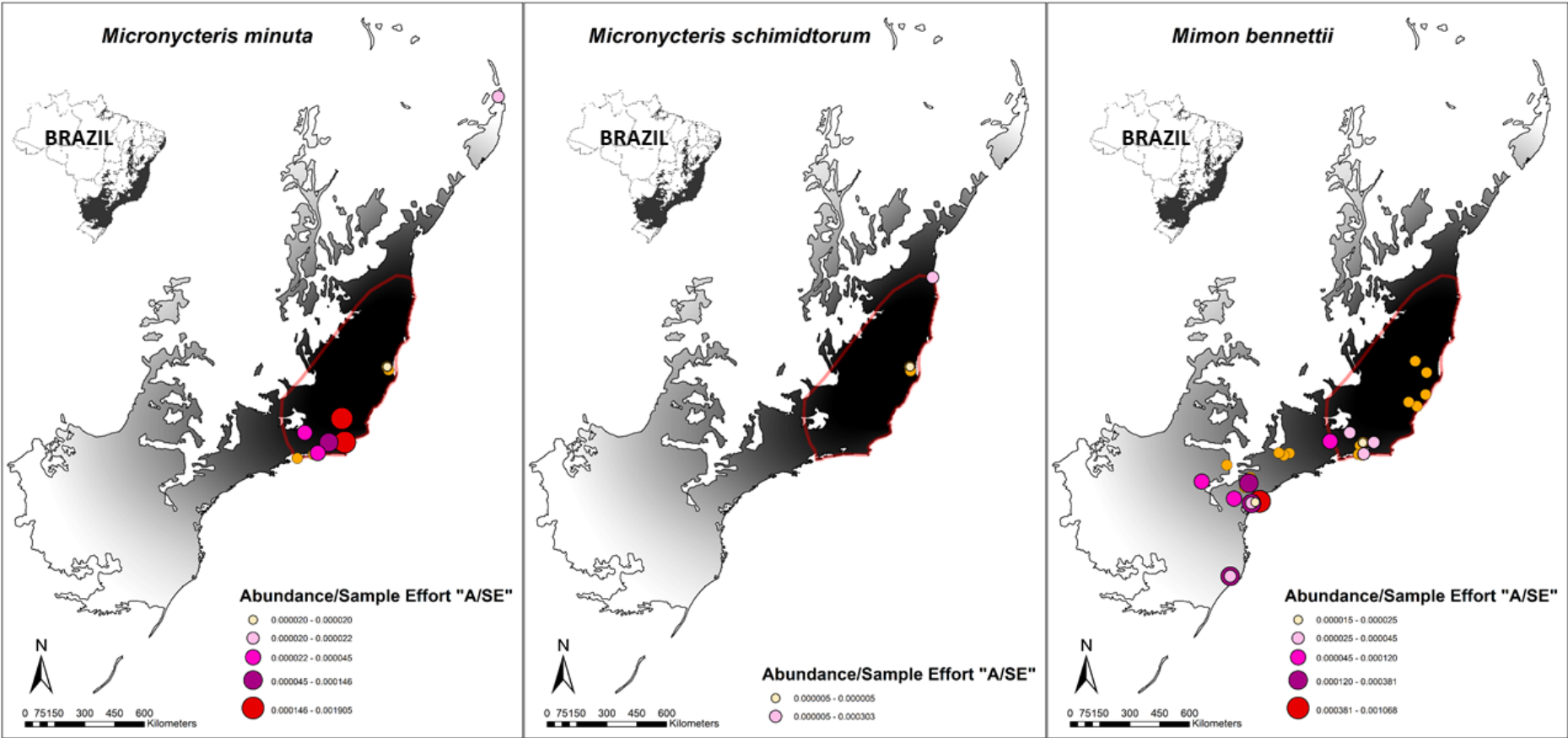


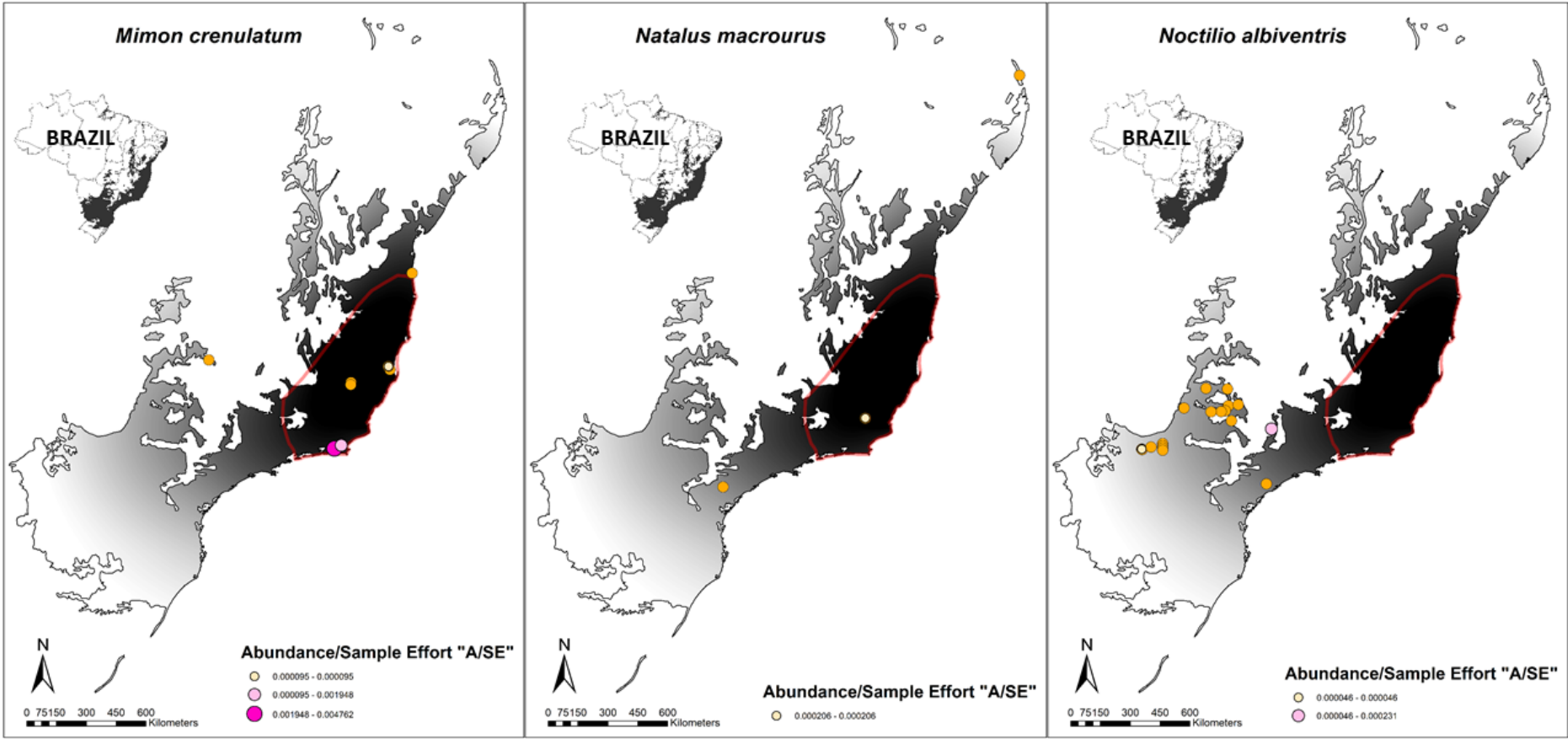


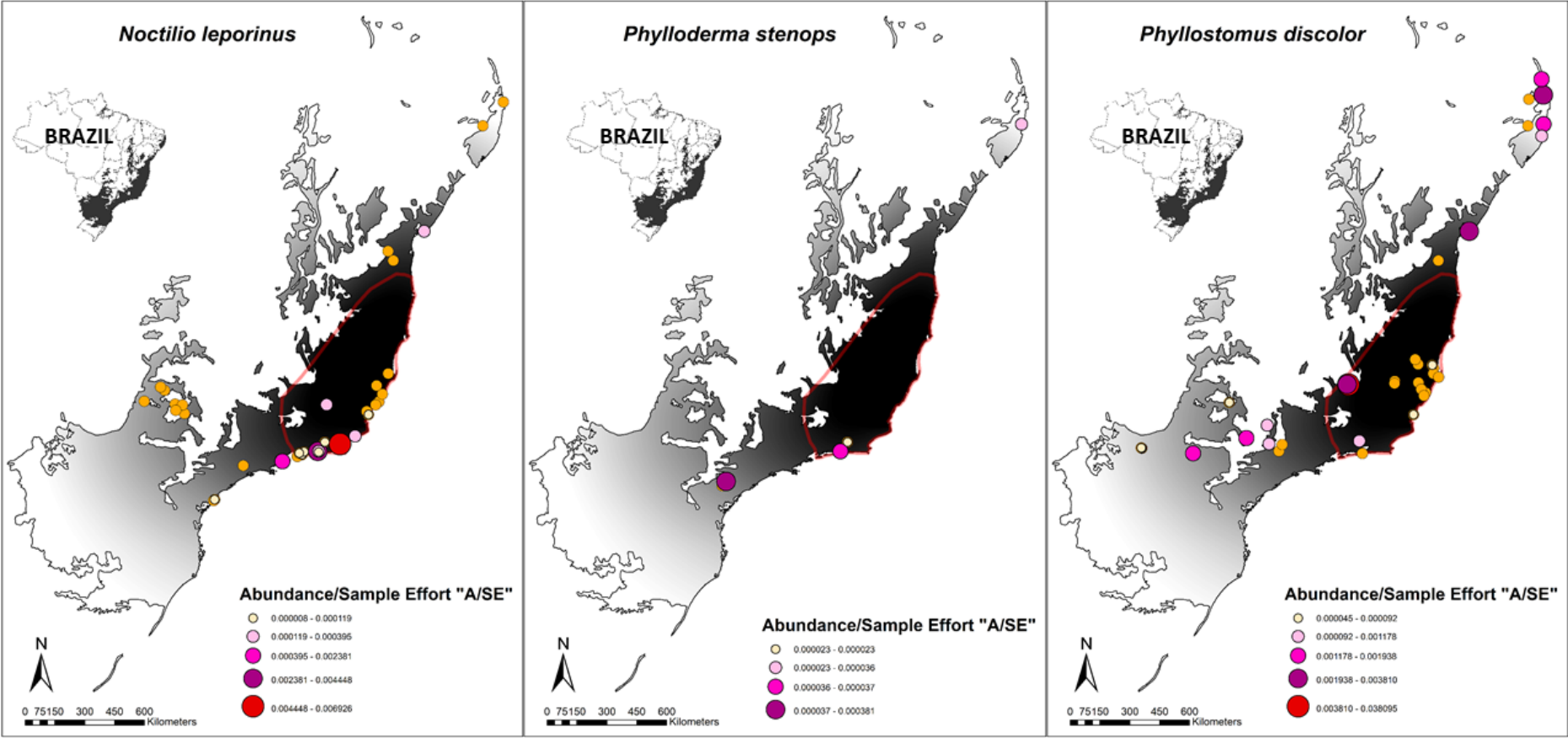


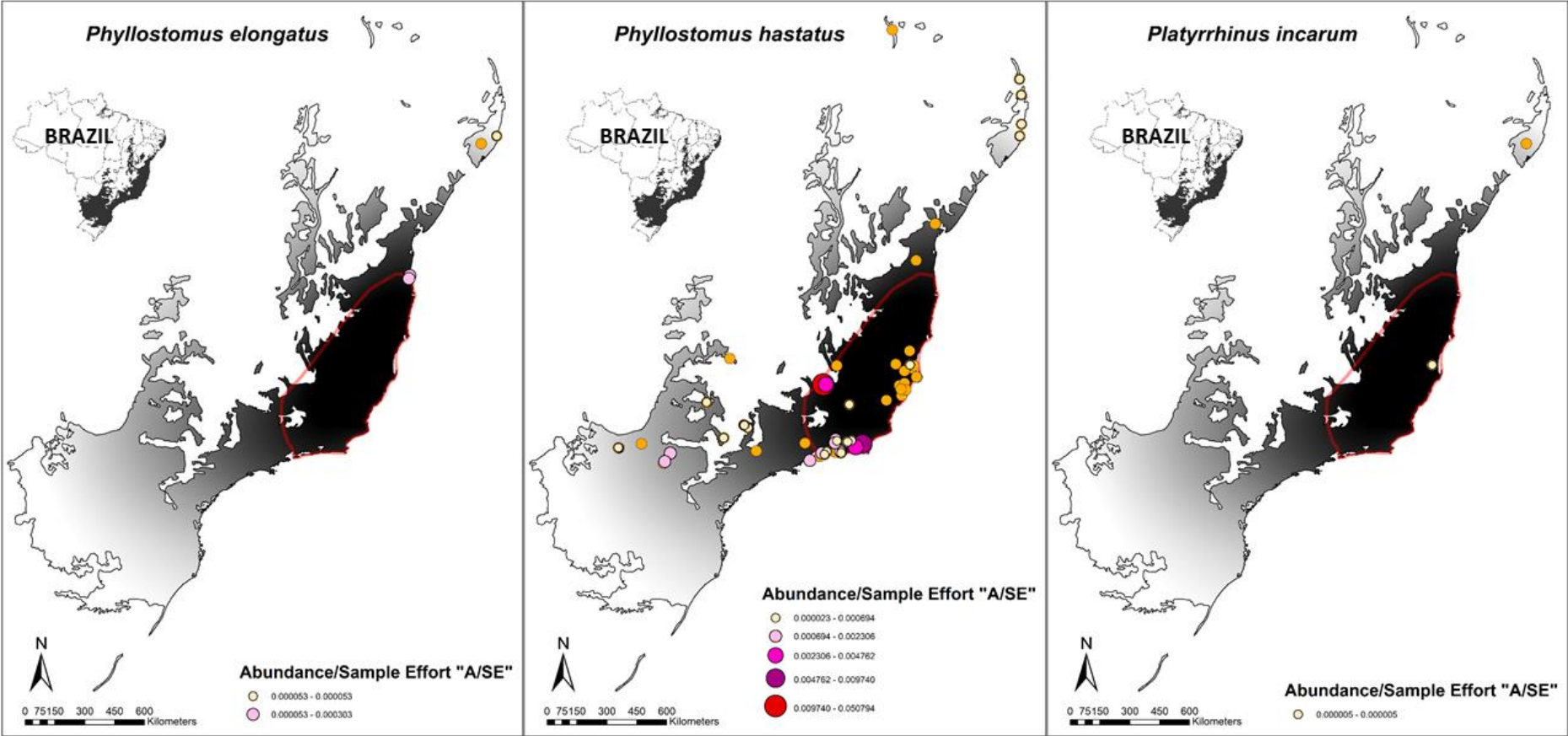


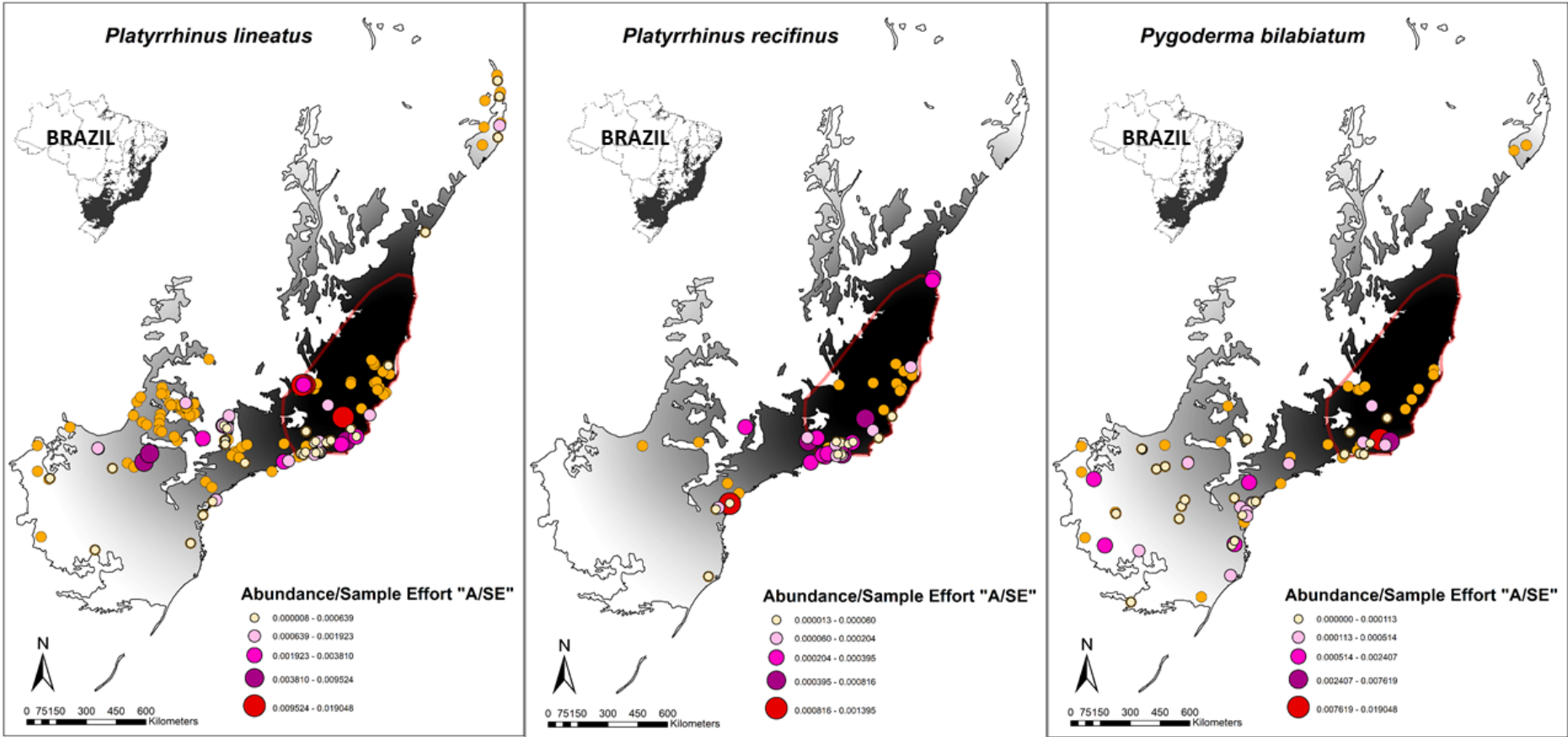


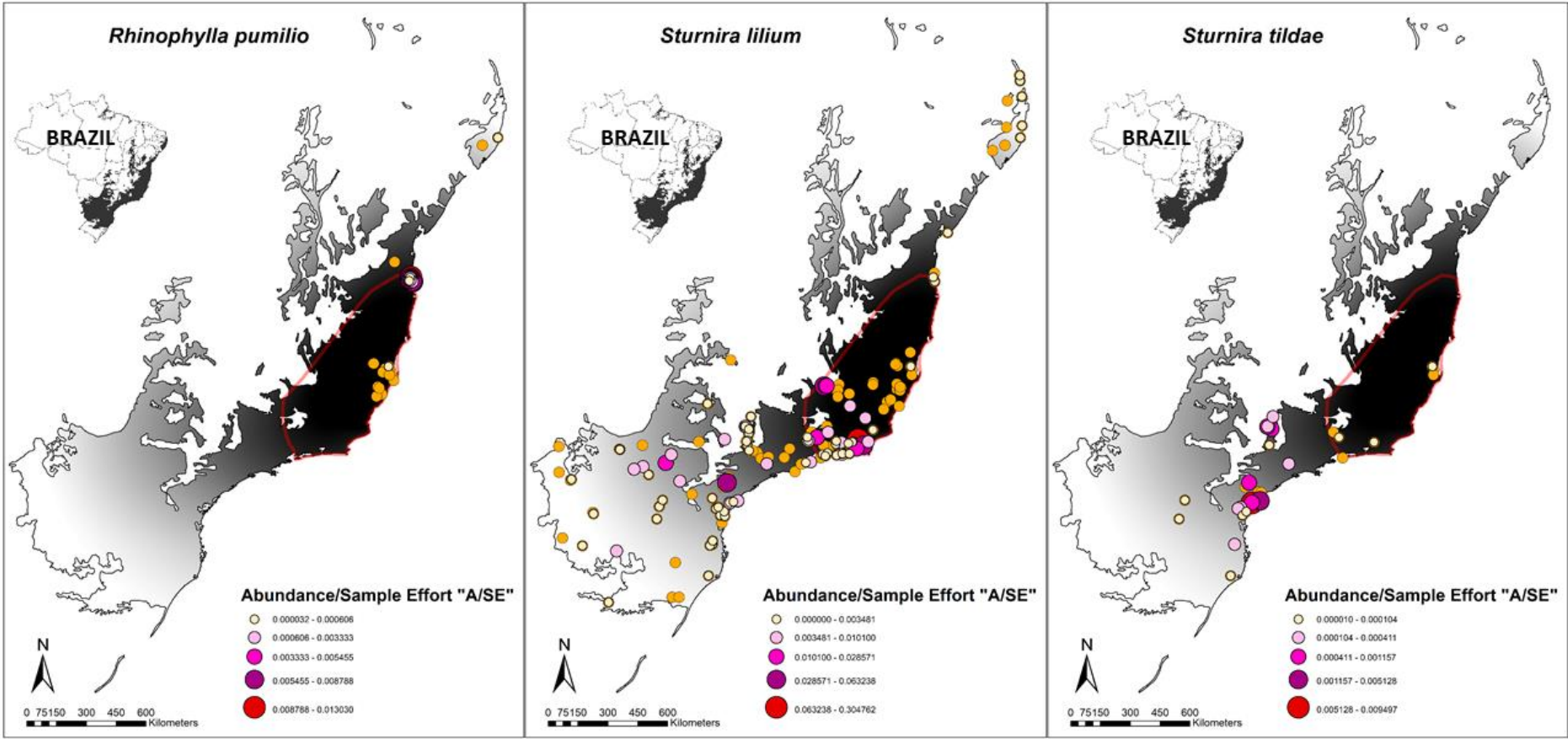


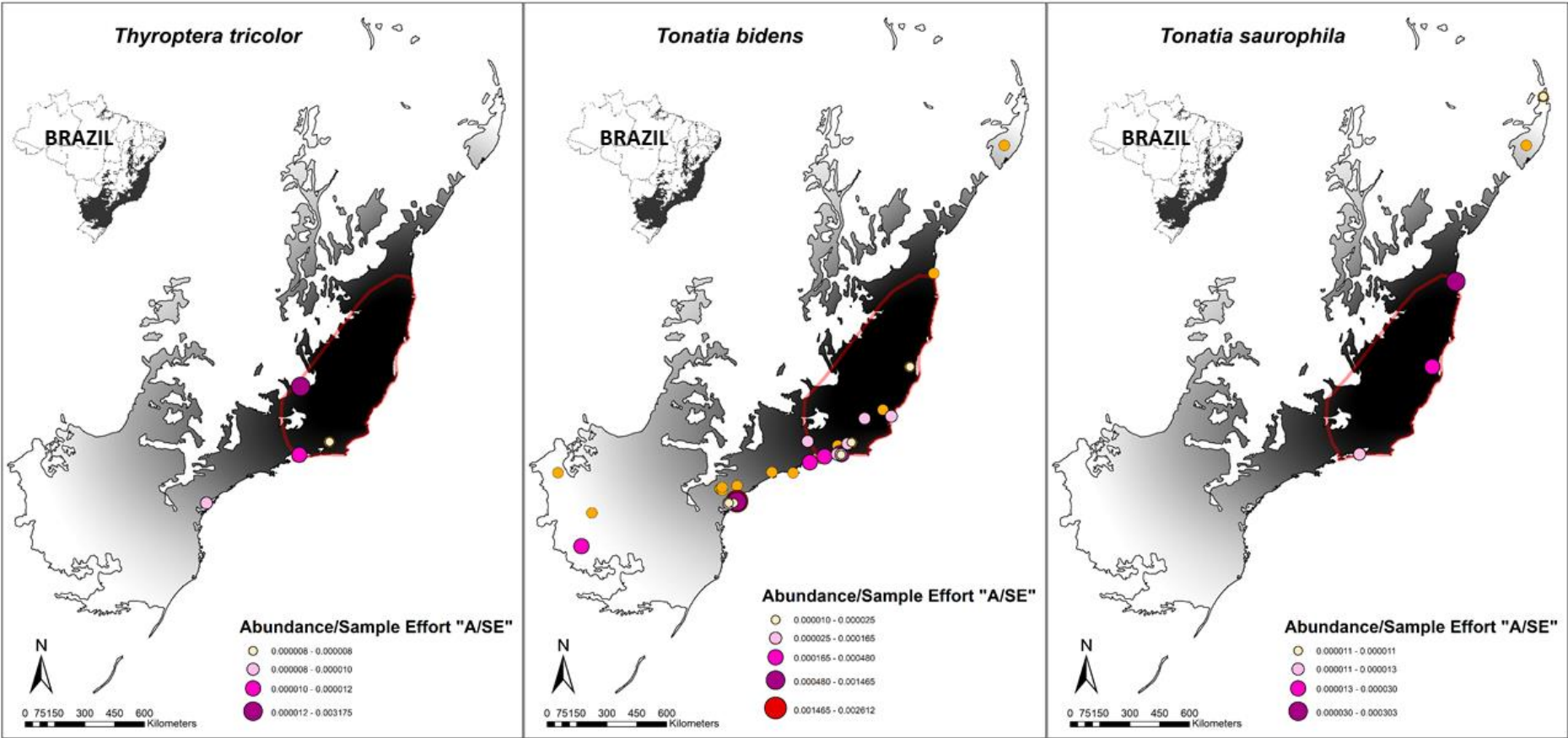


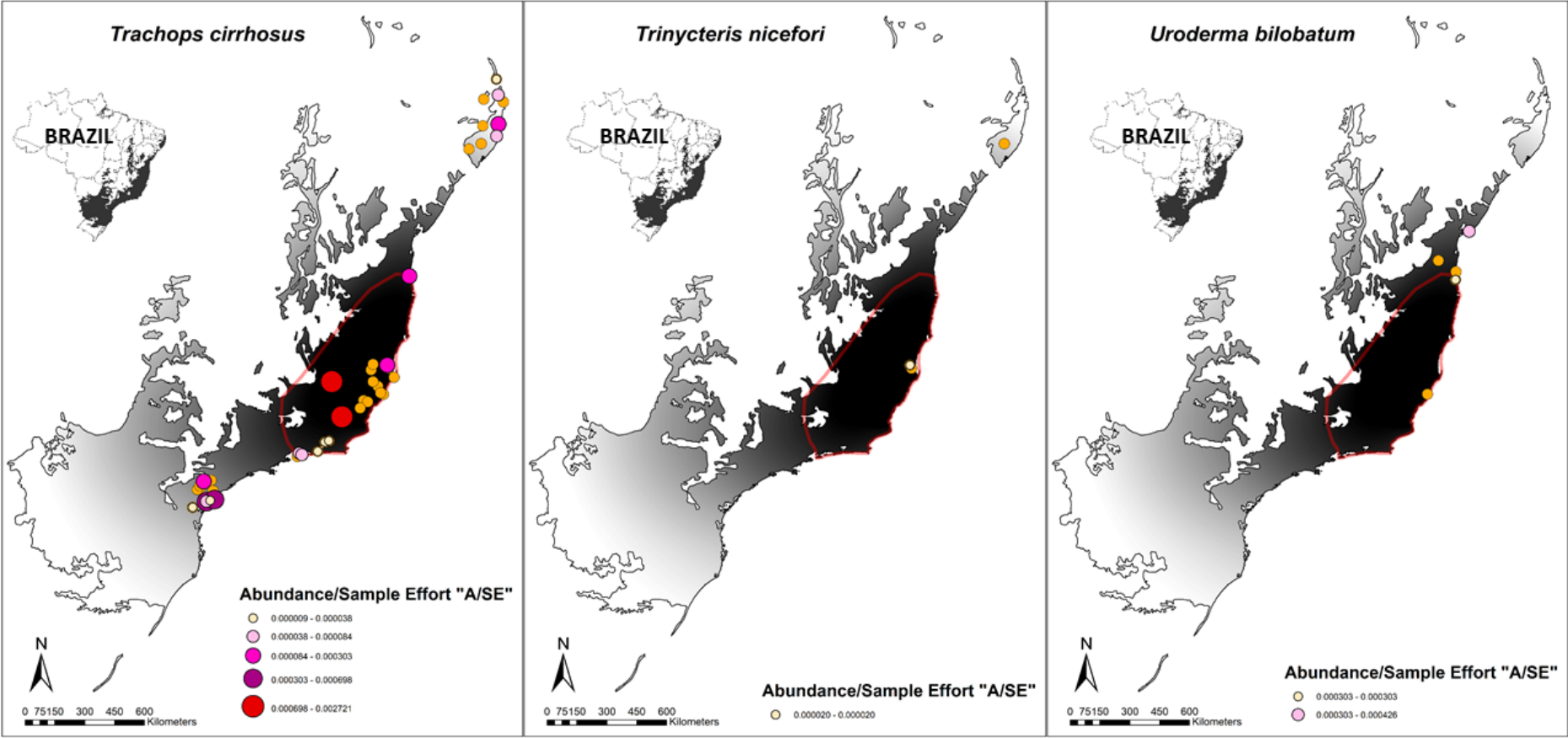


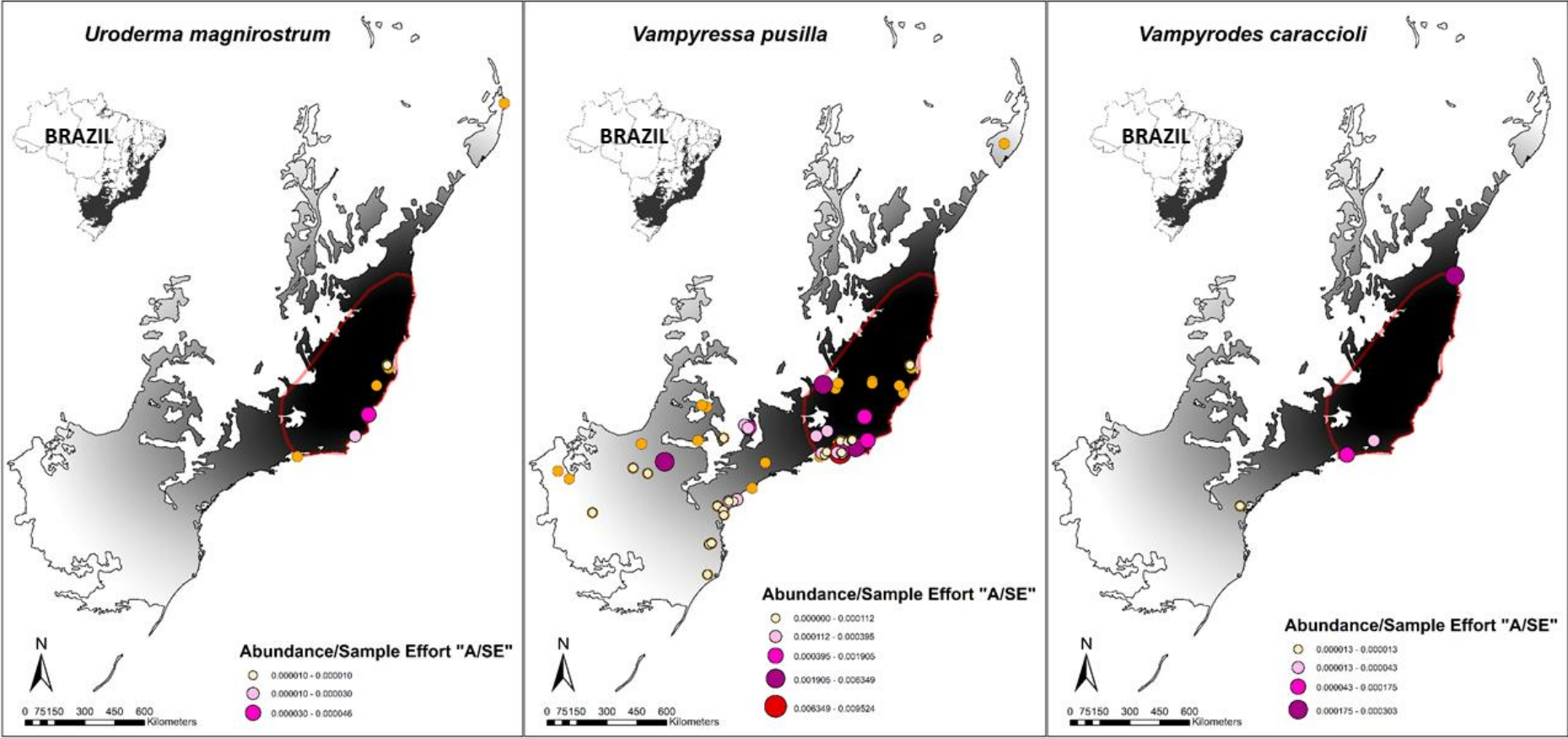












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Table 1A. List of 59 bat species and occurrences (georeferenced locations) for all locations in the Atlantic Forest, Congruence Area and only for Protected Areas (PAs). The occurrence numbers represent the total number of georeferenced sites where each species was recorded. The table also shows other information for the species: food preference, occurrence biomes (Reis et al., 2017), conservation status (Global assessment - IUCN, 2016; National assessment- ICMBio/MMA, 2018), class of the Geographical Interpolation Endemism (GIE) and Status in the Area of Congruence. Food Preference: O / C - Omnivore / Carnivore; I - insectivore; H - hematophagous; O - Omnivorous; N - nectarivore; F - Frugivore; P - Piscivore. Occurrence in other biomes: A - Amazon; C - Cerrado; Ca - Caatinga; P - Pantanal; Pa - Pampas. Conservation status: NI – Not included; LC - Least Concern; DD - Data Deficient; NT - Near Threatened and EN - Endangered. Class GIE: class 1- radius of distribution of the species varies from 0 to 400 km; Class 2- 401 to 800 km; Class 3- from 801 to 1,200 km; Class 4- from 1,201 to 1,600 km; Class 5- from 1,601 to 2,000 km; and Class 6- and from 2,001 to 2,200 km. Status in the Congruence Area: P - present; A - absent; R - restricted.

Species	Occurrence in other biomes	Food preference	Conservation status (IUCN)	Conservation status (ICMBio/MMA)	Total Occurrences	Occurrences in Congruence area	Total of occurrences in PAs	Total of occurrences in Pas in Congruence area	Occurrences in PtAs	Occurrences in PtAs in Cogruence area	Abundance/Sample Effort index "A/EA"	Class (GIE)	Status in Congruence area
Family Phyllostomidae													
Subfamily Micronycterinae													
<i>Lamproncyteris brachyotis</i> (Dobson, 1879)	C and A	O/C	LC	NI	4	2	1	1	0	0	C	3	P
<i>Micronycteris hirsuta</i> (Peters, 1869)	A	I	LC	NI	8	8	3	3	0	0	C	2	R
<i>Micronycteris megalotis</i> (Gray, 1842)	A, C, Ca and P	I	LC	NI	58	30	11	4	6	3	C	7	P
<i>Micronycteris microtis</i> Miller, 1898	A and C	I	LC	NI	19	11	7	4	1	0	C	4	P

<i>Micronycteris minuta</i> (Gervais, 1856)	A, C, Ca and P	I	LC	NI	11	10	5	4	1	1	C	6	P
<i>Micronycteris schmidtorum</i> Sanborn, 1935	A, C and Ca	I	LC	NI	3	2	1	1	0	0	C	2	R
Subfamily Desmodontinae													
<i>Desmodus rotundus</i> (É. Geoffroy, 1810)	A, Ca, C, P and Pa	H	LC	NI	203	80	51	11	0	9	C	7	P
<i>Diaemus youngii</i> (Jentink, 1893)	A, Ca, C and P	H	LC	NI	22	7	7	3	3	0	C	7	P
<i>Diphylla ecaudata</i> Spix, 1823	A, Ca, C and P	H	LC	NI	46	26	10	3	7	4	C	6	P
Subfamily Lonchorhininae													
<i>Lonchorhina aurita</i> Tomes, 1863	A, Ca, C and P	I	LC	VU	10	3	6	0	0	0	C	5	P
Subfamily Phyllostominae													
<i>Chrotopterus auritus</i> (Peters, 1856)	A, Ca, C, P and Pa	C/I	LC	NI	75	18	20	5	2	2	C	7	P
<i>Lophostoma brasiliense</i> Peters, 1866	A, C and Ca	I	LC	NI	8	4	5	3	0	0	C	4	P
<i>Lophostoma silvicola</i> d'Orbigny, 1836	A, Ca, C and P	I	LC	NI	9	1	5	1	0	0	C	7	P
<i>Macrophyllum macrophyllum</i> (Schinz, 1821)	A, C and P	I	LC	NI	2	1	0	0	0	0	C	3	P
<i>Mimon bennettii</i> (Gray, 1838)	A, Ca, C and P	I	LC	NI	30	12	9	2	2	2	C	3	P
<i>Mimon crenulatum</i> (É. Geoffroy, 1803)	A, Ca, C and P	I	LC	NI	8	6	2	1	2	2	C	3	P
<i>Phylloderma stenops</i> (Peters, 1865)	A, C and P	O	LC	NI	5	2	2	0	0	0	C	6	P
<i>Phyllostomus discolor</i> (Wagner, 1843)	A, Ca, C and P	O	LC	NI	38	18	7	1	3	2	C	5	P
<i>Phyllostomus elongatus</i> (É. Geoffroy, 1810)	A, C and P	O	LC	NI	4	1	3	1	0	0	C	2	P
<i>Phyllostomus hastatus</i> (Pallas, 1767)	A, Ca, C and P	O	LC	NI	66	38	17	6	4	1	C	6	P
<i>Tonatia bidens</i> (Spix, 1823)	P, C and Ca	O	DD	NI	39	17	22	7	3	1	C	7	P
<i>Tonatia saurophila</i> Koopman & Williams, 1951	Ca, C and A	O	LC	NI	5	3	4	2	0	0	C	4	P
<i>Trachops cirrhosus</i> (Spix, 1823)	A, C, Ca and P	O	LC	NI	51	26	25	9	3	2	C	6	P
Subfamily Glossophaginae													
<i>Anoura caudifer</i> (É. Geoffroy, 1818)	A, C, P and Pa	N	LC	NI	79	39	26	8	7	5	C	7	P
<i>Anoura geoffroyi</i> Gray, 1838	A, Ca, C, P and Pa	N	LC	NI	46	24	15	3	1	1	C	7	P
<i>Dryadonycteris capixaba</i> Nogueira,	Ca and C	N	DD	NI	5	5	1	1	2	2	C	1	R

Lima, Peracchi & Simmons, 2012													
<i>Glossophaga soricina</i> (Pallas, 1766)	A, Ca, C, P and Pa	N	LC	NI	223	75	47	21	21	5	C	7	P
<i>Lichonycteris degener</i> Miller, 1931	A	N	LC	NI	2	1	2	1	0	0	C	2	P
Subfamily Lonchophyllinae													
<i>Lonchophylla mordax</i> Thomas, 1903	Ca	N	LC	NI	13	11	2	2	1	1	C	4	P
<i>Lonchophylla peracchii</i> Dias, Esbérard & Moratelli, 2013	Ca	N	LC	NI	22	19	9	7	0	0	C	2	P
Subfamily Carolliinae													
<i>Carollia brevicauda</i> (Schinz, 1821)	A, Ca, C, P	F	LC	NI	14	12	9	8	0	0	C	5	P
<i>Carollia perspicillata</i> (Linnaeus, 1758)	A, Ca, C, P and Pa	F	LC	NI	302	145	90	35	5	0	C	7	P
Subfamily Glyphonycterinae													
<i>Glyphonycteris daviesi</i> (Hill, 1964)	A	I	LC	NI	2	2	1	1	0	0	C	1	R
<i>Glyphonycteris sylvestris</i> Thomas, 1896	A	I	LC	NI	4	2	3	1	0	0	C	4	P
<i>Trinycteris nicefori</i> (Sanborn, 1949)	A	I	LC	NI	3	2	2	1	0	0	C	3	P
Subfamily Rhinophyllinae													
<i>Rhinophylla pumilio</i> Peters, 1865	A; C and Ca	F	LC	NI	54	44	26	24	0	0	C	4	P
Subfamily Stenodermatinae													
<i>Dermanura cinerea</i> Gervais, 1856	A, C and Ca	F	LC	NI	61	36	32	20	19	2	C	5	P
<i>Artibeus fimbriatus</i> Gray, 1838	A, Ca, C, P and Pa	F	LC	NI	150	76	35	9	15	11	C	7	P
<i>Dermanura gnoma</i> (Handley, 1987)	A, C and Ca	F	LC	NI	7	5	5	4	0	4	C	2	R
<i>Artibeus lituratus</i> (Olfers, 1818)	A, Ca, C, P and Pa	F	LC	NI	201	106	60	26	12	9	C	7	P
<i>Artibeus obscurus</i> (Schinz, 1821)	A, Ca, C, P and Pa	F	LC	NI	106	65	52	30	2	2	C	6	P
<i>Artibeus planirostris</i> (Spix, 1823)	A, Ca, C, P and Pa	F	LC	NI	89	31	34	11	5	2	C	7	P
<i>Chiroderma doriae</i> Thomas, 1891	C and P	F	LC	NI	60	14	25	6	1	6	C	7	P
<i>Chiroderma villosum</i> Peters, 1860	A, C and P	F	LC	NI	36	22	17	11	8	1	C	6	P
<i>Platyrrhinus incarum</i> (Thomas, 1912)	A	F	LC	NI	2	1	2	1	0	0	C	3	P
<i>Platyrrhinus lineatus</i> (É. Geoffroy, 1810)	A, Ca, C, P	F	LC	NI	171	72	35	10	16	5	C	7	P

<i>Platyrrhinus recifinus</i> (Thomas, 1901)	and Pa C and Ca	F	LC	NI	55	40	15	8	3	3	C	4	P
<i>Pygoderma bilabiatum</i> (Wagner, 1843)	C, Ca and P	F	LC	NI	72	25	20	4	7	5	C	7	P
<i>Sturnira lilium</i> (É. Geoffroy, 1810)	A, Ca, C, P and Pa	F	LC	NI	179	78	45	11	13	10	C	7	P
<i>Sturnira tildae</i> de la Torre, 1959	A, Ca, C and P	F	LC	NI	31	7	12	2	1	0	C	4	P
<i>Uroderma bilobatum</i> Peters, 1866	A, Ca, C, P and Pa	F	LC	NI	5	2	2	1	0	0	C	2	P
<i>Uroderma magnirostrum</i> Davis, 1968	A, Ca, C, P and Pa	F	LC	NI	7	6	3	3	0	0	C	5	P
<i>Vampyressa pusilla</i> (Wagner, 1843)	C	F	DD	NI	62	31	19	7	5	3	C	7	P
<i>Vampyrodes caraccioli</i> (Thomas, 1889)	A	F	LC	NI	4	2	0	0	0	0	C	3	P
Family Noctilionidae													
<i>Noctilio albiventris</i> Desmarest, 1818	A, Ca, C and P	I	LC	NI	19	0	5	0	3	0	C	2	A
<i>Noctilio leporinus</i> (Linnaeus, 1758)	A, Ca, C and P	P	LC	NI	35	18	6	3	0	0	C	6	P
Family Furipteridae													
<i>Furipterus horrens</i> (Cuvier, 1828)	A, C and Ca	I	LC	NI	11	2	5	2	3	0	NC	7	P
Family Thyropteridae													
<i>Thyroptera tricolor</i> Spix, 1823	A	I	LC	NI	4	3	1	0	0	0	C	2	P
Family Natalidae													
<i>Natalus macrourus</i> (Gervais, 1856)	A, C and Ca	I	NT	VU	3	1	2	1	0	0	C	5	P

ANEXOS

ANEXO A

Apresentação das normas de formatação do periódico “Neotropical Biology and Conservation”. Disponível em: <http://revistas.unisinos.br/index.php/neotropical/about/submissions#onlineSubmissions> Acessado em: 20 Abril, 2020.

Author Guidelines**Submission Preparation Checklist**

As part of the submission process, authors are required to check off their submission's compliance with all of the following items, and submissions may be returned to authors that do not adhere to these guidelines.

The work is original and unpublished and it is not under consideration for publication anywhere else.

The manuscript file is in Microsoft Word or RTF.

The original text must contain in the following order:

Title in the same language of the article, and in English (if the article is in English, title must also be in Portuguese). The titles (on the original language and English) must have a maximum of 240 characters including spaces.;

Abstract with no more than 20 lines, followed by at least three key-words;

Abstract in English (or in Portuguese, in case the article is written in English), following the same norms of the Abstract in the original language;

Full text, typed in Times New Roman format, font 12, and with spacing of 1,5;

Images, tables, charts and graphs must include captions in the same language of the article and in English, they must be sent separately in their original files (.jpeg, .png, .tiff), not inserted in the text, with minimal resolution of 300 dpi.

The identification of authorship of the work has been removed from the file, ensuring the journal's confidentiality criteria, if the manuscript is subjected to peer review (e.g. articles), according to instructions in Ensuring a Blind Peer Review.

Short quotations (containing less than three lines) must be incorporated into the text in quotation marks;

Long quotations (longer than three lines) must be presented in an isolated paragraph, with simple spacing, font size 11 and the left margin larger than the surrounding text (4cm).

Footnotes must be used only when strictly necessary, for information that is complementary to the text. Do not use footnotes as a substitute for a reference list.

The manuscript is in accordance to the style and bibliographic requirements

described in the Author's Guideline section.

References

The list of references must only include works that are cited in the text. Reference list entries must be in alphabetical order according to the last names of the first author of each work.

References must be typed in Times New Roman (pt 12) and single spacing between lines.

References must contain the work's complete data according to the journal's rules.

Examples:

Journal articles:

SURNAME, Initial(s) name. Year of publication. Article title. Journal Title, volume(issue):initial page-final page.

Articles from publications related to events:

SURNAME, Initial(s) name. Year of publication. Title of work. In: name of conference (meeting, symposium, etc.), number of the event, city, year. Proceedings ... City, Symbol. volume:: initial page-final page.

Books:

SURNAME, Initial(s) name. Year of publication. Book title. Edition, city, publisher, the total number of pages.

Book chapters:

SURNAME, Initial(s) name. Year of publication. Chapter title. In: Initial(s) Name. SURNAME (ed.), Book title. City, publisher, initial page-final page.

Thesis and dissertations:

SURNAME, Initial(s) name. Year of publication Title of the thesis. City, state. Type of thesis (MA, PhD). University, the total number of pages.

Sites and electronic texts:

Explicit author:

SURNAME, Initial(s) name. Year of publication. Title of work. Available at: <http://>. Accessed on: month day, year.

Anonymous author:

SOURCE/SITE. Initial(s) name. Year of publication. Title of work. Available at: <http://>. Accessed on: month day, year.

Newspapers and magazines, organs and institutions:

Explicit author:

SURNAME, Initial(s) name. Year of publication. Title of work. Source (newspaper, organs, institutions, etc.). Session (Column, etc.). City, day month.

Anonymous author:

SOURCE (newspaper, organs, institutions, etc.). Year of publication. Title of work. City, day month, number page.

ANEXO B

Apresentação das normas de formatação do periódico “Journal of Biogeography”. Disponível em: <https://onlinelibrary.wiley.com/page/journal/13652699/homepage/forauthors.html>
Acessado em: 20 Abril, 2020.

Author Guidelines

1. SUBMISSION

Authors should kindly note that submission implies that the content has not been published or submitted for publication elsewhere except as a brief abstract in the proceedings of a scientific meeting or symposium. All submissions must be concisely and clearly written in grammatically correct English.

Once the submission materials have been prepared in accordance with the Author Guidelines, manuscripts should be submitted online at <https://mc.manuscriptcentral.com/jbi>

The submission system will prompt authors to use an ORCID iD (a unique author identifier) to help distinguish their work from that of other researchers. Click here to find out more.

For help with submissions, please contact the Editorial Office at jbiooffice@wiley.com

2. AIMS AND SCOPE

SCOPE:

The Journal of Biogeography publishes research at the intersection of biology and geography that is scientifically important and of broad general interest. We seek papers describing patterns and revealing mechanisms that shape biodiversity, through time, throughout the planet, from the deep past into the future, and from local to global scales. Diverse approaches are encouraged—including ecological, evolutionary, genomic, geographic, empirical, theoretical—considering any aspect of biogeography, from molecules to ecosystems and from microbes to plants and megafauna. Through this broad and inclusive scope, we aim for papers in Journal of Biogeography to address understudied, vexing, and urgent questions and to advance basic understanding of the origins, distributions, and fates of life on Earth.

Manuscripts submitted to Journal of Biogeography should be original and innovative, concise, well written, rigorously analyzed and argued, and consequential. While many such studies will be multifaceted, comparative, and draw generalities, we also welcome exceptional case studies that illustrate particularly interesting deviations that, in their aggregate, shift preconceptions.

The Journal of Biogeography is edited and reviewed for the community by a team of practising biogeographers. We support open data, accessibility to publish and read, and a constructive peer-review process.

VISION:

The Journal of Biogeography is the discipline's first and foremost journal. It's established history publishing influential papers in biogeography, its topical breadth, and its strong reputation in the community, provide the foundation for the journal to continue to grow as the most respected journal in the field. Nonetheless, as disciplinary and publishing trends change, to remain at the forefront of biogeography, the journal must innovate such that it represents not only core biogeography but also novel advances in emerging areas. Biogeography is an integrative discipline and the journal aims to increasingly complement its strong foundations with the most exciting multidisciplinary research.

3. MANUSCRIPT CATEGORIES AND REQUIREMENTS

The Journal publishes articles under the following main headers: 1) Research Paper, 2) Methods and Tools, 3) Data, 4) Synthesis, 5) Perspective, 6) Commentary and 7) Correspondence. All submissions are subject to peer review.

1) Research Paper. Research papers present new biogeographic research resulting from the analysis of a question in biogeography. For a typical Research paper, in which illustrative material (Tables and Figures) occupies about 3 pages of the journal when printed at final journal sizing, the text, inclusive of abstract and reference list, should not exceed 7000 words. Manuscripts should include a biosketch (see below); tables with their legends above; list of figure legends; and embedded figures, and the main headers in the main text of Research Papers should normally be Introduction, Materials and Methods, Results, Discussion, Acknowledgements, References. Methods need to be described in a manner that allows a competent practitioner in the field to repeat the study. Authors must allow repeatability by either providing a thorough description of the methods or by providing relevant computer code.

Structured abstracts. Abstracts should be of no more than 300 words, presented as a series of factual statements under the following headings: Aim, Location, Taxon, Methods, Results and Main conclusions. The Aim should give a clear statement of the principal research question(s) or hypotheses, the Taxon indicate the main group (eg angiosperms), the Methods should give details of materials/sampling/methods of analysis, and the Main conclusions should give the main take-home message.

Biosketch/Biosketches. A short Biosketch/Biosketches entry (30-100 words for one author/150 words total for the first three authors, respectively) describing the research interests of the author(s) should be provided. For papers with four or more authors, biosketch details should be supplied for the first author only and/or a general statement of the focus of the research team (which may include a link to a group web page) plus, in all cases, a statement of author contributions, e.g. Author contributions: A.S. and K.J. conceived the ideas; K.J. and R.L.M. collected the data; R.L.M. and P.A.K. analysed the data; and A.S. and K.J. led the writing.

For an example click [here](#).

2) Methods and Tools. These are structured as in Research papers, but the main focus is to present or investigate a new method, rather than to explore a biogeographical problem. Papers in this section are expected to apply new methods

to the analysis of biogeographic data and discuss the potential of those methods for advancing the study of the field. For an example [click here](#).

3) Data. Datasets that are likely to be of interest to the broader biogeography community are published under this category. Data papers allow scientists to publish and receive credit for work in which the nature of the collected, mobilized, or integrated data more than a specific analysis may be most impactful. The structure of a Data paper should be similar to that of a Research paper. Data papers must include a characterization of overall scope (e.g. organismal, spatial), a description of how the data were collected (protocols), a detailed characterization of all data fields and metadata, information on data records (e.g. SI or in a suitable repository), a section on technical validation, and usage notes addressing potential caveats for analysis and interpretation. Additional analyses that exemplify potential uses of the data are encouraged but not essential. Note that the Data papers category is intended for novel datasets - data used in a published or submitted research paper should be fully addressed and made available there.

4) Synthesis. Papers that have the character of a theoretical synthesis or review, even if incorporating an element of original analysis within them, should use the article type Synthesis. Guidelines are as for Research papers but submissions to the Synthesis section may be of up to 10,000 words, or exceptionally more, if the additional length is fully justified. Authors of synthesis papers are encouraged to discuss their planned paper with one of the Chief Editors, especially if the length will exceed 10,000 words. For an example [click here](#).

5) Perspective. Perspectives papers are should be stimulating and reflective essays providing personal perspectives on key research fields and issues within biogeography. When published, Perspectives should be of no more than eight printed pages (main text maximum 5000 words; word count including abstract, main text and references 7000 words maximum but note that shorter articles are encouraged), and they should include a short, single-paragraph abstract. A biosketch (see below) may be included after the references providing the overall paper length limit is not exceeded. For an example [click here](#).

6) Commentary. Commentary submissions should provide readily intelligible comment on the latest original research in biogeography. The prose style should be light, and the article should be written with the minimum of technical language and jargon, so as to be understandable to a general audience or an undergraduate taking an introductory course in biogeography. Contributions will be subject to rapid peer review. Commentaries should occupy a maximum of two pages of the journal, and should have a maximum of 10 references: thus the overall word count should not exceed 1600. No biosketch is included for commentaries. Should you wish to include a small figure or other illustration, this can be accommodated by a reduction in the number of words on a pro rata basis. For an example [click here](#).

7) Correspondence. The Journal welcomes short items of correspondence prompted by papers previously published in this or occasionally in other journals. The text should not normally exceed 2500 words, inclusive of a short one-paragraph abstract (up to 150 words), and a list of 6–10 keywords. No biosketch is necessary for Correspondence papers. For an example [click here](#).

4. PREPARING THE SUBMISSION

Article Preparation Support

Wiley Editing Services offers expert help with English Language Editing, as well as translation, manuscript formatting, figure illustration, figure formatting, and graphical abstract design – so you can submit your manuscript with confidence. Also, check out our resources for Preparing Your Article for general guidance about writing and preparing your manuscript.

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A cover letter to the editor, indicating in less than 100 words why this paper is of interest to the readers of the Journal, must be uploaded separately.

Parts of the Manuscript

The manuscript should be submitted in separate files: main text file with embedded figures; supporting information.

LaTeX users do not have to translate their manuscripts into MSWord, but may upload them as PDF files. Any explanatory notes, companion papers etc. for the attention of reviewers should be uploaded under 'Comments to reviewers'.

Main Text File

The text file should be single spaced, or 1.3 spaced, and presented in the following order:

- i. Title
- ii. A short running title of less than 40 characters
- iii. The full names of the authors, only 1 corresponding author may be included
- iv. The author's institutional affiliations where the work was carried out, with a footnote for the author's present address if different from where the work was carried out
- v. Acknowledgements
- vi. Abstract and keywords
- vii. Main text
- viii. Tables embedded in the text (each table complete with title and footnotes)
- ix. Figures embedded in the text, each with a figure legend
- x. Data availability statement
- xi. References
- xii. Biosketch

xiii. Appendices (if relevant)

xiv. Supporting information should be supplied as separate files.

Title. The title should be short and informative, containing major keywords related to the content. The title should not contain abbreviations (see Wiley's best practice SEO tips).

Authorship. For details on eligibility for author listing, please refer to the journal's Authorship policy outlined in the Editorial Policies and Ethical Considerations section. Only 1 corresponding author may be included.

Acknowledgements. Contributions from individuals who do not meet the criteria for authorship should be listed, with permission from the contributor, in an Acknowledgements section. Financial and material support should also be mentioned. Thanks to anonymous reviewers are not appropriate.

Conflict of Interest Statement. Authors will be asked to provide a conflict of interest statement during the submission process. See 'Conflict of Interest' section in Editorial Policies and Ethical Considerations for details on what to include in this section. Authors should ensure they liaise with all co-authors to confirm agreement with the final statement.

Abstract and Keywords

Abstracts and keywords are required for some manuscript types. For details on manuscript types that require abstracts and/or keywords, as well as how to prepare them, please refer to the 'Manuscript Categories and Requirements' section. Please provide 6-10 keywords, arranged alphabetically, separated by commas. Note that optimally the most important keywords are repeated in the title and the keywords.

Main Text

The journal uses British spelling; however, authors may submit using either option, as spelling of accepted papers is converted during the production process.

References

References are styled according to the sixth edition of the Publication Manual of the American Psychological Association. List all sources in the reference alphabetically by name.

In text citations should follow the author-date method. This means that the author's last name and the year of publication for the source should appear in the text, for example, (Jones, 1998), and a complete reference should appear in the reference list at the end of the paper.

When a work has two authors, cite both names every time the reference occurs in text. When a work has three, four, or five authors, cite all authors the first time the reference occurs; subsequent citations include only the surname of the first author followed by et al., (not Italicized and with a period after "al.") and the year if it is the first citation of the reference within a paragraph.

If there are two or more citations that shorten to the same lead author and date, give as many additional names as needed to identify them, e.g., (Smith, Jones, et al., 1991) and (Smith, Burke, et al., 1991).

Unpublished data, works in preparation and papers submitted but not yet accepted may be cited in the text as personal communication, giving the author's initials and surname, but should not be included in the reference list. It is the author's responsibility to obtain permission from colleagues to include their work as a personal communication. Please add the person's initials, surname and if applicable institute for personal communications.

The basic reference form for a journal paper is: Author (date). Paper title. Journal, Volume, page; and for a book citation: Author (date). Book title. Place of publication, publisher.

Please note that for journal articles, issue numbers are not included unless each issue in the volume begins with page one. Journals names are written out in full.

Please ensure that in the paper titles only proper names are capitalized, and that all scientific binomials are in italics.

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Journal article:

Light, M. A., & Light, I. H. (2008). The geographic expansion of Mexican immigration in the United States and its implications for local law enforcement. *Law Enforcement Executive Forum Journal*, 8(1), 73–82.

Book:

Goldstein, H. (1990). *Problem-oriented policing*. New York, NY: McGraw-Hill. Miles, M. B., & Huberman, A. M. (1994). *Qualitative data analysis* (2nd ed.). Thousand Oaks, CA: Sage.

Edited Book:

Gilbert, D. G., McClernon, J. F., Rabinovich, N. E., Sugai, C., Plath, L. C., Asgaard, G., ... Botros, N. (1983). Situational crime prevention: Its theoretical basis and practical scope. In M. Tonry & N. Morris (Eds.), *Crime and justice: An annual review of research* (Vol. 4, pp. 225–256). Chicago, IL: University of Chicago Press.

Citations to data sources

Some studies (e.g., meta-analyses) use data drawn from multiple published sources. If these sources are not otherwise cited in the main text, they should be listed in one or more appendices with titles similar to the following: "Appendix 1 – Data sources". These data appendices will be printed in the main paper (so that citation indexing services will capture them), but in a reduced font. These appendices should be cited in the main text (e.g. "A list of the data sources is found in Appendix 1."), and be placed after the biosketch in the manuscript.

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Bar scales for maps and photographs are preferred to numerical scales and must be given on all such items. Maps that display area data and organism distribution at a continental, hemispheric, or world scale must always use an equal-area map projection (e.g. Mollweide or Aitoff's). Note especially that Mercator's projection is not acceptable for such data. Please indicate the precise projection employed in the caption. On these maps, the equatorial scale should be indicated, while scale information should be provided, preferably as a scale bar within the figure, for all maps of whatever size and area; use 'km' or 'kilometres', not 'kilometers'. Maps should include adequate geo-referencing information (preferably the latitude and longitude).

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**BIOLOGICAL CONSERVATION****AUTHOR
INFORMATION PACK****TABLE OF
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ISSN: 0006-3207

1 DESCRIPTION

Biological Conservation is a leading international journal in the discipline of **conservation science**. The journal publishes articles spanning a diverse range of fields that contribute to the biological, sociological, ethical and economic dimensions of **conservation**. The primary aim of *Biological Conservation* is the publication of high-quality papers that advance the science and practice of conservation, or which demonstrate the application of conservation principles and policy. *Biological Conservation* invites the submission of research articles, reviews (including systematic reviews and perspectives), short communications, policy perspectives, and letters to the editor dealing with all aspects of conservation science, including theoretical and empirical investigations into the consequences of human actions for the diversity, structure and function of terrestrial, aquatic or marine ecosystems. Such papers may include quantitative assessments of extinction risk, fragmentation effects, spread of invasive organisms, conservation genetics, conservation management, global change effects on biodiversity, landscape or reserve design and management, restoration ecology, or resource economics. We also welcome papers coming from social sciences including those reporting on advances in conservation politics, ethics, policy, human social structure and biodiversity, and political culture among other subjects. *Biological Conservation* covers interdisciplinary topics within conservation biology and also provides practical applications of conservation research for land/resource managers and policy makers. We publish articles and thematic special issues that have a global relevance in terms of the topics or issues addressed, and thus demonstrate applications of conservation science and management beyond the specific system or species studied. *Biological Conservation* is an affiliate publication of the Society for Conservation Biology (SCB). SCB members can obtain a personal subscription to this journal through the Society.

2 AUDIENCE

Environmentalists, conservationists, botanists, marine scientists, ecologists, biologists, zoologists.

3 IMPACT FACTOR

2018: 4.451 © Clarivate Analytics Journal Citation Reports 2019

4 ABSTRACTING AND INDEXING

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5.2 INTRODUCTION

Please read all information carefully and follow the instructions in detail when preparing your manuscript.

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Biological Conservation encourages the submission of high-quality manuscripts that advance the science and practice of conservation, or which demonstrate the application of conservation principles and policy. **Conservation implications should be clearly emphasized and discussed.** Given the broad international readership of the journal, published articles should have global relevance in terms of the topics or issues addressed, and thus demonstrate applications for conservation beyond the specific system or species studied.

5.2.1 Types of paper

Word counts include text, references, figures and tables. Each figure or table should be considered equal to 300 words.

5.2.1.1 **1. Full length articles (Research papers)**

Research papers report the results of original research. The material must not have been previously published elsewhere. Full length articles are up to 8,000 words.

5.2.1.2 **2. Review articles**

Reviews summarize the status of research in a field of current interest. They may be submitted or invited. Review articles are up to 12,000 words and must include a methods section explaining how the literature for review was selected. We also consider Systematic Reviews, which apply a methodology to synthesize and appraise the scientific evidence on a specific question or hypothesis. More about systematic reviews can be found here: <http://www.environmentalevidence.org/information-for->

authors.

3. Perspectives:

These articles provide an opportunity for authors to present a novel, distinctive viewpoint on any subject within the journal's scope. The article should be well grounded in evidence and adequately supported by citations but may focus on a stimulating and thought-provoking line of argument that represents a significant advance in thinking about conservation problems and solutions. Perspectives articles should not exceed 8000 words.

5.2.1.3 4. Short communications

Short communications highlight both novel research and replication studies that report preliminary findings that are particularly compelling and highly relevant to conservation science and practice. If submitting a replication study, please include in your cover letter the rationale for undertaking the study. Short communications should not exceed 4,000 words.

5.2.1.4 5. Policy Analysis

These are short commentary pieces on contemporary, internationally relevant conservation or conservation-related policy issues that enable researchers, policy makers, and practitioners to make timely contributions to policy debates and actions. Contributions are based on research, expert analysis, literature review, or practitioner reflections regarding specific policy issues. Pure opinion pieces will not be considered for this paper type. Forum articles should be written in an accessible style and supported by real world examples and/or referenced scientific evidence and should not exceed 4,000 words.

5.2.1.5 6. Fast-Tracked Papers

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5.2.1.6 7. Book Reviews

Book reviews will be included in the journal on a range of relevant titles that are not more than two years old. These are usually less than 2,000 words. Please submit your requests/ideas to David Johns at johnsd@embarqmail.com.

5.2.1.7 8. Editorials

Opinion pieces by experts on a topic, usually invited by the Editor. The topic is usually timely and offers important insights into the field.

5.2.1.8 9. Correspondence

Letters to the Editor (Correspondence) papers are responses to recently published papers. Letters must be short (a maximum 800 words) and include only key references (5 maximum) and one figure if necessary. The content must be constructive, discuss significant issues, and respectful in tone. Provided the editors agree that publication of the letter is warranted, it will generally also receive a response from the authors of the original article, and both letter and response will be published in the same issue.

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There are no strict requirements on reference formatting at submission. References can be in any style or format as long as the style is consistent. Where applicable, author(s) name(s), journal title/ book title, chapter title/article title, year of publication, volume number/book chapter and the article number or pagination must be present. Use of DOI is highly encouraged. The reference style used by the journal will be applied to the accepted article by Elsevier at the proof stage. Note that missing data will be highlighted at proof stage for the author to correct. If you do wish to format the references yourself they should be arranged according to the following examples:

5.4.15.7 Reference style

Text: All citations in the text should refer to:

1. *Single author:* the author's name (without initials, unless there is ambiguity) and the year of publication;
2. *Two authors:* both authors' names and the year of publication;
3. *Three or more authors:* first author's name followed by 'et al.' and the year of publication. Citations may be made directly (or parenthetically). Groups of references can be listed either first alphabetically, then chronologically, or vice versa.

Examples: 'as demonstrated (Allan, 2000a, 2000b, 1999; Allan and Jones, 1999).... Or, as demonstrated (Jones, 1999; Allan, 2000)... Kramer et al. (2010) have recently shown ...'

List: References should be arranged first alphabetically and then further sorted chronologically if necessary. More than one reference from the same author(s) in the same year must be identified by the letters 'a', 'b', 'c', etc., placed after the year of publication.

Examples:

Reference to a journal publication:

Van der Geer, J., Hanraads, J.A.J., Lupton, R.A., 2010. The art of writing a scientific article. *J. Sci.*

Commun. 163, 51–59. <https://doi.org/10.1016/j.Sc.2010.00372>.

Reference to a journal publication with an article number:

Van der Geer, J., Hanraads, J.A.J., Lupton, R.A., 2018. The art of writing a scientific article. *Heliyon*. 19, e00205. <https://doi.org/10.1016/j.heliyon.2018.e00205>.

Reference to a book:

Strunk Jr., W., White, E.B., 2000. *The Elements of Style*, fourth ed. Longman, New York.

Reference to a chapter in an edited book:

Mettam, G.R., Adams, L.B., 2009. How to prepare an electronic version of your article, in: Jones, B.S., Smith, R.Z. (Eds.), *Introduction to the Electronic Age*. E-Publishing Inc., New York, pp. 281–304.

Reference to a website:

Cancer Research UK, 1975. Cancer statistics reports for the UK. <http://www.cancerresearchuk.org/aboutcancer/statistics/cancerstatsreport/> (accessed 13 March 2003).

Reference to a dataset:

[dataset] Oguro, M., Imahiro, S., Saito, S., Nakashizuka, T., 2015. Mortality data for Japanese oak wilt disease and surrounding forest compositions. *Mendeley Data*, v1. <https://doi.org/10.17632/xwj98nb39r.1>.

5.4.16 Video

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Include interactive data visualizations in your publication and let your readers interact and engage more closely with your research. Follow the instructions here to find out about available data visualization options and how to include them with your article.

5.4.18 Supplementary material

Supplementary material such as applications, images and sound clips, can be published with your article to enhance it. Submitted supplementary items are published exactly as they are received (Excel or PowerPoint files will appear as such online). Please submit your material together with the article and supply a concise, descriptive caption for each supplementary file. If you wish to make changes to supplementary material during any stage of the process, please make sure to provide an updated file. Do not annotate any corrections on a previous version. Please switch off the 'Track Changes' option in Microsoft Office files as these will appear in the published version.

The supplementary material should be cited as an online Appendix to the paper, usually in the Methods. If it contains several tables, images and/or figures, these should be cited as Table A1, Figure A1 and so on.

Authors are strongly encouraged to make the data supporting their paper available to readers through an open-access data repository and/or as an Appendix to the paper. For more details on journal data policy see the paragraphs on *Data Depositing and Linking to and depositing data at PANGAEA*.

5.4.18.1 **Data Depositing**

Ideally, data should be freely available online through a specialist data centre that provides a permanent archive (repository) for the dataset, and may integrate the data with other datasets using international standards. Examples include PANGAEA, and GBIF and its major contributors such as OBIS and VertNet. Some Ocean Data Centres may also provide this service. Where such a data centre does not exist, we ask that the data be made freely available online from a permanent archive (repository). Where possible, it should follow international data standards. This may be an institutional repository for its staff. The data should be accompanied by sufficient information (metadata) for the reader to understand its composition and origins, and determine if it is fit for their purpose. In particular, the data should allow the results of the publication to be reproduced. Data being downloadable from departmental or personal websites is not regarded as permanently archived.

5.4.19 Research data

This journal encourages and enables you to share data that supports your research

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5.4.19.3 Data in Brief

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5.4.19.4 MethodsX

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5.4.19.5 Data statement

To foster transparency, we encourage you to state the availability of your data in your submission. This may be a requirement of your funding body or institution. If your data is unavailable to access or unsuitable to post, you will have the opportunity to indicate why during the submission process, for example by stating that the research data is confidential. The statement will appear with your published article on ScienceDirect. For more information, visit the Data Statement page.

5.5 AFTER ACCEPTANCE

5.5.1 Online proof correction

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ANEXO D

Declaração de tradução profissional do capítulo 1

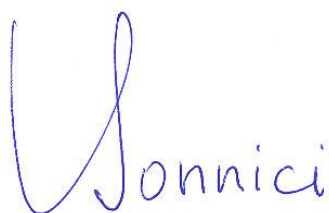
CERTIFICATE

To the Editorial Board

This is to certify that the paper **“THIRTY-SIX YEARS OF RESERACH ON BATS IN THE REGION OF LONDRINA, BRAZIL: WHAT HAVE WE LEARNT?”** by Carolina Blefari Batista, Isaac Passos de Lima and Marcos Robalinho Lima, was translated by the undersigned, who is a native speaker of English. The translation is grammatically and stylistically to standard.

Maringá PR Brazil, 22nd April 2020.

Yours truly,



Dr. Thomas Bonnici

I.D. card 563 250

Department of English, State University of Maringá, Maringá PR Brazil.

ANEXO E

Declaração de tradução profissional do capítulo 3

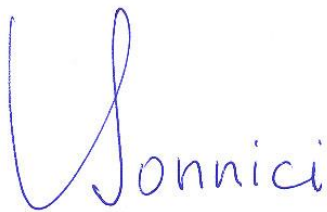
CERTIFICATE

To the Editorial Board

This is to certify that the paper "**DOWNSCALING THE ATLANTIC FOREST BIODIVERSITY HOTSPOT: USING THE DISTRIBUTION OF BATS TO FIND SMALLER HOTSPOTS OF CONSERVATION PRIORITY**", by Carolina Blefari Batista, Isaac Passos de Lima, Rafael Arruda and Marcos Robalinho Lima, was translated by the undersigned, who is a native speaker of English. The translation is grammatically and stylistically to standard.

Maringá PR Brazil, 25th April 2020.

Yours truly,



Dr. Thomas Bonnici

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